





Final Technical Report

2008

Surface Water Ambient Monitoring Program (SWAMP) Synthesis Report on Stream Assessments in the San Diego Region

March 2008



SURFACE WATER AMBIENT MONITORING PROGRAM (SWAMP) SYNTHESIS REPORT ON STREAM ASSESSMENTS IN THE SAN DIEGO REGION

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Prepared for the California Regional Water Quality Control Board, San Diego Region (Region 9).

This project was funded by the Surface Water Ambient Monitoring Program.

March 2008

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1. EXECUTIVE SUMMARY

Watershed managers require regional data to develop biomonitoring tools and contextualize local assessments. However, they often rely on data generated by programs with a more local emphasis, such as studies mandated by pollution discharge permits. These programs typically study only specific sites or stream reaches, The goal of this study was to compile individual data sets from site-specific programs to see if they could be merged into a regional-scale program. The compiled data was used to ask three questions: 1) perform a regional assessment of stream health in southern California; 2) identify potential stressors to aquatic life in these streams; and 3) how can existing programs be modified to better address the first two questions.

Water quality, toxicity, physical habitat, and benthic macroinvertebrate samples were collected from over 140 sites in coastal watersheds in San Diego, Riverside, and Orange counties by six different programs. Data were collected by six different programs including the State's Surface Water Ambient Monitoring Program (SWAMP), California Department of Fish and Game, and multiple National Pollutant Discharge Elimination System (NPDES) monitoring programs. Watershed health was assessed by comparing each indicator to applicable thresholds. Scatterplots and multivariate ordinations were used to examine relationships among potential stressors and benthic macroinvertebrate communities.

The results indicated widespread impacts to many water chemistry constituents with some, like ammonia-N and specific conductivity, exceeded aquatic life thresholds in more than 60% of samples. More than 50% of water and sediment samples were toxic to at least one indicator species (*Ceriodaphnia dubia, Hyallela azteca* or *Selenastrum capricornutum*). Physical habitat was degraded (i.e., mean physical habitat score < 10) at 30% of the assessed sites. Of the 708 bioassessment samples included in the study, 80% were in poor condition (i.e., index of biotic integrity < 40). Impacts for all indicators were most severe in urban areas along the coast.

Although data merged from multiple monitoring programs provided an assessment of stream health in the region, interpretations were limited due to a lack of coordination among programs. For example, the data set was undoubtedly biased because the individual programs typically focused on identifying likely impacted areas in order to address site-specific issues. As a result, the data set was not spatially representative, but instead focused on potentially disturbed areas at the exclusion of potentially healthy areas in the region.

Analysis for potential stressors indicated that multiple stressors likely impacted aquatic life. Nonmetric multidimensional scaling of benthic

macroinvertebrate communities identified two stressor gradients: a strong gradient associated with toxic contaminants in the water (e.g., metals, high specific conductivity, and organics) and a weaker gradient related to nutrient enrichment (e.g., elevated nitrate and low dissolved oxygen). The toxic contaminant gradient was strongly associated with development in the watershed, and watersheds with more than 20% developed area were invariably in poor biological health.

Despite the potential for bias that limited the regional assessment, the merged data from multiple programs could be effective at evaluating stressors. The data captured important gradients in the region. However, diagnosing the specific causes of impairment at individual sites will require additional studies, where synoptic data are collected and analyzed.

Four improvements were recommended for the SWAMP program to enhance its ability to merge data from multiple programs. First, SWAMP should coordinate with other programs operating in the region to increase efficiency, share sampling sites, and save resources. One such program, the Stormwater Monitoring Coalition (SMC) beginning in 2009, provides a mechanism for this type of collaboration. Second, the monitoring program should utilize a probabilistic design where sites are selected randomly, rather than a targeted design, will eliminate bias and provide a more accurate picture of the overall condition of the region's watersheds. Third, the monitoring program should identify a core set of indicators sampled synoptically at all sites to determine impacts to beneficial uses. These indicators should be selected in coordination with the other programs in the region. Many indicators examined in this report may work well, but others such as algae to detect nutrient enrichment should be considered. Fourth, ensure that there is an infrastructure to support collaborative programs. Consistent data structures, quality assurance requirements, and comparable field and lab methodology is essential for collaboration across multiple programs and will greatly increase efficiency of bioassessments in the region.

2. INTRODUCTION

With few exceptions (e.g., EPA 2006), stream monitoring has been applied in largely piecemeal fashion around the country in response to regulatory-based requirements. Southern California serves as a good example. Collectively, more than 12 agencies collect over 300 samples in the 18 major coastal watersheds in just the 6 from San Diego to Ventura. For the most part, these programs employ independent, site-specific monitoring designs that target specific discharge locations (SMC 2007).

Despite the lack of programs that focus beyond specific sites or watersheds, there is a tremendous need for regional scale evaluations of stream health. The first need for regional stream monitoring is to address questions posed by the public that tend to focus on streams as a whole as opposed to just the reaches where there are potential sources of impacts. The second need for regional stream monitoring is to develop assessment tools that watershed managers need for evaluating potential impacts. One such tool is stressorresponse relationships (Van Sickle et al. 2006). Regional scale programs are one way to collect the information necessary for developing stressor-response relationships because they not only capture the full breadth of natural variation, but a wide range of anthropogenically induced impacts. It is this range of stressor impact that is important for anchoring the spectrum of stream responses. The third need for regional stream monitoring is to help set management priorities, which is especially important in these times of limited resources. Regional monitoring programs provide the context of the worst and best streams that allow managers to effectively target the locations of greatest need.

One mechanism to achieve regional scale assessments is to link multiple local- or watershed-scale programs. However, there are many challenges associated with this approach. Individual programs may not measure similar indicators or, if they do, they may not measure the indicators using similar methods. Different programs may also have differing levels of quality assurance. The result is to force all of the data to the lowest level of QA common among them, which may be insufficient for management-level assessments. Finally, even if indicators, methods, and quality assurance were similar, data management can present an enormous hurdle. Undoubtedly, each monitoring program stores its data in different ways, from simple to sophisticated, making the collation of data an unusually burdensome (if not impossible) task.

The goal of this study was to determine if multiple, local scale data sets could be combined to make regional scale assessments. In this case study, we selected data sets collected by the San Diego Regional Water Quality Control Board's Surface Water Ambient Monitoring Program, as well as by numerous dischargers regulated by the National Pollutant Discharge Elimination System (NPDES) permits in the San Diego region to answer three questions of regional importance:

- 1) What is the health of streams in the San Diego region?
- 2) What are the primary stressors responsible for biological responses in the San Diego Region?
- 3) How can existing programs be modified to better address the first two questions?

The challenge was to combine data collected over 11 different hydrologic units and a time period of 9 years. None of the data was collected in concert, none of the designs was integrated, and none of the data systems were connected in any way. Thus, this study was an evaluation of the ability of such programs to address regional needs

3. METHODS

Setting

The San Diego region includes all coastal watersheds north of the Mexican border and south of the Santa Ana River. Covering portions of Orange, Riverside, and San Diego Counties, the region encompasses nearly 4,000 mi² and ranges from the mountains of the Peninsular Range mountains to the Pacific Coast.

Southern California is characterized by an arid mediterranean climate, with hot dry summers and cool wet winters. Average monthly rainfalls measured at the Lindberg Airport (SDG) in San Diego, California between 1905 and 2006 show that nearly all rain fell between the months of October and April, with hardly any falling between the months of May and September (California Department of Water Resources 2007). The wettest month was January, with an average rainfall of 2.05"). Average annual rainfall at this station was 10.37".

The San Diego Region consists of several coastal rivers and streams that are grouped into 11 hydrologic units (Figure 1, Table 1). The Tijuana River is the largest in the region. Other large rivers include the Santa Margarita, San Luis Rey, San Dieguito, San Diego, Sweetwater, and Otay Rivers (Figure 1). The watersheds extend from the Lagunas, the Cuyamacas, and other mountains of the Peninsular Range. Most of the larger rivers are regulated by large dams. The streams of the San Diego Region have profound effects on coastal ecology and the Southern California Bight (Ackerman and Schiff 2003). Discharging over 300 million m³ annually in typical years, the rivers are an important source of freshwater for San Diego and Mission Bay, as well as several estuaries and coastal wetlands.

Urban development extends along almost the entire coastal strip of the region (23% of the region), although large undeveloped areas remain in coastal northern San Diego County in Camp Pendleton Marine Corps Base. Many

smaller coastal watersheds are entirely urbanized. Agricultural land use occurs in 9% of the region, and is most extensive in the San Luis Rey and San Dieguito watersheds. Open space predominates in the interior, as well as in the aforementioned Camp Pendleton, covering 68% of the region (SANDAG 1998). The extent of undeveloped open space varies among each watershed, from a low of 12% in Pueblo San Diego to a high of 92% in San Juan (Table 1).

Table 1. Watersheds in the San Diego region. Land uses are calculated from data provided by SANDAG (1998)

| Watersheds | Abbreviation | HUC | Area (mi ²) | % Open | % Developed | % Agricultural |
|------------------|--------------|-----|-------------------------|--------|-------------|----------------|
| San Juan | SJ | 901 | 496 | 92 | 7 | 1 |
| Santa Margarita | SM | 902 | 750 | 81 | 13 | 6 |
| San Luis Rey | SLR | 903 | 560 | 61 | 15 | 24 |
| Carlsbad | СВ | 904 | 211 | 38 | 50 | 12 |
| San Dieguito | STO | 905 | 346 | 18 | 61 | 21 |
| Los Peñasquitos | LP | 906 | 162 | 43 | 53 | 4 |
| San Diego | SD | 907 | 440 | 72 | 26 | 2 |
| Pueblo San Diego | PSD | 908 | 56 | 12 | 88 | 0 |
| Sweetwater | SW | 909 | 230 | 67 | 29 | 4 |
| Otay | OT | 910 | 154 | 70 | 20 | 10 |
| Tijuana | TJ | 911 | 463 | 90 | 6 | 4 |
| TOTAL | | | 3868 | 68 | 23 | 9 |

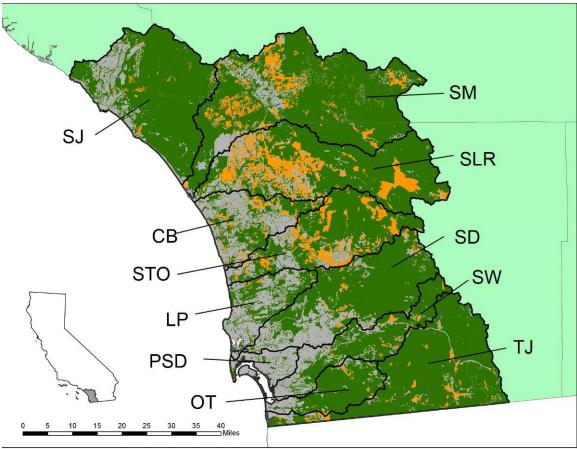


Figure 1. Hydrologic units and land use within the San Diego region. Abbreviations are given in Table 1. Dark green is undeveloped open space. Orange is agricultural land. Gray is developed land. Inset shows location of the San Diego region within California.

Sources of data

The primary source of data for this report was collected by the San Diego Regional Water Quality Control Board under the Surface Water Ambient Monitoring Program (SWAMP). In addition, data from the California Department of Fish and Game (CDFG), and NPDES monitoring by San Diego and Orange Counties, Camp Pendleton Marine Corps Base, and the Padre Dam Municipal Water District were obtained (Table 2). All these assessment programs used a targeted design to select sites of interest for sampling. Most sites were selected in order to assess known disturbances in the watershed, although a few undisturbed sites were targeted to set reference expectations for specific studies. A total of 62 sites were sampled under SWAMP for water chemistry and toxicity. Physical habitat was assessed at all but nine of these sites. Bioassessment samples were collected at 144 sites, of which 35 were located at or within 500 m of sites with water chemistry and toxicity data (Figure 2). All four indicators were measured at 29 sites.

Table 2. Summary of data sources used in this analysis. Additional data from each of these

programs was analyzed in watershed-specific reports.

| Program | Years | Watersheds | Indicator | Sites | Samples |
|-----------------------------------|-----------|------------|------------------|-------|---------|
| SWAMP | 2000-2006 | 11 | Bioassessment | 17 | 17 |
| | | | Water chemistry | 62 | 233 |
| | | | Toxicity | 62 | 235 |
| | | | Physical habitat | 53 | 53 |
| California Dept. of Fish and Game | 1998-2005 | 11 | Bioassessment | 98 | 408 |
| San Diego County NPDES | 2002-2005 | 9 | Bioassessment | 45 | 169 |
| Orange County NPDES | 2002-2005 | 1 | Bioassessment | 18 | 87 |
| Camp Pendleton | 2004-2006 | 1 | Bioassessment | 7 | 14 |
| Padre Dam MWD | 2004-2006 | 1 | Bioassessment | 2 | 10 |
| | | | | | |
| All programs | 1998-2006 | 11 | Bioassessment | 144 | 708 |
| | | | Water chemistry | 62 | 233 |
| | | | Toxicity | 62 | 235 |
| | | | Physical habitat | 53 | 53 |

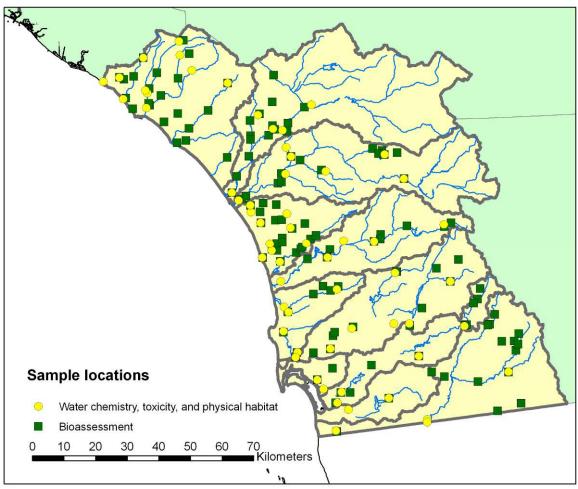


Figure 2. Locations of sampling sites. Yellow circles are sites sampled under SWAMP for water chemistry, toxicity, and physical habitat. Green squares are sites sampled for bioassessment under SWAMP and other programs.

To aggregate data collected under multiple programs, sites within 500 m of each other were treated as a single site. This distance was based on published measures of spatial correlation of benthic communities in streams (Gebler 2004). Although data used in this assessment cover many years (1998-2006), there was little indication that conditions had changed over the course of this study apart from a few sites in the upper Sweetwater watershed, which were affected by the 2003 Cedar Fires (see reports on specific watersheds for details about specific sites).

Indicators

Multiple indicators were used to assess the sites in the San Diego region. Water chemistry, water and sediment toxicity, benthic macroinvertebrate communities, and physical habitat.

Water chemistry was measured as per the SWAMP Quality Assurance Management Plan (QAMP) (Puckett 2002). Measured indicators included physical measures of water quality (e.g., pH, temperature dissolved oxygen, etc.), inorganics, pesticides, polycyclic aromatic hydrocarbons (PAHs), dissolved metals, pesticides, and polychlorinated biphenyls (PCBs). The Appendix contains a complete list of water chemistry constituents.

To evaluate water and sediment toxicity to aquatic life, toxicity assays were conducted on samples from each site as per the SWAMP QAMP (EPA 1993, Puckett 2002). Water toxicity was evaluated with 7-day exposures on the water flea, *Ceriodaphnia dubia*, and 96-hour exposures to the alga *Selenastrum capricornutum*. Both acute and chronic toxicity to *C. dubia* was measured as decreased survival and fecundity (i.e., eggs per female) relative to controls, respectively. Chronic toxicity to *S. capricornutum* was measured as changes in total cell count relative to controls. Sediment toxicity was evaluated with 10-day exposures on the amphipod *Hyallela azteca*. Both acute and chronic toxicity to *H. azteca* was measured as decreased survival and growth (mg per individual) relative to controls, respectively. Chronic toxicity endpoints (i.e., *C. dubia* fecundity, *H. azteca* growth, and *S. capricornutum* total cell count) were used to develop a summary index of toxicity at each site.

Physical habitat was assessed using semi-quantitative observations of 10 components relating to habitat quality, such as embeddedness, bank stability, and width of riparian zone. The assessment protocols are described in The California Stream Bioassessment Procedure (California Department of Fish and Game 2003). Each component was scored on a scale of 0 (highly degraded) to 20 (not degraded). 53 sites were assessed, although data were incomplete at 26 sites. Sites were assessed by the average component score.

To assess the ecological health of the streams in the San Diego region, 708 benthic macroinvertebrate samples were collected at 144 sites. Samples

were collected using SWAMP-comparable protocols, as per the SWAMP QAMP (Puckett 2002). Three replicate samples were collected from riffles at each site; at least 300 individuals were sorted and identified from each replicate, creating a total count of over 900 individuals per site. Using a Monte Carlo simulation, all samples were reduced to 500 count for calculation of the Southern California Index of Biotic Integrity (IBI; Ode et al. 2005), a composite of seven metrics summed and scaled from 0 (poor condition) to 100 (good condition).

A GIS analysis was used to calculate simple landscape metrics for each site. Land use data came from the San Diego Association of Governments, and the Tijuana River Watershed GIS Database (SANDAG 1998, CESAR 2000). Both data sources use compatible procedures for identifying and naming land uses. Land use categories were aggregated into three classes: open space, developed land, and agricultural land. Metrics were calculated for the entire contributing watershed, as well as at a local scale (i.e., within 500 m of the sampling site).

Data Analysis

Water quality was assessed by comparing water chemistry constituents to known thresholds, when possible (SDRWQCB 1994, EPA 1997, CCR 2007). Watersheds were compared by plotting distributions of concentrations of selected constituents. Toxicity was assessed by plotting frequency of samples with endpoints significantly different from controls for each indicator species. Bioassessment samples were assessed by calculating the Southern and Central California Index of Biotic Integrity (IBI, Ode et al. 2005), and comparing samples to a threshold of 40 (i.e., poor or very poor condition versus fair or better condition).

To assess the influence of water chemistry, toxicity, physical habitat, and land use on benthic communities, nonmetric multidimensional scaling (NMS) was used to ordinate bioassessment samples. NMS is an ordination method that represents gradients in community structure in a small number of axes (Kruskal 1964, McCune and Grace 2002). In an NMS plot, ecological distance between samples is represented as distance between points along these axes. Samples are initially placed randomly in ordination space, and the configuration is iteratively adjusted to optimize the ability of the ordination plot to represent ecological distance. The final scaling-stress of the ordination (a measure of how well distances in the ordination plot represent ecological distances between samples) is used to identify the optimal configuration. Only biological data are used to create these axes that define ecological gradients. Subsequently, environmental data (such as water quality, or physical habitat) can be related to the axes using correlation analysis to determine which environmental variables influence community structure. NMS was run using 500-count subsamples were averaged to produce mean abundances for each site because of the lack of synoptic data for many sites. Number of samples per site ranged from 1 to 13.

NMS was run with the following parameters: 1000 runs with real data, 100 runs with randomized data, 4 maximum number of axes, 250 maximum number of iterations, 0.2 step length, 0.000001 stability criterion. Bray-Curtis dissimilarity was used as an ecological distance metric. NMS was run in PC-ORD v 5.12 (McCune and Mefford 2006).

To assess the influence of environmental variables on biotic structure, water chemistry, toxicity, physical habitat, and landscape variables were correlated with NMS axis scores using Spearman's rank correlation (ρ). In addition, these variables were correlated to IBI scores to determine their relationship with biological condition. Correlation strength (based on ρ^2) rather than statistical significance was used to identify strong relationships, as the high number of tests may yield spurious significance and low power. Results for all environmental variables are shown, except for water chemistry variables with ρ^2 < 0.02.

As an additional way to investigate the role of land use on biological condition, sites were plotted on triangular ternary plot, where each axis represents the portion of the watershed with developed, agricultural, or open land use. Ternary plots are used to show the distribution of samples along three dimensions that add up to a constant (e.g., proportional data). Mean IBI scores, as well as frequency of samples in poor or very poor biological condition.

The SWAMP QAMP guided QA/QC for all data collected under SWAMP (See SWAMP QAMP for detailed descriptions of QA/QC protocols, Puckett 2002). QA/QC officers flagged non-compliant physical habitat, water chemistry, toxicity, and tissue results. No data were excluded as a result of QA/QC violations.

4. RESULTS

Assessment of the watersheds

Many sites showed signs of degraded water chemistry. For example, most sites had elevated nutrients, metals, and other constituents. Several of these constituents occurred in concentrations known to harm aquatic life. For example, more than 60% of samples exceeded applicable aquatic life thresholds of 0.025 mg/L of ammonia-N. Exceedances for specific conductivity, sulfate, selenium, and total phosphorus were nearly as frequent. Some anthropogenic organic constituents, such as diazinon, lack thresholds but were detected at many sites (Figure 3). Table 3 shows selected water chemistry constituents in each watershed. The full list of constituents is included in the appendix.

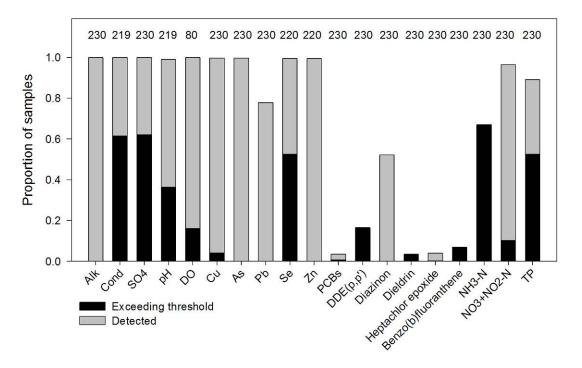


Figure 3. Proportion of samples where water chemistry constituents was detected (gray bars) or exceeded thresholds (black bars). Numbers above bars are numbers of samples. Abbreviations for constituents are in Table 3. Thresholds used for each constituent are shown in Table 3.

Concentrations of many constituents were high in most watersheds, although come values were extreme in sites from the Tijuana watershed that receive surface flows from Mexico (e.g., ammonia-N and phosphorus, Figure 4). Although elevated values were found throughout the region, some samples from the Tijuana River were 1-2 orders of magnitude more concentrated than other samples. Samples with elevated concentrations of many constituents were also found in the San Juan, Carlsbad, and Los Peñasquitos watersheds (Table 3, Appendix).

Table 3. Magnitudes of selected water chemistry constituents. SD = Standard deviation. n = number of samples. Boldface indicates mean values that exceed thresholds. Watershed abbreviations are given in Table 1. * = Thresholds do not apply to all sites in the watershed.

| SJ SM SLR CB | | | | | | | | | | | | | | | | |
|------------------------|--------------------|-------------|-------|--------------|-------|--------|----|-------|-------|----|------|-------|----|-------|--------|----|
| Constituent | Symbol | Threshold L | Jnits | Source | Mean | SD | n | Mean | | n | Mean | | n | Mean | SD | n |
| Physical water quality | | | | | | | | | | | | | | | | |
| Alkalinity as CaCO3 | Alk | 20000 n | ng/l | EPA 2002 | 208 | 109 | 39 | 187 | 49 | 21 | 198 | 84 | 25 | 252 | 43 | 41 |
| Oxygen, Dissolved | DO | 6 n | ng/L | SDRWQCB 1994 | | | 0 | | | 0 | 9.4 | 3.6 | 24 | | | 0 |
| pH | | 6 or 8 p | ρŬ | SDRWQCB 1994 | 7.6 | 1.5 | 38 | 7.4 | 0.8 | 20 | 7.7 | 0.4 | 24 | 7.9 | 1.5 | 40 |
| Salinity | Sal | None p | opt | | 1.05 | 1.77 | 32 | 2.62 | 6.16 | 20 | 0.83 | 0.52 | 24 | | | 0 |
| Specific conductivity | Cond | 1600 µ | uS/cm | CCR 2007 | 2032 | 2839 | 38 | 4399 | 9779 | 20 | 1516 | 968 | 24 | 3800 | 4232 | 40 |
| Sulfate | SO_4 | 250* n | ng/l | SDRWQCB 1994 | 497 | 491 | 39 | 352 | 398 | 21 | 382 | 264 | 25 | 469 | 277 | 41 |
| Nutrients | | | - | | | | | | | | | | | | | |
| Ammonia as N | NH ₃ -N | 0.025 n | mg/l | SDRWQCB 1994 | 0.19 | 0.52 | 39 | 0.02 | 0.05 | 21 | 0.05 | 0.06 | 25 | 0.12 | 0.11 | 41 |
| Nitrate as NO3 | NO_3 | None n | mg/l | | 1.79 | 2.20 | 39 | 22.48 | 19.92 | 20 | | | 0 | 16.00 | | 1 |
| Total Phosphorus as P | TP | 0.1 n | ng/l | SDRWQCB 1994 | 0.22 | 0.27 | 39 | 0.21 | 0.22 | 21 | 0.21 | 0.24 | 25 | 0.14 | 0.09 | 41 |
| Metals . | | | • | | | | | | | | | | | | | |
| Arsenic | As | 50 µ | ug/L | SDRWQCB 1994 | 3.4 | 2.5 | 39 | 2.5 | 3.9 | 21 | 1.3 | 0.9 | 25 | 4.7 | 2.6 | 41 |
| Cadmium | Cd | 5 µ | ug/L | SDRWQCB 1994 | 0.26 | 0.34 | 39 | 0.04 | 0.03 | 21 | 0.03 | 0.02 | 25 | 0.05 | 0.04 | 41 |
| Chromium | Cr | 50 µ | ug/L | SDRWQCB 1994 | 0.25 | 0.22 | 39 | 0.23 | 0.35 | 21 | 0.33 | 0.28 | 25 | 1.06 | 1.05 | 41 |
| Copper | Cu | 9 μ | ug/L | EPA 1997 | 4.05 | 2.85 | 39 | 3.10 | 2.37 | 21 | 4.03 | 2.60 | 25 | 3.55 | 1.50 | 41 |
| Lead | Pb | 2.5 µ | ug/L | EPA 1997 | 0.02 | 0.02 | 39 | 0.01 | 0.01 | 21 | 0.06 | 0.05 | 25 | 0.05 | 0.12 | 41 |
| Manganese | Mn | 5* µ | ug/L | EPA 2002 | 148 | 329 | 39 | 92 | 139 | 21 | 133 | 270 | 25 | 127 | 147 | 41 |
| Nickel | Ni | 52 µ | ug/L | EPA 1997 | 5.55 | 6.76 | 39 | 0.71 | 1.22 | 21 | 0.97 | 2.14 | 25 | 2.16 | 1.37 | 41 |
| Selenium | Se | 5 μ | ug/L | EPA 2002 | 7.5 | 10.4 | 38 | 5.9 | 16.8 | 20 | 4.9 | 4.8 | 24 | 10.6 | 10.1 | 40 |
| Silver | Ag | 3.4 µ | ug/L | EPA 1997 | 0.20 | 1.22 | 39 | 0.09 | 0.29 | 21 | 0.07 | 0.34 | 25 | 0.05 | 0.28 | 41 |
| Zinc | Zi | 120 µ | ug/L | EPA 2002 | 4.1 | 3.1 | 38 | 2.3 | 1.5 | 20 | 2.7 | 1.9 | 24 | 6.5 | 6.4 | 40 |
| Organics | | | | | | | | | | | | | | | | |
| Benzo(b)fluoranthene | | 0.0044 n | _ | EPA 2002 | 3.5 | 9.6 | | 1.0 | 3.2 | | 0 | 0 | | 0 | 0 | 41 |
| PCBs | | 0.014 n | ng/L | EPA 2002 | 2.45 | 6.68 | 39 | 0 | 0 | 21 | 0 | | 25 | 0 | | 41 |
| Diazinon | | | ng/L | | 43.97 | 103.77 | | 7.78 | 18.10 | | 0.67 | | | 68.39 | 101.40 | 40 |
| DDE(p,p') | | 0.00059 n | ng/L | EPA 2002 | 0.29 | 0.66 | 39 | 0.90 | 2.64 | | 0.04 | 0.20 | | 1.37 | 2.24 | 41 |
| DDTs | | | ng/L | | 0.58 | 1.28 | | 1.48 | 3.89 | | 0.04 | 0.20 | | 1.93 | 2.79 | |
| Dieldrin | | 0.00014 n | ng/L | EPA 2002 | 0.15 | 0.42 | 39 | 0 | 0 | 21 | 0 | 0 | 25 | 0.02 | 0.16 | 41 |
| Disulfoton | | | ng/L | | 3.95 | 10.28 | 38 | 0 | 0 | 20 | 0 | 0 | 24 | 33.33 | 38.29 | 40 |
| Heptachlor epoxide | | 0.0038 n | ng/L | EPA 1997 | 0.21 | 0.70 | | 0.10 | 0.30 | | 0 | 0 | _ | 0 | _ | 41 |
| Secbumeton | | None n | ng/L | | 3.11 | 13.48 | 38 | 1.75 | 7.83 | 20 | 4.50 | 12.50 | 24 | 85.00 | 153.57 | 40 |

Table 3, continued.

| Table 3, continued. | | | | | | STO | | | LP | | | SD | |
|------------------------|----------|----------------|----------|----------|-------|-------|----|--------|--------|----|-------|-------|------|
| Constituent | Symbol | Threshold Unit | s Source | | Mean | SD | n | Mean | SD | n | Mean | SD | n |
| Physical water quality | | | | | | | | | | | | | |
| Alkalinity as CaCO3 | Alk | 20000 mg/ | EPA 20 | 02 | 243 | 129 | 18 | 221 | 88 | 20 | 242 | 85 | 27 |
| Oxygen, Dissolved | DO | 6 mg/ | SDRW | QCB 1994 | | | 0 | | | 0 | 9.5 | 3.0 | 26 |
| рН | | 6 or 8 pH | SDRW | QCB 1994 | 8.0 | 0.4 | 17 | 7.9 | 0.4 | 19 | 8.0 | 0.3 | 26 |
| Salinity | Sal | None ppt | | | 2.53 | 5.47 | 17 | | | 0 | 1.00 | 0.45 | 26 |
| Specific conductivity | Cond | 1600 µS/d | m CCR 20 | 07 | 4316 | 8781 | 17 | 2981 | 1256 | 19 | 1872 | 830 | 26 |
| Sulfate | SO_4 | 250* mg/ | SDRW | QCB 1994 | 358 | 367 | 18 | 674 | 407 | 20 | 283 | 159 | 27 |
| Nutrients | | | | | | | | | | | | | |
| Ammonia as N | NH_3-N | 0.025 mg/ | SDRW | QCB 1994 | 0.10 | 0.12 | 18 | 0.09 | 0.06 | 20 | 0.05 | 0.05 | 27 |
| Nitrate as NO3 | NO_3 | None mg/ | | | 1.96 | 2.58 | 17 | 0 | | 1 | | | 0 |
| Total Phosphorus as P | TP | 0.1 mg/ | SDRW | QCB 1994 | 0.24 | 0.35 | 18 | 0.10 | 0.16 | 20 | 0.12 | 0.09 | 27 |
| Metals | | | | | | | | | | | | | |
| Arsenic | As | 50 μg/L | SDRW | QCB 1994 | 2.1 | 1.7 | 18 | 3.4 | 0.8 | 20 | 3.2 | 2.4 | - 27 |
| Cadmium | Cd | 5 µg/L | SDRW | QCB 1994 | 0.03 | 0.03 | 18 | 0.02 | 0.01 | 20 | 0.03 | 0.02 | 27 |
| Chromium | Cr | 50 μg/L | SDRW | QCB 1994 | 0.18 | 0.17 | 18 | 0.89 | 0.97 | 20 | 0.64 | 0.74 | 27 |
| Copper | Cu | 9 μg/L | EPA 19 | 97 | 2.41 | 1.76 | 18 | 4.03 | 1.58 | 20 | 4.16 | 2.06 | 27 |
| Lead | Pb | 2.5 µg/l | EPA 19 | 97 | 0.03 | 0.03 | 18 | 0.05 | 0.08 | 20 | 0.09 | 0.08 | 27 |
| Manganese | Mn | 5* μg/L | EPA 20 | 02 | 135 | 135 | | 141 | 156 | 20 | 60 | | 27 |
| Nickel | Ni | 52 μg/L | EPA 19 | 97 | 0.70 | 0.79 | 18 | 3.38 | 3.55 | 20 | 1.15 | 1.92 | |
| Selenium | Se | 5 µg/l | EPA 20 | 02 | 3.7 | 5.4 | 17 | 7.8 | 3.7 | 19 | 8.1 | 6.8 | 26 |
| Silver | Ag | 3.4 µg/l | | 97 | 0.00 | 0.00 | | 0.06 | 0.25 | | 0.00 | 0.00 | |
| Zinc | Zi | 120 µg/l | EPA 20 | 02 | 2.2 | 1.7 | 17 | 8.4 | 8.7 | 19 | 3.9 | 2.3 | 26 |
| Organics | | | | | | | | | | | | | |
| Benzo(b)fluoranthene | | 0.0044 ng/L | | | 3.0 | 5.9 | | 0 | 0 | | 0.8 | 4.0 | 27 |
| PCBs | | 0.014 ng/L | | 02 | 0 | 0 | | 0 | 0 | | 0 | 0 | |
| Diazinon | | None ng/L | | | 12.64 | 14.90 | | 50.81 | 51.65 | | 1.73 | 5.50 | |
| DDE(p,p') | | 0.00059 ng/L | | 02 | 0.17 | 0.51 | | 5.30 | 12.71 | 20 | 0 | 0 | |
| DDTs | | None ng/L | | | 0.28 | 0.96 | | 5.65 | 12.75 | | 0 | 0 | |
| Dieldrin | | 0.00014 ng/L | EPA 20 | 02 | 0.11 | 0.32 | | 0 | 0 | | 0 | 0 | |
| Disulfoton | | None ng/L | | | 0 | | 17 | 52.92 | 64.68 | | 0 | 0 | |
| Heptachlor epoxide | | 0.0038 ng/L | EPA 19 | 97 | 0.11 | 0.32 | | 0 | 0 | 20 | 0 | 0 | |
| Secbumeton | | None ng/L | | | 0 | 0 | 17 | 131.16 | 147.98 | 19 | 12.92 | 48.18 | 26 |

Table 3, continued.

| Table 3, continued. | | | | F | PSD | | | SW | | | ОТ | | TJ | |
|------------------------|--------------------|-----------------|--------------|-------|-------|---|------|-------|----|-------|------|---------------|-------|------|
| Constituent | Symbol | Threshold Units | Source | Mean | SD | n | Mean | SD | n | Mean | SD | n Mean | SD | n |
| Physical water quality | | | | | | | | | | | | | | |
| Alkalinity as CaCO3 | Alk | 20000 mg/l | EPA 2002 | 191 | 15 | 5 | 191 | 92 | 15 | 233 | 21 | 7 400 | 164 | 12 |
| Oxygen, Dissolved | DO | 6 mg/L | SDRWQCB 1994 | 113.3 | 198.9 | 4 | 9.2 | 2.0 | 14 | | | 0 8.3 | 4.3 | 12 |
| рН | | 6 or 8 pH | SDRWQCB 1994 | 8.8 | 0.8 | 4 | 8.0 | 0.4 | 14 | 7.8 | 0.3 | 5 7.9 | 0.7 | 12 |
| Salinity | Sal | None ppt | | 4.19 | 5.66 | 4 | 1.85 | 1.57 | 14 | 1.30 | 0.85 | 5 0.76 | 0.46 | 12 |
| Specific conductivity | Cond | 1600 µS/cm | CCR 2007 | 6925 | 9619 | 4 | 3930 | 3649 | 14 | 2478 | 1539 | 5 1482 | 887 | 12 |
| Sulfate | SO_4 | 250* mg/l | SDRWQCB 1994 | 437 | 394 | 5 | 276 | 190 | 15 | 217 | 40 | 7 201 | 133 | 12 |
| Nutrients | | | | | | | | | | | | | | |
| Ammonia as N | NH ₃ -N | 0.025 mg/l | SDRWQCB 1994 | 0.05 | 0.05 | 5 | 0.06 | 0.05 | 15 | 0.08 | 0.10 | 7 11.33 | 16.63 | 12 |
| Nitrate as NO3 | NO_3 | None mg/l | | 0 | | 1 | | | 0 | 9.21 | 9.14 | 7 | | 0 |
| Total Phosphorus as P | TP | 0.1 mg/l | SDRWQCB 1994 | 0.25 | 0.07 | 5 | 0.06 | 0.04 | 15 | 0.01 | 0.02 | 7 3.31 | 3.70 | 12 |
| Metals | | · · | | | | | | | | | | | | |
| Arsenic | As | 50 μg/L | SDRWQCB 1994 | 2.5 | 1.5 | 5 | 11.5 | 16.8 | 15 | 7.7 | 6.0 | 7 3.7 | 2.6 | 12 |
| Cadmium | Cd | 5 µg/L | SDRWQCB 1994 | 0.08 | 0.05 | 5 | 0.02 | 0.02 | 15 | 0.02 | 0.01 | 7 0.06 | 0.03 | 12 |
| Chromium | Cr | 50 μg/L | SDRWQCB 1994 | 1.22 | 0.79 | 5 | 0.96 | 1.06 | 15 | 0.37 | 0.30 | 7 2.67 | 2.94 | - 12 |
| Copper | Cu | 9 μg/L | EPA 1997 | 8.23 | 3.94 | 5 | 4.25 | 3.05 | 15 | 2.91 | 1.16 | 7 5.14 | 6.25 | 12 |
| Lead | Pb | 2.5 µg/L | EPA 1997 | 0.51 | 0.28 | 5 | 0.07 | 0.13 | 15 | 0.02 | 0.02 | 7 0.25 | 0.27 | 12 |
| Manganese | Mn | 5* μg/L | EPA 2002 | 61 | 67 | 5 | 54 | 71 | 15 | 41 | 67 | 7 238 | 228 | 12 |
| Nickel | Ni | 52 μg/L | EPA 1997 | 3.99 | 2.81 | 5 | 0.78 | 0.93 | 15 | 1.80 | 3.22 | 7 9.16 | 11.13 | 12 |
| Selenium | Se | 5 μg/L | EPA 2002 | 77.5 | 115.8 | 4 | 26.6 | 27.8 | 14 | 9.2 | 7.3 | 6 7.2 | 4.6 | 12 |
| Silver | Ag | 3.4 µg/L | EPA 1997 | 0.13 | 0.27 | 5 | 0.00 | 0.00 | 15 | 0.39 | 1.02 | 7 0.02 | 0.04 | 12 |
| Zinc | Zi | 120 µg/L | EPA 2002 | 13.6 | 4.1 | 4 | 2.9 | 2.0 | 14 | 2.0 | 8.0 | 6 4.5 | 6.6 | 12 |
| Organics | | | | | | | | | | | | | | |
| Benzo(b)fluoranthene | | 0.0044 ng/L | EPA 2002 | 5.8 | 13.0 | 5 | 0 | | 15 | 0 | 0 | | 30.3 | 12 |
| PCBs | | 0.014 ng/L | EPA 2002 | 0 | 0 | 5 | 0 | 0 | 15 | 0 | 0 | 7 0 | 0 | 12 |
| Diazinon | | None ng/L | | 26.25 | 29.17 | | | 11.19 | | 20.83 | - | | 20.93 | |
| DDE(p,p') | | 0.00059 ng/L | EPA 2002 | 0 | 0 | 5 | 0 | | 15 | 1.43 | 2.51 | 7 0 | 0 | 12 |
| DDTs | | None ng/L | | 0 | 0 | 5 | 0 | 0 | 15 | 2.00 | 2.83 | | 0 | 12 |
| Dieldrin | | 0.00014 ng/L | EPA 2002 | 0 | 0 | - | 0 | 0 | 15 | 0 | 0 | 7 0 | 0 | 12 |
| Disulfoton | | None ng/L | | 13.25 | 26.50 | | 4.14 | 15.50 | 14 | 0 | 0 | 6 16.17 | 40.76 | |
| Heptachlor epoxide | | 0.0038 ng/L | EPA 1997 | 0 | 0 | 5 | 0 | 0 | 15 | 0 | 0 | 7 0 | 0 | 12 |
| Secbumeton | | None ng/L | | | | 0 | | | 0 | 0 | 0 | 6 | | 0 |

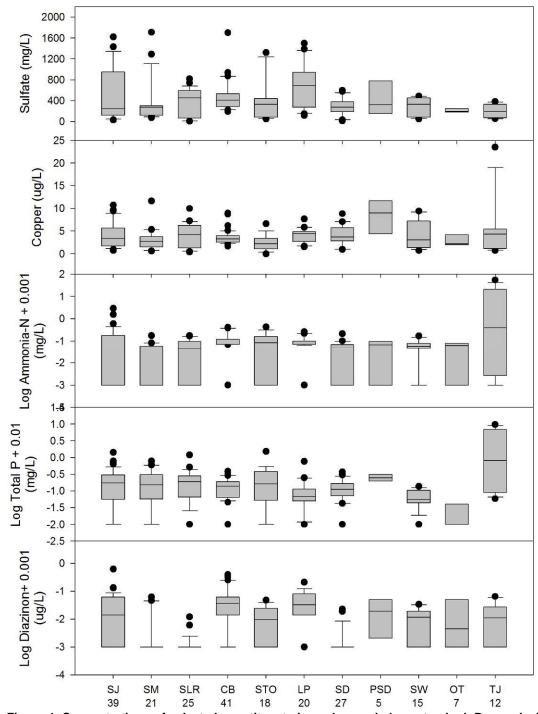


Figure 4. Concentrations of selected constituents in each sample by watershed. Box and whiskers represent 5th, 25th, 50th, 75th, and 95th percentiles. Dots represent values above the 95th or below the 5th percentile. Numbers below the X-axis represent the number of samples from that watershed for all constituents, except for diazinon (which had one less sample in each watershed). Watershed abbreviations are given in Table 1.

Water and sediment toxicity was frequently observed throughout the region, although certain indicator species were more sensitive than others. For example, 59% of all water samples were toxic to the algae *S. capricornutum*. The two arthropod indicators were less sensitive, with the amphipod *H. azteca* suffering increased mortality when exposed to 27% of sediment samples, and the water flea *C. dubia* showing reduced fecundity when exposed to 34% of water samples (Figure 5).

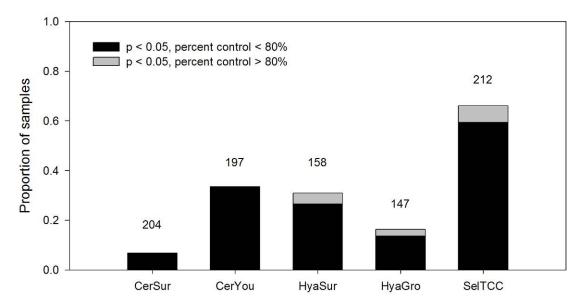


Figure 5. Frequency of toxicity for each endpoint and indicator species are shown in the black bars. Toxicity was determined if sample endpoints were less than 80% of controls, and the difference was significant at the 0.05 level. Weaker (but still significant) results are shown in the gray bars. Empty space above the bars indicate the proportion of samples not indicating toxicity. Numbers above bars indicate number of samples.

Toxicity was observed in every watershed. The frequency of toxicity to chronic endpoints (i.e., *C. dubia* young per female, *H. azteca* growth, and *S. capricornutum* total cell count) ranged from 24% of samples in the Carlsbad watershed to 90% of samples in the Tijuana watershed (Table 4).

Table 4. Mean percent control of each toxic indicator in each watershed. SD = standard deviation. n = number of samples. Freq = frequency of toxicity. -- = No toxicity detected.

| | | _ | | | | | | | | | | | | | | | | | | | | |
|-----------|------|----------|------|----|------|-------|-------|----|-----------------|-----------|------|----|------|----|------|-----------------|------|------------------|------|----|----------|-----|
| | | C. dubia | | | | | | | | H. azteca | | | | | | | | S. capricornutum | | | | |
| | - 5 | Surv | ival | | You | ıng/f | emale | • | Survival Growth | | | | | | wth | Total cell cour | | | | t | Indicato | |
| Watershed | Mean | SD | Freq | n | Mean | SD | Freq | n | Mean | SD | Freq | n | Mean | SD | Freq | n | Mean | SD | Freq | n | Freq | n |
| SJ | 89 | 27 | 0.11 | 35 | 92 | 26 | 0.35 | 34 | 153 | 193 | 0.29 | 25 | 80 | 34 | 0.12 | 28 | 64 | 35 | 0.64 | 39 | 0.32 | 161 |
| SM | 93 | 22 | 0.05 | 21 | 95 | 23 | 0.20 | 20 | 106 | 34 | 0.00 | 13 | 101 | 6 | 0.15 | 13 | 96 | 61 | 0.52 | 21 | 0.20 | 88 |
| SLR | 98 | 6 | 0.00 | 34 | 80 | 26 | 0.32 | 34 | 99 | 19 | 0.00 | 16 | 106 | 5 | 0.06 | 16 | 82 | 41 | 0.50 | 32 | 0.21 | 132 |
| CB | 96 | 34 | 0.10 | 39 | 87 | 43 | 0.44 | 36 | 121 | 62 | 0.45 | 40 | 68 | 34 | 0.13 | 40 | 70 | 30 | 0.55 | 40 | 0.33 | 195 |
| STO | 91 | 25 | 0.06 | 16 | 92 | 25 | 0.20 | 15 | 130 | 31 | 0.00 | 7 | 101 | 6 | 0.00 | 8 | 61 | 38 | 0.88 | 17 | 0.30 | 63 |
| LP | 95 | 34 | 0.16 | 19 | 104 | 56 | 0.33 | 18 | 129 | 79 | 0.38 | 16 | 82 | 22 | 0.25 | 16 | 56 | 27 | 0.84 | 19 | 0.40 | 88 |
| SD | 101 | 4 | | 30 | 82 | 26 | 0.37 | 30 | 94 | 39 | 0.21 | 11 | 92 | 28 | 0.36 | 14 | 82 | 44 | 0.45 | 29 | 0.28 | 114 |
| PSD | 42 | 60 | 0.50 | 2 | 45 | 64 | 0.50 | 2 | 110 | 39 | 0.00 | 2 | 109 | 2 | 0.00 | 2 | 57 | 35 | 0.50 | 2 | 0.30 | 10 |
| SW | 108 | 4 | | 2 | 123 | 17 | | 2 | 86 | 15 | 0.14 | 5 | 102 | 14 | 0.20 | 7 | 121 | 100 | 0.50 | 4 | 0.21 | 20 |
| OT | 96 | 9 | | 5 | 84 | 28 | 0.40 | 5 | 106 | 36 | 0.50 | 3 | 81 | 24 | 0.33 | 4 | 66 | 26 | 0.71 | 7 | 0.41 | 24 |
| TJ | 105 | | | 1 | 103 | | | 1 | 125 | 26 | 0.40 | 9 | 70 | 48 | 0.00 | 10 | 64 | 79 | 0.50 | 2 | 0.20 | 23 |

Physical habitat ranged from very poor to very good, although the majority of sites showed some signs of degradation. Every watershed contained some sites in good condition, except for watersheds where few sites were assessed. For example, all sites in the Santa Margarita watershed were in very good condition, with mean physical habitat scores greater than 15 (Figure 6, Table 5).

Some components of physical habitat were more often degraded than other components. For example, a large majority of sites had poor scores (< 5) for embeddedness. Degradation of velocity-depth regimes were nearly as bad, with the majority of sites scoring below 10. In contrast, sediment deposition, channel flow, and bank stability were in good condition (score > 15) at the majority of sites (Figure 7, Table 5).

Table 5. Physical habitat scores at sampled sites. Symbols above the columns indicate the watershed. Numbers indicate number of sites assessed within each watershed. SD = standard deviation. Watershed abbreviations are given in Table 1.

| | | SJ | S | M | SLI | ₹ | CE | 3 | ST | O . | LP | | SD |) | SW | TO | | TJ |
|-----------------------|----------|---------|--------|-------|------|-----|------|-----|------|-----|--------|-----|------|-----|------|------|----|------|
| | | 11 | | 4 | 6 | | 10 | | 5 | | 6 | | 7 | | 1 | 2 | | 1 |
| Component | Symbol | Mean SI |) Meai | n SD | Mean | SD | Mean | SD | Mean | SD | Mean S | SD | Mean | SD | Mean | Mean | SD | Mean |
| Mean score | AvePHAB | 10.5 4. | 7 15.3 | 3 0.4 | 13.3 | 2.0 | 11.4 | 2.6 | 14.7 | 3.1 | 10.9 4 | 4.7 | 11.1 | 3.2 | | 8.7 | 5 | |
| Epifaunal cover | EpiCov | 11.0 5. | 1 16.3 | 3 2.2 | 13.3 | 3.2 | 12.1 | 5.6 | 13.8 | 6.0 | 10.7 | 7.3 | 10.3 | 4.6 | 13 | 8.5 | 6 | 10 |
| Embeddedness | Embed | 9.8 6. | 3. | 5 1.9 | 5.5 | 6.4 | 3.5 | 4.1 | 5.0 | 5.7 | 6.7 | 7.8 | 5.4 | 7.0 | 13 | 3 | 1 | 3 |
| Velocity-Depth Regime | VelDep | 8.2 4. | 4 14.8 | 3 2.5 | 10.0 | 1.1 | 12.1 | 4.6 | 11.2 | 4.0 | 10.3 4 | 4.5 | 7.6 | 4.0 | 8 | 5.5 | 4 | 9 |
| Sediment Deposition | SedDep | 11.0 6. | 2 13. | 5 1.7 | 15.2 | 1.3 | 13.8 | 4.9 | 17.2 | 0.8 | 12.7 8 | 8.2 | 15.6 | 4.8 | 13 | 8.5 | 8 | 5 |
| Channel Flow | ChanFlo | 11.3 6. | 0 17.3 | 3 2.1 | 15.3 | 3.9 | 13.9 | 5.4 | 16.0 | 6.2 | 12.8 | 5.5 | 14.7 | 4.3 | 13 | 8 | 6 | 18 |
| Channel Alteration | ChanAlt | 11.9 7. | 7 19.3 | 3 1.0 | 14.0 | 2.4 | 7.8 | 6.4 | 17.6 | 1.9 | 10.5 | 7.1 | 11.3 | 5.9 | 18 | 6.5 | 9 | 19 |
| Riffle Frequency | RifFreq | 11.2 7. | 3 17.8 | 3 1.7 | 12.2 | 4.7 | 10.2 | 6.0 | 13.2 | 7.9 | 11.0 | 7.0 | 7.9 | 6.1 | 13 | 8.5 | 9 | 16 |
| Bank Stability | BankStab | 11.0 7. | 5 16. | 5 2.4 | 14.5 | 2.1 | 16.3 | 5.4 | 18.0 | 3.4 | 12.8 | 5.4 | 15.3 | 2.8 | 18 | 16.5 | 4 | 16 |
| Vegetative Protection | VegPro | 11.4 6. | 8 17.8 | 3 1.3 | 15.3 | 2.9 | 15.4 | 5.3 | 19.2 | 1.8 | 12.8 | 5.3 | 11.3 | 3.3 | 18 | 14 | 6 | 15 |
| Riparian Vegetation | RipVeg | 10.5 8. | 2 16. | 5 2.4 | 15.3 | 3.8 | 8.9 | 4.7 | 15.4 | 6.3 | 8.2 (| 6.6 | 11.7 | 3.4 | 18 | 8 | 8 | 19 |

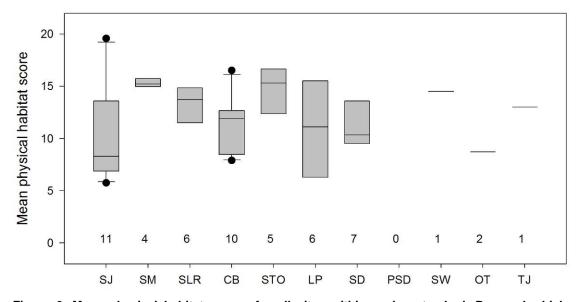


Figure 6. Mean physical habitat scores for all sites within each watershed. Box and whiskers represent 5th, 25th, 50th, 75th, and 95th percentiles. Points represent values above the 95th and below the 5th percentiles. No sites in the Pueblo San Diego watershed were assessed. Numbers indicate number of sites assessed in each watershed. Watershed abbreviations are given in Table 1.

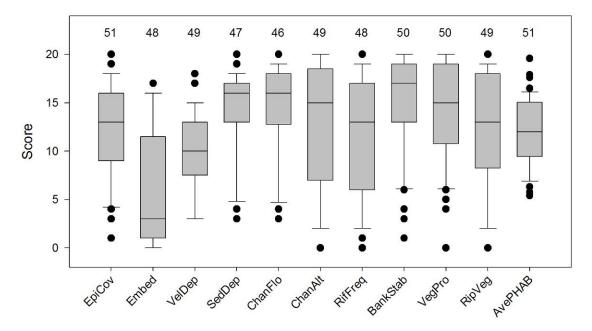


Figure 7. Scores for each component of physical habitat. Abbreviations are given in Table 5. Box and whiskers represent 5th, 25th, 50th, 75th, and 95th percentiles. Points represent values above the 95th and below the 5th percentiles. Numbers indicate number of sites assessed for each component. Component abbreviations are given in Table 5.

IBI scores covered nearly the entire range of the index, with a low score of 0 and a high score of 93 (Table 6). However, the overwhelming majority (80%) of the 708 samples were below the impairment threshold of 39. Poor conditions were observed in every sample from 61% of sites. Good conditions (IBI > 39) were observed in samples from 39% of sites. Although samples in poor condition were found in every watershed, samples in good condition were absent from smaller coastal watersheds, like Carlsbad, Los Peñasquitos, and Pueblo San Diego. A majority of samples (65%) from the Tijuana watershed were in good condition, as were a near-majority (44%) in the Santa Margarita watershed (Figure 8). Sites with samples in good condition were largely restricted to the interior mountains of the larger watersheds. However, a few samples in good condition were occasionally detected in smaller coastal watersheds in undeveloped portions of southern Orange County and in the Camp Pendleton Marine Corps Base (Figure 9).

Table 6. Bioassessment scores by watershed. SD = Standard deviation. Frequency = frequency of samples in poor condition (i.e., IBI < 40). Watershed abbreviations are given in Table 1.

| | | | | IBI | | |
|-----------|-------|---------|-----------|------|----|-----------|
| Watershed | Sites | Samples | Years | Mean | SD | Frequency |
| SJ | 26 | 132 | 1998-2005 | 25 | 18 | 0.83 |
| SM | 16 | 113 | 1998-2006 | 37 | 16 | 0.56 |
| SLR | 16 | 75 | 1998-2006 | 28 | 20 | 0.73 |
| CB | 22 | 125 | 1998-2005 | 13 | 8 | 1.00 |
| STO | 9 | 27 | 2000-2005 | 24 | 13 | 0.89 |
| LP | 10 | 62 | 1998-2006 | 16 | 8 | 1.00 |
| SD | 20 | 75 | 1996-2006 | 17 | 13 | 0.96 |
| PSD | 2 | 7 | 2003-2005 | 18 | 6 | 1.00 |
| SW | 10 | 46 | 1998-2005 | 28 | 20 | 0.74 |
| OT | 2 | 3 | 2000-2001 | 25 | 18 | 0.67 |
| TJ | 16 | 43 | 1999-2006 | 43 | 18 | 0.35 |

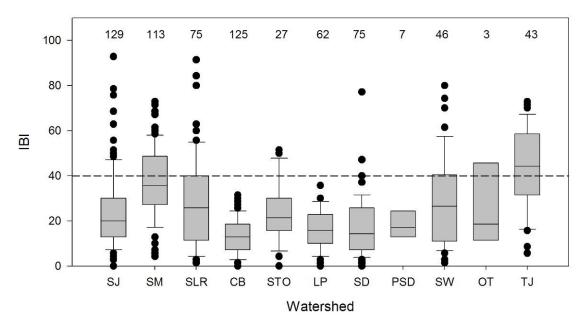


Figure 8. Boxplot of IBI scores of bioassessment samples in each watershed. Boxes and whiskers represent 5th, 25th, 50th, 75th, and 95th percentiles; points represent scores above the 95th percentile or below the 5th percentile. The dashed line represents the threshold for impaired conditions (i.e., 40). Watershed abbreviations are given in Table 1.

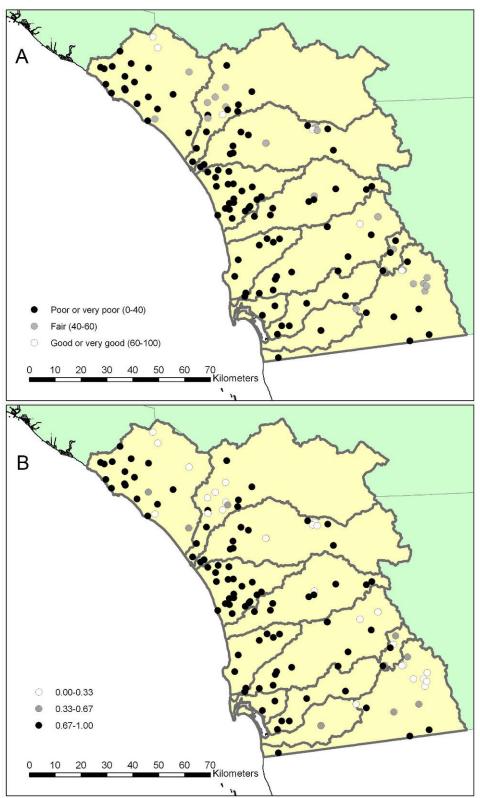


Figure 9. Distribution of IBI scores in the San Diego region. A) Mean IBI scores at each site. B) Frequency of samples in poor or very poor condition (i.e., IBI < 40).

Stressor relationships

Inspection of scatterplots revealed that some stressor variables had a strong influence on the IBI. For example, IBI scores were never above 30 where more than 20% of the watershed was developed. Similar relationships were observed with other variables, such as arsenic concentration, frequency of acute toxicity to *C. dubia*, sulfate, and arsenic. In contrast, more linear relationships were observed for other variables, such as distance from coast. Wedge-shaped relationships, were observed for several nutrients, such as ammonia-N and total phosphorus (Figure 10). Several variables showed no discernible relationship, such as dissolved oxygen and frequency of toxicity to *S. capricornutum*.

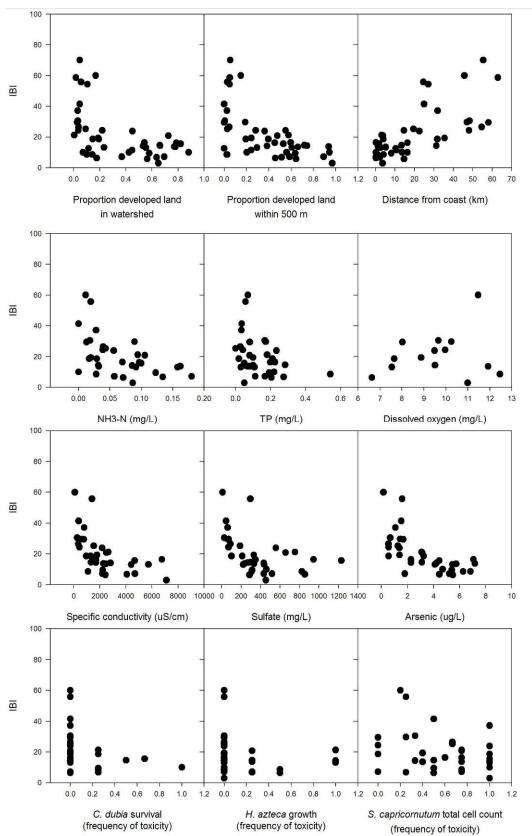


Figure 10. Relationships between IBI scores and landscape, nutrient, water quality, and toxicity variables.

Land use was associated with elevated levels of some contaminants, suggesting that development within the watershed may be a mechanism for many of the observed stressors on aquatic life. For example, specific conductivity was elevated at more developed sites (Figure 11). However, most relationships between land use and environmental stressors were weak and not statistically significant.

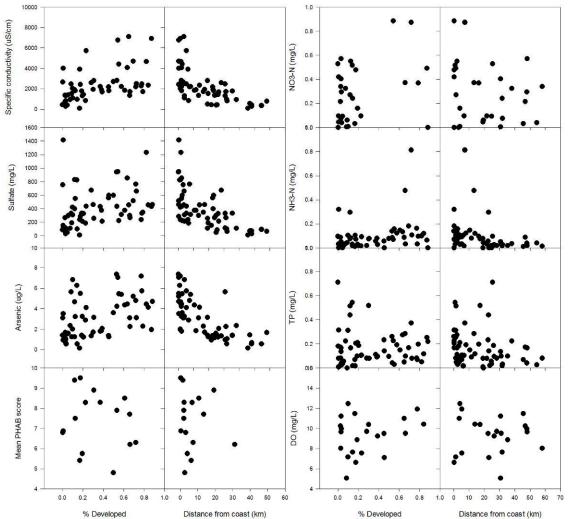


Figure 11. Relationship between two land use variables (% developed land in watershed, and distance from coast) versus selected environmental stressor variables.

Ternary plots further suggested a strong role for development in the watershed as limiting biological health (Figure 12). For example, sites in fair or good condition (mean IBI > 40) were tightly clustered on the right side of plots, where developed land was lowest. Similar patterns were observed at both watershed-wide and local scales.

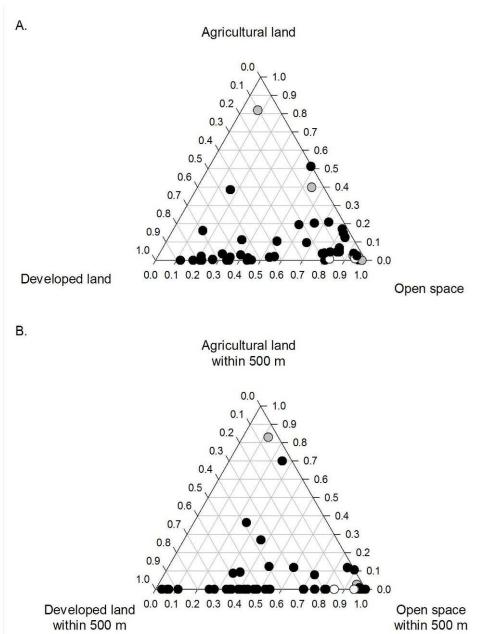


Figure 12. Ternary plots of sites showing land use in A) the contributing watershed, and B) within 500 m. Black dots indicate sites in poor or very poor condition (mean IBI 0 - 40). Gray dots indicate sites in fair condition (mean IBI 40 - 60). White dots indicate sites in good condition (mean IBI 60 - 100).

Ordination of mean taxa abundances yielded a three-axis solution. This ordination was a good representation of the biological data, as indicated by the low final scaling-stress (i.e., 17.6). The three axes combined represented 82.6% of the variance in the site-by-taxa matrix, with the third axis representing the largest portion (40%) of this variance, followed by axis 2 (23.9%) and axis 1 (18.7%). No clustering of sites in ordination plots by watershed was evident (Figure 13).

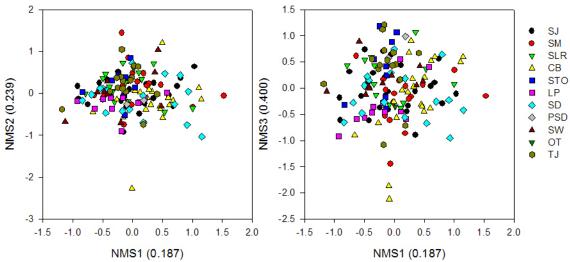


Figure 13. NMS Ordinations of sites in the San Diego region. Each point represents the ordination of the mean abundance of all samples collected at that site. Symbols represent different watersheds. Final scaling-stress was 17.6. Numbers in the axis titles is the proportion of variability represented by the axis.

Examination of weighted scores for taxa showed that certain species drove the gradients observed in the data. For example, several pollution-sensitive taxa (particularly mayflies, stoneflies, and caddisflies) were located high on axis 2, while several non-insects were low on axis 2. However, there was considerable diversity within all taxonomic groups; for example, the caddisfly *Hydroptila* had relatively low score of -0.16 on axis 2, and the dipteran *Dasyhelea* had a high score of 0.57. Dipterans, like *Simulium* and Muscidae dominated the low end of axis 1, and non-insects, such as Oligochaeta, Cladocera, Corbicula, and Nematoda were at the high end. No obvious pattern characterized axis 3; the stonefly *Malenka* and the dipteran *Dasyhelea* were at the high end, while several caddisflies *Ochrotrichia*, the mayfly *Baetis*, and the clam *Corbiculum* were at the low end. (Figure 14).

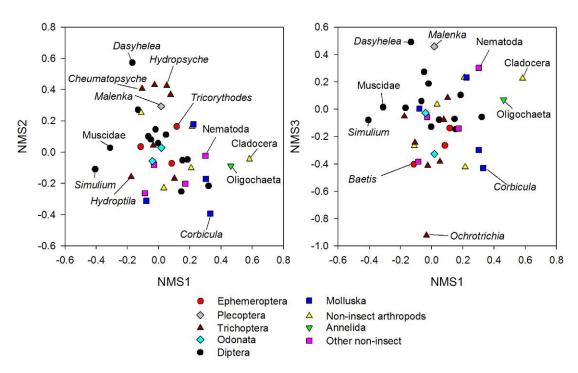


Figure 14. Weighted averages of selected taxa in ordination space. Symbols represent taxonomic groups. Only taxa appearing in 40 or more sites are shown.

Correlation of ordination axes with IBI and metric scores showed that most metrics responded strongly to axis 2 (Figure 15). In general, higher values on axis 2 corresponded to better ecological condition. For example, the IBI as well as the metrics EPT richness and % intolerant individuals had strong positive correlations with axis 2 (ρ^2 of 0.50, 0.50, and 0.54 respectively), and % noninsect taxa had a moderately strong negative relationship (ρ^2 = 0.35). Two metrics (i.e., % collectors and % tolerant taxa) showed no strong relationships with any axis, perhaps because of the ubiquity of collector and tolerant taxa at both disturbed and undisturbed sites. No metric showed strong relationships with axis 1 or 3 (Table 7, Figure 15). However, the IBI showed a unimodal relationship with axis 1 (Figure 16).

Table 7. Correlations of biological metrics with NMS axes. n = Number of sites used to calculate correlations.

| | | Spearman rank correlations (ρ) | | | | | | | | | |
|------------------------------|----------|--------------------------------|-------|-------|----|--|--|--|--|--|--|
| Biological metric | Symbol | NMS1 | NMS2 | NMS3 | n | | | | | | |
| Index of biotic integrity | IBI | -0.04 | 0.70 | 0.02 | 44 | | | | | | |
| Frequeny of impaired samples | | -0.10 | -0.60 | -0.07 | 44 | | | | | | |
| EPT Taxa | EPT tx | 0.16 | 0.58 | 0.13 | 44 | | | | | | |
| Coleoptera Taxa | Coleo tx | -0.11 | 0.71 | -0.08 | 44 | | | | | | |
| Predator Taxa | Pred tx | -0.03 | 0.53 | -0.08 | 44 | | | | | | |
| % Collectors | % Coll | -0.12 | -0.15 | 0.11 | 44 | | | | | | |
| % Intolerant | % Intol | 0.02 | 0.74 | 0.20 | 44 | | | | | | |
| % Non-Insecta Taxa | % NI tx | 0.33 | -0.59 | -0.04 | 44 | | | | | | |
| % Tolerant Taxa (8-10) | % Tol tx | 0.22 | -0.39 | 0.09 | 44 | | | | | | |

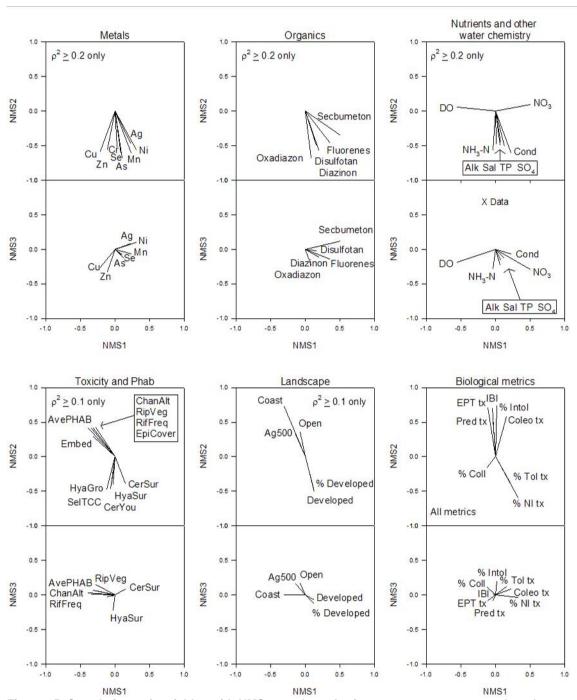


Figure 15. Correlations of variables with NMS axes. Length of vectors represent strength and direction of correlation, as measured by Spearman's rank correlation coefficient (ρ).

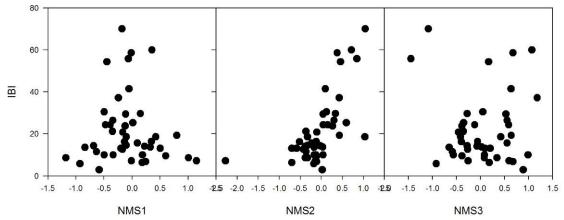


Figure 16. Relationship between IBI scores and ordination axes.

Correlation analyses revealed that many environmental variables related to water chemistry were correlated with axis 2. For example, many metals, ammonia-N, total phosphorus, specific conductivity, sulfate, and several organic constituents (such as diazinon and secbumeton) had strong negative correlations (Spearman's $\rho^2 > 0.2$) with this axis (Table 8). In addition, frequency of toxicity for all endpoints were negatively correlated with this axis, further suggesting that axis 2 represents a toxic contamination gradient (Table 9). In contrast, only two variables (nitrate-N and dissolved oxygen) were strongly correlated with axis 1, suggesting that this axis may represent a nutrient enrichment gradient. No water chemistry or toxicity variables were strongly correlated with axis 3 (Figure 15). Many of the variables that had strong correlations with axis 2 also had significant correlations with the IBI (e.g., arsenic, ammonia-N, etc., Table 8-9).

Table 8. Correlations of selected water chemistry constituents with NMS axes and IBI. N = number of sites used to calculate correlations. A) Physical water quality, metals, and nutrients. B) PAHs, PCBs, and pesticides.

| PCBs, and pesticides. | | | | | | |
|--|----------|--------------------------------|-------|-------|-------|----|
| | | Spearman rank correlations (ρ) | | | | |
| A. Water qualityNon-organic constituents | Symbol | NMS1 | NMS2 | NMS3 | IBI | n |
| Physical water quality and inorganics | | | | | | |
| Alkalinity as CaCO3 | Alk | 0.02 | -0.48 | 0.00 | -0.47 | 35 |
| Sulfate | SO_4 | 0.14 | -0.51 | -0.14 | -0.58 | 35 |
| Oxygen, Dissolved | DO | -0.56 | 0.05 | -0.19 | 0.10 | 14 |
| рН | | -0.32 | -0.15 | -0.06 | -0.19 | 34 |
| Salinity | Sal | 0.06 | -0.49 | -0.23 | -0.76 | 24 |
| Specific conductivity | Cond | 0.23 | -0.61 | -0.07 | -0.72 | 34 |
| Temperature | | 0.10 | -0.32 | -0.05 | -0.55 | 34 |
| Total Suspended Solids | | 0.10 | -0.11 | 0.05 | -0.13 | 27 |
| Turbidity | | -0.02 | -0.19 | -0.06 | 0.02 | 34 |
| Velocity | | 0.06 | 0.02 | -0.05 | -0.08 | 35 |
| Metals | | | | | | |
| Aluminum | | 0.16 | 0.20 | 0.27 | 0.10 | 35 |
| Arsenic | As | 0.11 | -0.68 | -0.08 | -0.79 | 35 |
| Cadmium | Cd | 0.15 | -0.34 | -0.33 | -0.31 | 35 |
| Chromium | Cr | 0.03 | -0.52 | 0.02 | -0.68 | 35 |
| Copper | Cu | -0.21 | -0.58 | -0.26 | -0.58 | 35 |
| Lead | Pb | -0.24 | 0.06 | 0.20 | -0.04 | 35 |
| Manganese | Mn | 0.23 | -0.60 | -0.07 | -0.46 | 35 |
| Nickel | Ni | 0.31 | -0.55 | 0.10 | -0.59 | 35 |
| Selenium | Se | 0.08 | -0.61 | -0.08 | -0.78 | 34 |
| Silver | Ag | 0.22 | -0.47 | 0.08 | -0.51 | 35 |
| Zinc | Zi | -0.11 | -0.56 | -0.33 | -0.62 | 34 |
| Nutrients | | | | | | |
| Ammonia as N | NH_3-N | -0.03 | -0.57 | -0.28 | -0.53 | 35 |
| Nitrate + Nitrite as N | | 0.01 | -0.14 | -0.28 | -0.31 | 35 |
| Nitrate as N | | 0.01 | 0.03 | -0.18 | -0.28 | 27 |
| Nitrate as NO3 | NO_3 | 0.50 | 0.09 | -0.29 | -0.20 | 13 |
| Nitrite as N | | -0.07 | -0.08 | -0.29 | -0.24 | 27 |
| Nitrogen, Total Kjeldahl | | -0.15 | -0.38 | | -0.38 | |
| OrthoPhosphate as P | | 0.20 | -0.36 | | -0.50 | |
| Phosphorus as P, Total | TP | 0.08 | -0.46 | -0.16 | -0.40 | 35 |

Table 8, continued.

| Jillinded. | Spearman rank correlations (ρ) | | | | | |
|--------------------------------------|--------------------------------|-------|-------|-------|----|--|
| B. Water qualityOrganic constituents | - | NMS2 | | IBI | n | |
| PAHs | | | | | | |
| Acenaphthene | -0.15 | -0.33 | 0.05 | -0.33 | 35 | |
| Benz(a)anthracene | -0.15 | -0.33 | 0.05 | -0.33 | 35 | |
| Benzo(b)fluoranthene | -0.07 | -0.39 | -0.08 | -0.22 | 35 | |
| Chrysene | -0.16 | -0.34 | 0.01 | -0.30 | 35 | |
| Fluorenes, C2 - | 0.36 | -0.45 | -0.14 | -0.41 | 33 | |
| Naphthalenes, C3 - | -0.15 | -0.41 | -0.10 | -0.15 | 33 | |
| Naphthalenes, C4 - | -0.15 | -0.34 | -0.22 | -0.12 | 33 | |
| Phenanthrene | -0.06 | -0.39 | 0.00 | -0.32 | 35 | |
| PCBs | | | | | | |
| PCBs | 0.05 | -0.18 | -0.20 | 0.02 | 35 | |
| Pesticides | | | | | | |
| DDE(p,p') | 0.33 | | -0.25 | -0.23 | 35 | |
| DDTs | 0.35 | -0.23 | -0.25 | -0.19 | 35 | |
| Demeton-s | 0.38 | | -0.09 | -0.26 | 34 | |
| Diazinon | 0.19 | -0.57 | -0.12 | -0.57 | 34 | |
| Dimethoate | 0.15 | -0.34 | | -0.21 | | |
| Dioxathion | -0.31 | -0.26 | 0.17 | -0.33 | 34 | |
| Disulfoton | 0.16 | -0.49 | -0.02 | -0.52 | 34 | |
| Endosulfan sulfate | -0.01 | | | 0.05 | 35 | |
| Endrin Aldehyde | 0.35 | -0.14 | 0.05 | -0.33 | 35 | |
| HCH, alpha | 0.06 | -0.42 | | -0.14 | | |
| HCH, delta | 0.24 | | | -0.20 | | |
| Oxadiazon | 0.08 | | -0.23 | -0.61 | 34 | |
| Oxychlordane | 0.32 | -0.13 | | -0.13 | | |
| Parathion, Methyl | 0.09 | 0.00 | -0.32 | -0.06 | 34 | |
| Prometon | 0.43 | -0.15 | 0.33 | -0.46 | 28 | |
| Propazine | 0.01 | -0.34 | | -0.37 | 28 | |
| Secbumeton | 0.51 | -0.35 | | -0.43 | | |
| Terbuthylazine | 0.32 | -0.29 | | -0.39 | | |
| Thiobencarb | -0.31 | -0.15 | -0.40 | 0.08 | 34 | |

Table 9. Correlations of toxicity endpoints with NMS axes and IBI. N = number of sites used to calculate correlations.

| | | Spearman rank correlations (ρ) | | | | |
|---|--------|--------------------------------|-------|-------|-------|----|
| Toxicity indicator | Symbol | NMS1 | NMS2 | NMS3 | IBI | n |
| C. dubia survival (% control) | | 0.00 | 0.25 | -0.11 | 0.16 | 32 |
| C. dubia young/female (% control) | | 0.01 | 0.27 | -0.20 | 0.29 | 32 |
| H. azteca survival (% control) | | -0.03 | 0.45 | 0.23 | 0.47 | 33 |
| H. azteca growth (% control) | | 0.28 | 0.02 | -0.07 | 0.13 | 32 |
| S. capricornutum total cell count (% control) | | 0.00 | 0.61 | 0.08 | 0.35 | 34 |
| C. dubia survival frequency | CerSur | 0.15 | -0.39 | 0.09 | -0.36 | 32 |
| C. dubia young/female frequency | CerYou | -0.06 | -0.46 | 0.00 | -0.40 | 32 |
| H. azteca survival frequency | HyaSur | -0.03 | -0.40 | -0.22 | -0.42 | 33 |
| H. azteca growth frequency | HyaGro | -0.04 | -0.35 | -0.02 | -0.36 | 32 |
| S. capricornutum total cell count frequency | SelTcc | -0.12 | -0.47 | -0.03 | -0.27 | 34 |

Physical habitat variables were correlated with both axis 1 and 2, although none with $\rho^2 > 0.2$. The strongest physical habitat variables were riparian vegetation, riffle frequency, channel alteration, and epifaunal cover with ρ^2 with axis 2 of 0.17-0.18; these variables all had significant relationships with the IBI as well. Correlations with axis 1 were weaker, with ρ^2 ranging from 0.10-0.15. Therefore, physical habitat degradation appears to be associated with both water chemistry contamination and nutrient enrichment (Table 10, Figure 12).

Table 10. Correlations of physical habitat assessments with NMS axes and IBI. N = number of sites used to calculate correlations.

| | | Spearman rank correlations (p) | | | | |
|----------------------------|----------|--------------------------------|------|-------|-------|----|
| Physical habitat component | Symbol | NMS1 | NMS2 | NMS3 | IBI | n |
| Epifaunal cover | EpiCover | -0.26 | 0.39 | 0.06 | 0.45 | 30 |
| Embeddedness | Embed | -0.31 | 0.29 | 0.07 | 0.53 | 28 |
| Velocity-depth regime | VelDep | 0.00 | 0.09 | -0.12 | 0.04 | 28 |
| Sediment deposition | SedDep | 0.07 | 0.08 | 0.12 | -0.13 | 26 |
| Channel flow | ChanFlo | 0.02 | 0.04 | -0.20 | -0.08 | 27 |
| Channel alteration | ChanAlt | -0.39 | 0.41 | 0.03 | 0.64 | 29 |
| Riffle frequency | RifFreq | -0.34 | 0.41 | 0.01 | 0.64 | 28 |
| Bank stability | BankStab | 0.01 | 0.22 | 0.16 | -0.02 | 30 |
| Vegetative protection | VegPro | -0.04 | 0.28 | 0.09 | 0.12 | 30 |
| Riparian zone | RipZone | -0.28 | 0.42 | 0.15 | 0.59 | 29 |
| Mean score | AvePHAB | -0.32 | 0.41 | 0.02 | 0.50 | 28 |

Analysis of landscape-scale variables suggest that the extent of development in the watershed strongly influences benthic community structure. For example, total developed area in the watershed, as well as percent of developed land in the watershed, both had strong negative relationships with axis 2 (ρ^2 of 0.26 and 0.21, respectively). Landscape metrics reflecting local land use were more weakly correlated with axis 2, with local developed land having a ρ^2 of 0.13. Axis 2 was most strongly correlated with distance from coast (ρ^2 = 0.53), perhaps reflecting the higher intensity of development along the coast. No landscape-scale variable correlated strongly with axis 1 or 3 (Table 11, Figure 12).

Table 11. Correlations of landscape metrics with NMS axes and IBI. N = number of sites used to calculate correlations.

| outoulate conficiations. | | | | | | | |
|--|-------------|---------------------|-------|-------|-------|-------|----|
| Variable | Symbol | Unit | NMS1 | NMS2 | NMS3 | IBI | n |
| Land use, watershed-wide | | | | | | | |
| Open space in watershed | Open | log km ² | 0.00 | 0.02 | 0.05 | 0.03 | 44 |
| Agricultural land in watershed | | log km ² | 0.00 | 0.03 | -0.02 | 0.02 | 44 |
| Developed land in watershed | Developed | log km ² | 0.13 | -0.51 | -0.07 | -0.70 | 44 |
| Percent open space in watershed | | % | -0.07 | 0.35 | 0.17 | 0.42 | 44 |
| Percent agricultural land in watershed | | % | 0.03 | 0.16 | -0.01 | 0.13 | 44 |
| Percent developed land in watershed | % Developed | % | 0.13 | -0.46 | -0.12 | -0.58 | 44 |
| Land use, local | | | | | | | |
| Percent open space within 500 m | | % | 0.06 | 0.23 | -0.02 | 0.41 | 44 |
| Percent agricultural land within 500 m | Ag500 | % | -0.14 | 0.32 | 0.16 | 0.23 | 44 |
| Percent developed land within 500 m | | % | 0.02 | -0.37 | -0.10 | -0.55 | 44 |
| Other landscape-scale variables | | | | | | | |
| Distance from coast | Coast | km | -0.30 | 0.72 | 0.00 | 0.72 | 44 |
| Watershed area | WSA | log km ² | 0.00 | -0.12 | -0.01 | -0.10 | 44 |

5. DISCUSSION

Assessment of the region

Impacts to streams in southern California in this study were pervasive, and were associated with a large suite of potential stressors, including multiple water chemistry constituents, toxic waters and sediments, and degraded physical habitat. Impacts to all indicators were observed, and in most cases impacts were widespread. In general, smaller coastal watersheds (e.g., Carlsbad, Los Peñasquitos, Pueblo San Diego, and Otay) were more impacted, suffering from elevated water contaminants, high toxicity, degraded physical habitat, and poor biological condition. However, all watersheds contained sites suffering impacts to multiple indicators.

Despite the prevalence of observed impacts, some watersheds contained sites in good health. Larger watersheds with extensive undeveloped areas (e.g., Santa Margarita, San Diego, and Tijuana) contained sites in moderate to good health, generally clustered in the interior. Bioassessment samples from these sites were frequently in fair or good condition, and contamination of the water column was less severe. The San Juan hydrologic unit was unique in that it contained sites in moderate to good health near the coast. Unlike all other watersheds, the San Juan hydrologic unit contains extensive undeveloped coastal areas, where these sites were located.

Using data from local programs for regional assessments was a qualified success. These programs generated considerable quantities of data within the San Diego region, measuring multiple indicators in all watersheds. However, extrapolating results from the sites in this study to the entire San Diego region should be done cautiously. All sites were targeted for sampling, often because impairment was suspected at many of these sites. Therefore, the assumption that the sites in the study represent the region as a whole is most likely violated, resulting in a regional assessment that may be worse than the true condition.

Much of the bioassessment sampling was driven by municipal stormwater permits (such as NPDES programs) or other impact assessments (including a large number of samples collected by California Department of Fish and Game), resulting in a high density of sites in densely populated and highly developed areas along the coast. Because of this focus on urban streams, vast areas of the interior, such as the upper San Luis Rey River and the Santa Margarita, contained no sites (Figure 2).

Relationship between stressors, ecological health, and land use

Despite the limits of the data in making regional assessments, they helped determine the relationships between ecological health and potential stressors to aquatic life. The targeted selection of sites by local programs was adequate to establish gradients of most stressors, such as nutrients or metals in the water column. Poor biological condition was associated with elevated metals, organic constituents, nutrients, specific conductivity, pH, and many other water chemistry constituents. Water and sediment toxicity was more frequent at biologically impacted sites than at sites in good biological condition. Degraded physical habitat was also associated with poor biological condition.

Nonmetric multidimensional scaling of benthic communities suggests ecosystem health was degraded in two different ways: toxic contamination of the water column and nutrient enrichment. These two gradients of degradation correspond to two of the three ordination axes (axes 2 and 1, respectively). Sites located on the low end of axis 2 had elevated concentrations of many contaminants, such as metals and organic constituents, as well as elevated specific conductivity, sulfates, and other water quality constituents. Furthermore, toxicity was frequently evident at sites on the low end of axis 2. In contrast, the nutrient enrichment gradient was strongly related to dissolved oxygen and to concentrations of nitrate; sites on the high end of axis 1 may be in a nutrientenriched state. Physical habitat scores were related to both axes, suggesting that degradation of physical habitat is associated with both nutrient enrichment and with contamination of the water column. The lack of strong relationships between the third axis and potential stressors suggest that this axis represents either responses to unmeasured stressors, or natural variability in stream communities. Such variability may arise from environmental heterogeneity, as well as biotic processes like dispersal, predation, and competition among stream biota (Power et al. 1998).

The fact that the IBI and its composite metrics were all strongly correlated to axis 2 suggests that the IBI is a good tool to detect impacts from altered water chemistry, as well as degraded physical habitat. However, the weak relationship with axis 1 suggests that it may not be sensitive to impacts caused by nutrient enrichment. A complementary index, either based on macroinvertebrate taxa that respond to this gradient (e.g., *Simulium*, Oligochaeta, Cladocera, etc.) or based on assemblages with higher sensitivity to nutrient enrichment (e.g., algae) may improve assessment of impacts related to nutrient enrichment.

The nature of the data collected for this assessment does not allow identification of stressors that were directly responsible for the observed impacts to ecological health. However, the extent of developed land had one of the strongest associations with poor health observed in this study. It is evident that increased development in watersheds—perhaps as little as 20 percent—could seriously impact stream ecosystems. Similar thresholds have been identified in many other regions of the world (e.g., Hatt et al. 2004, Walsh et al. 2007). Although agricultural land within the watershed was not shown to be associated

with impacts to aquatic life, watersheds with extensive agricultural activity were minimally represented in the data set, Furthermore, agricultural land may include areas that are minimally affected, such as lightly grazed pasture.

Land use may affect stream health at both local as well as watershed scales. Our data showed that both watershed-scale and local-scale land use measured were associated with poor biological integrity. Furthermore, local conditions, as reflected by physical habitat condition, was also associated with biological health. The role near-stream conditions and riparian buffers in biological integrity has long been recognized (Hickey and Doran 2004, Moore and Palmer 2005). However, recent research suggests that watershed condition is more important than local riparian condition (e.g., Walsh et al. 2007). This study supports the finding of other studies that protection of aquatic life may require addressing local habitat, as well as watershed-wide alterations of land use and stream hydrology (Taylor et al. 2004, Walsh et al. 2005).

Conclusions

Most sites within the San Diego region were in poor condition.

Multiple lines of evidence suggested that many sites in the San Diego region were in poor condition. For example, over half of all water samples exceeded applicable aquatic life thresholds for multiple water chemistry constituents, such as ammonia-n, selenium, specific conductivity, or sulfate. Water or sediment toxicity was evident in the majority of samples; toxicity to the alga *Selenastrum capricornutum* was the most widespread affecting 59% of all samples. Impacts to benthic macroinvertebrate communities were particularly prevalent, with 80% of over 700 bioassessment samples in poor or very poor condition. Good bioassessment condition was never observed at 87 of the 144 (60%) sites assessed.

Multiple stressors were associated with poor biological condition

Poor biological condition was associated with many potential stressors, including altered water chemistry, high toxicity and degraded physical habitat. Nonmetric multidimensional scaling showed that benthic communities responded to two different gradients of stressors: toxic contaminants in the water column (e.g., trace metals, organic constituents) and nutrient enrichment (e.g., elevated nitrate and low dissolved oxygen). The toxic contaminant gradient accounted for more of the variability observed in biological communities compared to the nutrient enrichment gradient and was closely related to frequency of toxicity, as well as the index of biotic integrity. Degradation of physical habitat was associated with both toxic contaminant and nutrient enrichment gradients.

Although this study could not causally link stressors to biological degradation at any specific sites, several stressors appear to be likely candidates, including both physical habitat, water chemistry, and toxicity. Several water chemistry constituents, such as nutrients (including nitrate, ammonia, and Phosphorus), specific conductivity, sulfate, pH, and Selenium frequently exceeded aguatic life thresholds at sites in poor biological health.

Development in the watershed was a strong predictor of biological health.

Sites with extensive development in the contributing watershed were invariably in poor condition. The data suggest that development in as little as 20% of the watershed was enough to degrade biological integrity, although other stressors may affect aquatic life in undeveloped watersheds. Development in the watershed correlated strongly with the toxic contaminant gradient.

Although biological degredation was not restricted to the urbanized coast, sites draining highly developed watersheds were invariably in poor biological health. Development in the watershed was associated with all the major stressors listed above, and urbanization has been shown to be a source of water quality and physical habitat impacts in multiple studies (e.g., Walsh et al. 2007).

Recommendations

SWAMP should integrate its monitoring with other monitoring programs in the region to increase cost-efficiency

SWAMP is an important foundation for monitoring in the San Diego Region. SWAMP provides a high-quality data set that is not constrained to specific waterbodies or pollutant categories. It cannot, however, monitor all waterbodies for all of the important attributes RWQCB staff need for decision-making. This study found that hundreds of samples were collected for NPDES monitoring in the San Diego Region, sometimes in the same locations at the same time as SWAMP. SWAMP should look to integrate its monitoring with NPDES monitoring to extend its resources. This was a similar recommendation to what the SPARC had provided to the SWRCB during its most recent external review of the SWAMP program.

SWAMP should redesign its monitoring program to improve effectiveness at addressing important monitoring questions.

One of the primary questions to be addressed by the SWAMP program was "what is the health of streams in the San Diego Region?". Answering this question was hindered due to a potentially biased monitoring design that

targeted sites for sampling. Often, these were sites with known sources of pollutants. A probabilistic monitoring design would provide a more accurate assessment of stream health overall while requiring fewer sample sites. The probabilistic design has been used by others including the US EPA's Perennial Stream Assessment (PSA) and the Southern California Stormwater Monitoring Coalition's Regional Watershed Monitoring Program (SMC).

A regional monitoring program such as that put together by the SMC may not be adequate to address all site-specific concerns within the San Diego Region. For example, identification of specific causes of impairment (rather than associations) will require toxicity identification evaluations (TIEs), or an analysis using the US EPA's Causal Analysis/Diagnosis Decision Information System (CADDIS, www.epa.gov/caddis). CADDIS and TIEs are able to infer causality of impairment and produce multiple lines of evidence for such purposes as 303d listing and TMDL preparation.

Identify a set of core indicators that can help determine impacts to beneficial uses

Two challenges affected the ability to assess regional stream health and examine stressor-response relationships. First, few bioassessment samples were collected synoptically with water chemistry, toxicity, or physical habitat assessments. The lack of synoptic data obscures potentially strong stressor relationships, as stresses may wax or wane between sampling events. Second, potentially important indicators were not measured. Impacts such as nutrient enrichment may be detected most effectively using other indicators like periphyton (attached algae). Once again, there is opportunity to integrate and collaborate with local scale monitoring programs such as NPDES and larger scale programs such as PSA and SMC to define a list of core indicators. These indicators may be used for the implementation of nutrient numeric endpoints and other assessment tools in development. Collaboration with a program such as SMC presents the best opportunity for evaluating emerging assessment tools (such as nutrient numeric endpoints), as they can be evaluated with a wide range of habitat types and stressors and at several spatial scales.

SWAMP should ensure that there is an infrastructure to support its collaborative programs

One problem encountered during this study was the inability to combine data sets from different programs. Differing data structures, QA requirements, plus field and laboratory methods hindered effective progress towards meaningful interpretations of the data. SWAMP should engage in shared information management systems, integrated quality assurance system checks, and common field and laboratory method manuals to ensure that integration among monitoring programs becomes seamless. These activities have already

made a start including California's Environmental Data Exchange Network (CEDEN) and the inter-calibration exercises being conducted by the SMC.

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7. APPENDIX

Water quality constituents at each watershed. SD = standard deviation. n = number of samples.

| water quality constituents at each waters | ica. OD = | Staridard devi | | Juan | Jui | | Margarita | <u> </u> |
|---|--------------------|----------------|--------|-------|-----|-------|-----------|----------|
| Physical water quality and inorganics | Symbol | Units | Mean S | | n | | SD | n |
| Alkalinity as CaCO3 | Alk | mg/l | 208 | 109 | | 187 | | 21 |
| Chloride | | mg/l | | | 0 | | | 0 |
| Fine-ASTM | | % | | | 0 | | | 0 |
| Fine-ASTM, Passing No. 200 Sieve | | % | 24.9 | 24.6 | 25 | 6.5 | 7.0 | 12 |
| Oxygen, Dissolved | DO | mg/L | | | 0 | | | 0 |
| Oxygen, Saturation | | % | 105 | 33 | 38 | 99 | 25 | 20 |
| pH | | pН | 7.6 | 1.5 | 38 | 7.4 | 0.8 | 20 |
| Salinity | Sal | ppt | 1.05 | 1.77 | 32 | 2.62 | 6.16 | 20 |
| Specific conductivity | Cond | μS/cm | 2032 | 2839 | 38 | 4399 | 9779 | 20 |
| Sulfate | SO_4 | mg/l | 497 | 491 | 39 | 352 | 398 | 21 |
| Suspended Sediment Concentration | | % | 70.5 | 117.2 | 8 | | | 0 |
| Temperature | | °C | 16.8 | 3.2 | 38 | 16.1 | 4.1 | 20 |
| Total Organic Carbon | | mg/L | | | 0 | | | 0 |
| Total Suspended Solids | | mg/L | 101 | 241 | 31 | 25 | 48 | 21 |
| Turbidity | | NTU | 35.5 | 76.8 | 38 | 33.1 | 101.0 | 20 |
| Velocity | | ft/s | 0.4 | 0.9 | 39 | 0.5 | 0.9 | 21 |
| Nutrients | | | | | | | | |
| Ammonia as N | NH ₃ -N | mg/l | 0.19 | 0.52 | 39 | 0.02 | 0.05 | 21 |
| Nitrate + Nitrite as N | | mg/l | 0.42 | 0.52 | 39 | 4.86 | 4.52 | 21 |
| Nitrate as N | | mg/l | 0.41 | 0.51 | 39 | 4.84 | 4.52 | 21 |
| Nitrate as NO3 | NO_3 | mg/l | 1.79 | 2.20 | | 22.48 | 19.92 | |
| Nitrite as N | | mg/l | 0.02 | 0.02 | 39 | 0.02 | 0.01 | 21 |
| Nitrogen, Total Kjeldahl | | mg/l | 0.80 | 0.90 | | 0.67 | 0.72 | |
| OrthoPhosphate as P | | mg/l | 3.45 | 20.80 | | 2.30 | 9.79 | |
| Phosphorus as P,Total | TP | mg/l | 0.22 | 0.27 | 39 | 0.21 | 0.22 | 21 |
| Metals | | • | | | | | | |
| Aluminum | | μg/L | 3.3 | 8.6 | 39 | 4.5 | 13.5 | 21 |
| Arsenic | As | μg/L | 3.4 | 2.5 | 39 | 2.5 | 3.9 | 21 |
| Cadmium | Cd | μg/L | 0.26 | 0.34 | 39 | 0.04 | 0.03 | 21 |
| Chromium | Cr | μg/L | 0.25 | 0.22 | 39 | 0.23 | 0.35 | |
| Copper | Cu | μg/L | 4.05 | 2.85 | 39 | 3.10 | 2.37 | |
| Lead | Pb | μg/L | 0.02 | 0.02 | | 0.01 | 0.01 | |
| Manganese | Mn | μg/L | 148 | 329 | 39 | 92 | 139 | |
| Nickel | Ni | μg/L | 5.55 | 6.76 | 39 | 0.71 | 1.22 | 21 |
| Selenium | Se | μg/L | 7.5 | 10.4 | | 5.9 | 16.8 | |
| Silver | Ag | μg/L | 0.20 | 1.22 | 39 | 0.09 | 0.29 | |
| Zinc | Zi | μg/L | 4.1 | 3.1 | 38 | 2.3 | 1.5 | 20 |
| Bacteria | | | | | | | | |
| Enterococcus | | MPN/100 ml | | | 1 | 10 | | 1 |
| Fecal Coliform | | MPN/100 ml | | | 1 | 50 | | 1 |
| Total Coliform | | MPN/100 ml | 190 | | 1 | 900 | | <u>1</u> |

| Appendix, continued. | | San | San Juan | | Santa Margarita | | |
|-------------------------------|--------------|--------|----------|-----|-----------------|----------|--|
| PAHs | Symbol Units | Mean S | | n N | | D n | |
| Acenaphthene | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Acenaphthylene | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Anthracene | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Benz(a)anthracene | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Benzo(a)pyrene | ng/L | 1.7 | 10.8 | 39 | 0 | 0 21 | |
| Benzo(b)fluoranthene | ng/L | 3.5 | 9.6 | 39 | 1.0 | 3.2 21 | |
| Benzo(e)pyrene | ng/L | 1.3 | 6.3 | 39 | 0 | 0 21 | |
| Benzo(g,h,i)perylene | ng/L | 4.3 | 14.4 | 39 | 0 | 0 21 | |
| Benzo(k)fluoranthene | ng/L | 1.4 | 8.9 | 38 | 0 | 0 20 | |
| Biphenyl | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Chrysene | ng/L | 0.8 | 3.4 | 39 | 0 | 0 21 | |
| Chrysenes, C1 - | ng/L | 1.9 | 7.4 | 38 | 0 | 0 20 | |
| Chrysenes, C2 - | ng/L | 2.7 | 9.3 | 38 | 0 | 0 20 | |
| Chrysenes, C3 - | ng/L | 19.6 | 111.4 | 38 | 0 | 0 20 | |
| Dibenz(a,h)anthracene | ng/L | 2.4 | 15.1 | 39 | 0 | 0 21 | |
| Dibenzothiophene | ng/L | 1.7 | 6.4 | 38 | 0 | 0 20 | |
| Dibenzothiophenes, C1 - | ng/L | 10.4 | 21.1 | 38 | 5.4 | 8.4 20 | |
| Dibenzothiophenes, C2 - | ng/L | 18.1 | 38.9 | 38 | 10.5 | 12.9 20 | |
| Dibenzothiophenes, C3 - | ng/L | 9.4 | 25.1 | 38 | 2.3 | 7.1 20 | |
| Dimethylnaphthalene, 2,6- | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Dimethylphenanthrene, 3,6- | ng/L | | | 0 | | 0 | |
| Fluoranthene | ng/L | 0.82 | 3.74 | 39 | 0 | 0 21 | |
| Fluoranthene/Pyrenes, C1 - | ng/L | 0.39 | 2.43 | 38 | 0 | 0 20 | |
| Fluorene | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Fluorenes, C1 - | ng/L | 2.06 | 6.06 | 38 | 0.57 | 2.55 20 | |
| Fluorenes, C2 - | ng/L | 0.93 | 3.26 | 38 | 0 | 0 20 | |
| Fluorenes, C3 - | ng/L | 5.89 | 13.27 | 38 | 4.27 | 12.54 20 | |
| Indeno(1,2,3-c,d)pyrene | ng/L | 3.79 | 17.96 | 39 | 0 | 0 21 | |
| Methyldibenzothiophene, 4- | ng/L | | | 0 | | 0 | |
| Methylfluoranthene, 2- | ng/L | | | 0 | | 0 | |
| Methylfluorene, 1- | ng/L | | | 0 | | 0 | |
| Methylnaphthalene, 1- | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Methylnaphthalene, 2- | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Methylphenanthrene, 1- | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Naphthalene | ng/L | 1.34 | 6.42 | 39 | 0 | 0 21 | |
| Naphthalenes, C1 - | ng/L | 1.30 | 6.53 | 38 | 0 | 0 20 | |
| Naphthalenes, C2 - | ng/L | 1.17 | 4.07 | | 0 | 0 20 | |
| Naphthalenes, C3 - | ng/L | 3.65 | 8.57 | 38 | 1.70 | 5.27 20 | |
| Naphthalenes, C4 - | ng/L | 7.44 | 21.33 | 38 | 0 | 0 20 | |
| Perylene | ng/L | 1.83 | 8.04 | | 0 | 0 20 | |
| Phenanthrene | ng/L | 0 | 0 | | 0 | 0 21 | |
| Phenanthrene/Anthracene, C1 - | ng/L | 4.99 | 9.68 | 38 | 5.01 | 8.55 20 | |
| Phenanthrene/Anthracene, C2 - | ng/L | 2.57 | 7.07 | | 0.50 | 2.24 20 | |
| Phenanthrene/Anthracene, C3 - | ng/L | 2.81 | 8.01 | | 0 | 0 20 | |
| Phenanthrene/Anthracene, C4 - | ng/L | 0.44 | 2.74 | | 0 | 0 20 | |
| Pyrene | ng/L | 2.65 | 8.10 | | 0 | 0 21 | |
| Trimethylnaphthalene, 2,3,5- | ng/L | 0 | 0 | | 0 | 0 20 | |

SWAMP Synthesis Report on Stream Assessments in the San Diego Region

| | | San Ju | ıan | | Santa Marg | arita |
|---------|--------------|---------|------|----|------------|-------|
| PCBs | Symbol Units | Mean SD | | n | Mean SD | n |
| PCB 005 | ng/L | 0.35 | 1.71 | | 0 | 0 20 |
| PCB 008 | ng/L | 0.41 | 2.50 | | 0 | 0 20 |
| PCB 015 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 018 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 027 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 028 | ng/L | 0 | 0 | 38 | | 0 20 |
| PCB 029 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 031 | ng/L | 0.11 | 0.65 | | 0 | 0 20 |
| PCB 033 | ng/L | 0 | | 38 | | 0 20 |
| PCB 044 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 049 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 052 | ng/L | 0.32 | 1.95 | | | 0 20 |
| PCB 056 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 060 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 066 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 070 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 074 | ng/L | 0 | | 38 | | 0 20 |
| PCB 087 | ng/L | 0.82 | 2.01 | | 0 | 0 20 |
| PCB 095 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 097 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 099 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 101 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 105 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 110 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 114 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 118 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 128 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 137 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 138 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 141 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 149 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 151 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 153 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 156 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 157 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 158 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 170 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 174 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 177 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 180 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 183 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 187 | ng/L | 0.16 | 0.72 | 38 | 0 | 0 20 |
| PCB 189 | ng/L | 0 | 0 | 38 | 0 | 0 20 |
| PCB 194 | ng/L | 0.18 | 1.14 | 38 | | 0 20 |
| PCB 195 | ng/L | 0.18 | 1.14 | | | 0 20 |
| PCB 200 | ng/L | 0 | | 38 | | 0 20 |
| PCB 201 | ng/L | 0 | | 38 | | 0 20 |
| PCB 203 | ng/L | 0 | | 38 | 0 | 0 20 |
| PCB 206 | ng/L | 0 | | 38 | | 0 20 |

| | | | Juan | | Margarita | |
|---------------------|--------------|-------|-----------|------|-----------|--|
| PCBs | Symbol Units | | SD n | | SD n | |
| PCB 209 | ng/L | 0 | 0 38 | | 0 20 | |
| PCB-1016 | ng/L | 0 | | 1 0 | 1 | |
| PCB-1221 | ng/L | 0 | | 1 0 | 1 | |
| PCB-1232 | ng/L | 0 | | 1 0 | 1 | |
| PCB-1242 | ng/L | 0 | | 1 0 | 1 | |
| PCB-1248 | ng/L | 0 | | 1 0 | 1 | |
| PCB-1254 | ng/L | 0 | | 1 0 | 1 | |
| PCB-1260 | ng/L | 0 | | 1 0 | 1 | |
| PCBs | ng/L | 2.45 | 6.68 39 | 9 0 | 0 21 | |
| Pesticides | | | | | | |
| Aldrin | ng/L | 0 | 0 39 | | 0 21 | |
| alpha-BHC | ng/L | 0 | | 1 0 | 1 | |
| Ametryn | ng/L | 0 | 0 38 | | 0 20 | |
| Aspon | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Atraton | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Atrazine | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Azinphos ethyl | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Azinphos methyl | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| beta-BHC | ng/L | 0 | | 1 0 | 1 | |
| Bolstar | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Carbophenothion | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Chlordane (tech) | ng/L | 0 | • | 1 0 | 1 | |
| Chlordane, cis- | ng/L | 0.29 | 1.49 38 | 3 0 | 0 20 | |
| Chlordane, trans- | ng/L | 0.03 | 0.16 38 | 3 0 | 0 20 | |
| Chlordene, alpha- | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Chlordene, gamma- | ng/L | 0.34 | 1.16 38 | 3 0 | 0 20 | |
| Chlorfenvinphos | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Chlorpyrifos | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Chlorpyrifos methyl | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Ciodrin | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Coumaphos | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| Dacthal | ng/L | 0.24 | 0.71 38 | 3 0 | 0 20 | |
| DDD(o,p') | ng/L | 0 | 0 38 | 3 0 | 0 20 | |
| DDD(p,p') | ng/L | 0.03 | 0.16 39 | 0.05 | 0.22 21 | |
| DDE(o,p') | ng/L | 0.03 | 0.16 38 | | 0 20 | |
| DDE(p,p') | ng/L | 0.29 | 0.66 39 | | 2.64 21 | |
| DDMU(p,p') | ng/L | 0 | 0 38 | | 0 20 | |
| DDT(o,p') | ng/L | 0 | 0 38 | | 0.45 20 | |
| DDT(p,p') | ng/L | 0.23 | 0.83 39 | | 1.57 21 | |
| DDTs | ng/L | 0.58 | 1.28 39 | | 3.89 21 | |
| delta-BHC | ng/L | 0 | | 1 0 | 1 | |
| Demeton-s | ng/L | 0 | 0 38 | | 0 20 | |
| Diazinon | ng/L | 43.97 | 103.77 38 | | 18.10 20 | |
| Dichlofenthion | ng/L | 0 | 0 38 | | 0 20 | |
| Dichlorvos | ng/L | 0 | 0 38 | | 0 20 | |
| Dicrotophos | ng/L | 0 | 0 38 | | 0 20 | |
| Dieldrin | ng/L | 0.15 | 0.42 39 | | 0 20 | |
| Dimethoate | ng/L | 1.05 | 6.49 38 | | 0 20 | |
| Dioxathion | ng/L | 0 | 0.49 30 | | 0 20 | |

| Appendix, continued. | | San | San Juan | | Santa Margarita | | |
|----------------------|--------------|--------|----------|----|-----------------|-----------|--|
| Pesticides | Symbol Units | | | n | | SD n | |
| Disulfoton | ng/L | 3.95 | 10.28 | 38 | 0 | 0 20 | |
| Endosulfan I | ng/L | 0.14 | 0.38 | 39 | 0.05 | 0.22 21 | |
| Endosulfan II | ng/L | 0.04 | 0.24 | 39 | 0 | 0 21 | |
| Endosulfan sulfate | ng/L | 0.10 | 0.31 | 39 | 0.05 | 0.22 21 | |
| Endrin | ng/L | 0.05 | 0.22 | 39 | 0 | 0 21 | |
| Endrin Aldehyde | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Endrin Ketone | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Ethion | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Ethoprop | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Famphur | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Fenchlorphos | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Fenitrothion | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Fensulfothion | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Fenthion | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Fonofos | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| gamma-BHC (Lindane) | ng/L | 0 | | 1 | 0 | 1 | |
| HCH, alpha | ng/L | 0.04 | 0.24 | 38 | 0 | 0 20 | |
| HCH, beta | ng/L | 0 | | 38 | 0.10 | 0.45 20 | |
| HCH, delta | ng/L | 0.13 | 0.34 | | 0 | 0 20 | |
| HCH, gamma | ng/L | 0 | | 38 | 0 | 0 20 | |
| Heptachlor | ng/L | 0 | 0 | 39 | 0 | 0 21 | |
| Heptachlor epoxide | ng/L | 0.21 | 0.70 | | 0.10 | 0.30 21 | |
| Hexachlorobenzene | ng/L | 0.18 | 0.68 | | 0 | 0 20 | |
| Leptophos | ng/L | 0 | | 38 | 0 | 0 20 | |
| Malathion | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Merphos | ng/L | 0 | | 38 | 0 | 0 20 | |
| Methidathion | ng/L | 0 | | 38 | 0 | 0 20 | |
| Methoxychlor | ng/L | 0 | | 39 | 0 | 0 21 | |
| Mevinphos | ng/L | 0 | | 38 | 0 | 0 20 | |
| Mirex | ng/L | 0 | | 38 | 0 | 0 20 | |
| Molinate | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Naled | ng/L | 0 | | 38 | 0 | 0 20 | |
| Nonachlor, cis- | ng/L | 0.03 | 0.16 | | 0.05 | 0.22 20 | |
| Nonachlor, trans- | ng/L | 0.05 | 0.23 | | 0.05 | 0.22 20 | |
| Oxadiazon | ng/L | 46.21 | 164.92 | | 22.15 | 68.41 20 | |
| Oxychlordane | ng/L | 0.11 | 0.31 | | 0.05 | 0.22 20 | |
| Parathion, Ethyl | ng/L | 0 | 0.01 | 38 | 0.00 | 0 20 | |
| Parathion, Methyl | ng/L | 0 | | 38 | 0 | 0 20 | |
| Phorate | ng/L | 0 | | 38 | 0 | 0 20 | |
| Phosmet | ng/L | 0 | | 38 | 0 | 0 20 | |
| Phosphamidon | ng/L | 0 | | 38 | 0 | 0 20 | |
| Prometon | ng/L | 0 | | 38 | 0 | 0 20 | |
| Prometryn | ng/L | 0 | 0 | 38 | 0 | 0 20 | |
| Propazine | ng/L | 0 | | 38 | 0 | 0 20 | |
| Secbumeton | ng/L | 3.11 | 13.48 | | 1.75 | 7.83 20 | |
| Simazine | ng/L | 0.89 | 5.52 | | 35.90 | 100.34 20 | |
| Simetryn | _ | | | 38 | | 0 20 | |
| Sulfotep | ng/L | 0 0 | | 38 | 0 | 0 20 | |
| | ng/L | | | | | | |
| Tedion | ng/L | 0.05 | 0.23 | JÖ | 0 | 0 20 | |

| | | San | Juan | Santa Margarita | | |
|-------------------|--------------|--------|----------|-----------------|------|--|
| Pesticides | Symbol Units | Mean S | D n | Mean SD | n | |
| Terbufos | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Terbuthylazine | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Terbutryn | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Tetrachlorvinphos | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Thiobencarb | ng/L | 3.95 | 24.33 38 | 0 | 0 20 | |
| Thionazin | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Tokuthion | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Toxaphene | ng/L | 0 | 1 | 0 | 1 | |
| Trichlorfon | ng/L | 0 | 0 38 | 0 | 0 20 | |
| Trichloronate | ng/L | 0 | 0 38 | 0 | 0 20 | |

| дрених, сониниси. | | | San Luis | Rev | | Car | Isbad | |
|---------------------------------------|--------------------|------------|----------|------|----|-------|-------|----|
| Physical water quality and inorganics | Symbol | Units | Mean SD | | n | | SD | n |
| Alkalinity as CaCO3 | Alk | mg/l | 198 | | 25 | 252 | 43 | 41 |
| Chloride | | mg/l | 95 | 129 | 3 | | | 0 |
| Fine-ASTM | | % | | | 0 | 24.4 | 27.1 | 30 |
| Fine-ASTM, Passing No. 200 Sieve | | % | | | 0 | 24.9 | 20.0 | 10 |
| Oxygen, Dissolved | DO | mg/L | 9.4 | 3.6 | 24 | | | 0 |
| Oxygen, Saturation | | % | 98 | 39 | 24 | 102 | 32 | 40 |
| рН | | pН | 7.7 | 0.4 | | 7.9 | 1.5 | 40 |
| Salinity | Sal | ppt | 0.83 | 0.52 | | | | 0 |
| Specific conductivity | Cond | μS/cm | 1516 | 968 | 24 | 3800 | 4232 | 40 |
| Sulfate | SO_4 | mg/l | 382 | 264 | 25 | 469 | 277 | 41 |
| Suspended Sediment Concentration | | % | | | 0 | | | 0 |
| Temperature | | °C | 17.1 | 4.5 | 24 | 18.5 | 3.0 | 40 |
| Total Organic Carbon | | mg/L | | | 0 | 2.24 | | 1 |
| Total Suspended Solids | | mg/L | 45 | 79 | 25 | 15 | | 1 |
| Turbidity | | NTU | 12.0 | 19.0 | 24 | 3.5 | 3.9 | 40 |
| Velocity | | ft/s | 0.5 | 0.9 | 25 | 0.7 | 1.4 | 41 |
| Nutrients | | | | | | | | |
| Ammonia as N | NH ₃ -N | mg/l | 0.05 | 0.06 | 25 | 0.12 | 0.11 | 41 |
| Nitrate + Nitrite as N | | mg/l | 5.13 | 5.50 | 25 | 6.35 | 11.47 | 41 |
| Nitrate as N | | mg/l | 5.10 | 5.49 | 25 | 3.70 | | 1 |
| Nitrate as NO3 | NO_3 | mg/l | | | 0 | 16.00 | | 1 |
| Nitrite as N | | mg/l | 0.03 | 0.04 | 25 | 0 | | 1 |
| Nitrogen, Total Kjeldahl | | mg/l | 0.89 | 1.18 | | 0.49 | 0.20 | 41 |
| OrthoPhosphate as P | | mg/l | 31.00 | | 1 | 0.12 | 0.07 | |
| Phosphorus as P,Total | TP | mg/l | 0.21 | 0.24 | 25 | 0.14 | 0.09 | |
| Metals | | Ü | | | | | | |
| Aluminum | | μg/L | 6.0 | 6.4 | 25 | 14.7 | 71.5 | 41 |
| Arsenic | As | μg/L | 1.3 | 0.9 | 25 | 4.7 | 2.6 | 41 |
| Cadmium | Cd | μg/L | 0.03 | 0.02 | 25 | 0.05 | 0.04 | 41 |
| Chromium | Cr | μg/L | 0.33 | 0.28 | 25 | 1.06 | 1.05 | 41 |
| Copper | Cu | μg/L | 4.03 | 2.60 | 25 | 3.55 | 1.50 | 41 |
| Lead | Pb | μg/L | 0.06 | 0.05 | 25 | 0.05 | 0.12 | 41 |
| Manganese | Mn | μg/L | 133 | 270 | 25 | 127 | 147 | 41 |
| Nickel | Ni | μg/L | 0.97 | 2.14 | 25 | 2.16 | 1.37 | 41 |
| Selenium | Se | μg/L | 4.9 | 4.8 | 24 | 10.6 | 10.1 | 40 |
| Silver | Ag | μg/L | 0.07 | 0.34 | 25 | 0.05 | 0.28 | 41 |
| Zinc | Zi | μg/L | 2.7 | 1.9 | 24 | 6.5 | 6.4 | 40 |
| Bacteria | | | | | | | | |
| Enterococcus | | MPN/100 ml | | | 1 | 490 | | 1 |
| Fecal Coliform | | MPN/100 ml | | | 1 | 900 | | 1 |
| Total Coliform | | MPN/100 ml | 1600 | | 1 | 1600 | | 1 |

| Appendix, continued. | | San Luis | Rev | Carl | sbad |
|-------------------------------|--------------|----------|--------------------|--------|----------|
| PAHs | Symbol Units | Mean SD | | | SD n |
| Acenaphthene | ng/L | 0 | 0 25 | 1 | 5 41 |
| Acenaphthylene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Anthracene | ng/L | 0 | 0 25 | 4 | 14 41 |
| Benz(a)anthracene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Benzo(a)pyrene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Benzo(b)fluoranthene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Benzo(e)pyrene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Benzo(g,h,i)perylene | ng/L | 0.3 | 1.3 25 | 0 | 0 41 |
| Benzo(k)fluoranthene | ng/L | 0 | 0 24 | 0 | 0 40 |
| Biphenyl | ng/L | 0 | 0 24 | 0 | 0 40 |
| Chrysene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Chrysenes, C1 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Chrysenes, C2 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Chrysenes, C3 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Dibenz(a,h)anthracene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Dibenzothiophene | ng/L | 0.3 | 1.7 24 | 0 | 0 10 |
| Dibenzothiophenes, C1 - | ng/L | 0 | 0 24 | Ö | 0 10 |
| Dibenzothiophenes, C2 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Dibenzothiophenes, C3 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Dimethylnaphthalene, 2,6- | ng/L | 0 | 1 24 | 0 | 0 40 |
| Dimethylphenanthrene, 3,6- | ng/L | 0 | 0 18 | · | 0 |
| Fluoranthene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Fluoranthene/Pyrenes, C1 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Fluorene | ng/L | 0 | 0 25 | 1 | 4 41 |
| Fluorenes, C1 - | ng/L | Ö | 0 24 | 0 | 0 10 |
| Fluorenes, C2 - | ng/L | 0 | 0 24 | 28.82 | 5.41 10 |
| Fluorenes, C3 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Indeno(1,2,3-c,d)pyrene | ng/L | 0 | 0 25 | 0 | 0 10 |
| Methyldibenzothiophene, 4- | ng/L | 0 | 0 18 | U | 0 41 |
| Methylfluoranthene, 2- | ng/L | 0 | 0 18 | | 0 |
| Methylfluorene, 1- | ng/L | 0 | 0 18 | | 0 |
| Methylnaphthalene, 1- | _ | 0 | 0 18 | 0 | 0 40 |
| Methylnaphthalene, 2- | ng/L ng/L | 0 | 0 24 | 0 0 | 0 40 |
| | - | 0 | 0 24 | 0 | 0 40 |
| Methylphenanthrene, 1- | ng/L | 0.20 | 1.01 25 | 4.27 | 11.60 41 |
| Naphthalene | ng/L | | | | |
| Naphthalenes, C1 - | ng/L | 0.46 | 2.25 24 3.81 24 | 0 | 0 10 |
| Naphthalenes, C2 - | ng/L | 1.34 | | 0 | 0 10 |
| Naphthalenes, C3 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Naphthalenes, C4 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Perylene | ng/L | 0 | 0 24 | 0 | 0 40 |
| Phenanthrene | ng/L | 0 | 0 25 | 4 | 14 41 |
| Phenanthrene/Anthracene, C1 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Phenanthrene/Anthracene, C2 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Phenanthrene/Anthracene, C3 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Phenanthrene/Anthracene, C4 - | ng/L | 0 | 0 24 | 0 | 0 10 |
| Pyrene | ng/L | 0 | 0 25 | 0 | 0 41 |
| Trimethylnaphthalene, 2,3,5- | ng/L | 0 | 0 24 | 0 | 0 40 |

SWAMP Synthesis Report on Stream Assessments in the San Diego Region

| Appendix, continued. | | San Luis R | Rey | Carlsb | ad |
|----------------------|--------------|------------|------|---------|------|
| PCBs | Symbol Units | Mean SD | | Mean SD | n |
| PCB 005 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 008 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 015 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 018 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 027 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 028 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 029 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 031 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 033 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 044 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 049 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 052 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 056 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 060 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 066 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 070 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 074 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 087 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 095 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 097 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 099 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 101 | ng/L | 0 | 0 24 | Ö | 0 40 |
| PCB 105 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 110 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 114 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 118 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 128 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 137 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 138 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 141 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 149 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 151 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 153 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 156 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 157 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 158 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 170 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 174 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 177 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 180 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 183 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 187 | _ | 0 | 0 24 | | 0 40 |
| PCB 189 | ng/L | 0 | 0 24 | 0 0 | 0 40 |
| | ng/L | | | | |
| PCB 194 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 195 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 200 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 201 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 203 | ng/L | 0 | 0 24 | 0 | 0 40 |
| PCB 206 | ng/L | 0 | 0 24 | 0 | 0 40 |

| | endix, continued. San Luis Rey | | | | | | |
|---------------------|--------------------------------|---------|----------|-------|-----------|--|--|
| PCBs | Symbol Units | Mean SI |) n | | SD n | | |
| PCB 209 | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| PCB-1016 | ng/L | 0 | 1 | 0 | 1 | | |
| PCB-1221 | ng/L | 0 | 1 | 0 | 1 | | |
| PCB-1232 | ng/L | 0 | 1 | 0 | 1 | | |
| PCB-1242 | ng/L | 0 | 1 | 0 | 1 | | |
| PCB-1248 | ng/L | 0 | 1 | 0 | 1 | | |
| PCB-1254 | ng/L | 0 | 1 | 0 | 1 | | |
| PCB-1260 | ng/L | 0 | 1 | 0 | 1 | | |
| PCBs | ng/L | 0 | 0 25 | 0 | 0 41 | | |
| Pesticides | | | | | | | |
| Aldrin | ng/L | 0 | 0 25 | 0.05 | 0.22 41 | | |
| alpha-BHC | ng/L | 0 | 1 | 0 | 1 | | |
| Ametryn | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| Aspon | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| Atraton | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| Atrazine | ng/L | 8.13 | 39.80 24 | 11.88 | 27.66 40 | | |
| Azinphos ethyl | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| Azinphos methyl | ng/L | 0 | 0 24 | 1.00 | 6.32 40 | | |
| beta-BHC | ng/L | 0 | 1 | 0 | 1 | | |
| Bolstar | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| Carbophenothion | ng/L | 0 | 0 24 | 4.00 | 12.15 40 | | |
| Chlordane (tech) | ng/L | 0 | 1 | 0 | 1 | | |
| Chlordane, cis- | ng/L | 0.04 | 0.20 24 | 0.03 | 0.16 40 | | |
| Chlordane, trans- | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| Chlordene, alpha- | ng/L | 0 | 0 24 | | 0 40 | | |
| Chlordene, gamma- | ng/L | 0 | 0 24 | 0.35 | 1.44 40 | | |
| Chlorfenvinphos | ng/L | 0 | 0 24 | | 0 40 | | |
| Chlorpyrifos | ng/L | 0 | 0 24 | | 0 40 | | |
| Chlorpyrifos methyl | ng/L | 0 | 0 24 | | 0 40 | | |
| Ciodrin | ng/L | 0 | 0 24 | | 0 40 | | |
| Coumaphos | ng/L | 0 | 0 24 | | 0 40 | | |
| Dacthal | ng/L | 0 | 0 24 | | 0.49 40 | | |
| DDD(o,p') | ng/L | 0 | 0 24 | 0.08 | 0.35 40 | | |
| DDD(p,p') | ng/L | 0 | 0 25 | 0.05 | 0.22 41 | | |
| DDE(o,p') | ng/L | 0 | 0 24 | | 0 40 | | |
| DDE(p,p') | ng/L | 0.04 | 0.20 25 | 1.37 | 2.24 41 | | |
| DDMU(p,p') | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| DDT(o,p') | ng/L | 0 | 0 24 | 0 | 0 40 | | |
| DDT(p,p') | ng/L | 0 | 0 25 | 0.44 | 1.07 41 | | |
| DDTs | ng/L | 0.04 | 0.20 25 | 1.93 | 2.79 41 | | |
| delta-BHC | ng/L | 0 | 1 | 0 | 1 | | |
| Demeton-s | ng/L | 0 | 0 24 | | 13.40 40 | | |
| Diazinon | ng/L | 0.67 | 2.43 24 | | 101.40 40 | | |
| Dichlofenthion | ng/L | 0.07 | 0 24 | | 0 40 | | |
| Dichlorvos | ng/L | 0 | 0 24 | | 0 40 | | |
| Dicrotophos | ng/L | 0 | 0 24 | | 6.32 40 | | |
| Dieldrin | ng/L | 0 | 0 24 | | 0.32 40 | | |
| Dimethoate | ng/L | 0 | 0 23 | | 14.46 40 | | |
| Dioxathion | ng/L | 0 | 0 24 | | 12.15 40 | | |

| Appendix, continued. | | San Lui | San Luis Rey | | Cai | Isbad |
|------------------------|--------------|---------|--------------|----------|-------|-----------|
| Pesticides | Symbol Units | Mean SE | | n | | SD n |
| Disulfoton | ng/L | 0 | 0 | 24 | 33.33 | 38.29 40 |
| Endosulfan I | ng/L | 0 | 0 | 25 | 0.05 | 0.22 41 |
| Endosulfan II | ng/L | 0 | 0 | 25 | 0.43 | 1.31 41 |
| Endosulfan sulfate | ng/L | 0 | 0 | 25 | 0.02 | 0.16 41 |
| Endrin | ng/L | 0 | 0 | 25 | 0.08 | 0.37 41 |
| Endrin Aldehyde | ng/L | 0 | 0 | 25 | 0.51 | 1.73 41 |
| Endrin Ketone | ng/L | 0 | 0 | 24 | 0 | 0 40 |
| Ethion | ng/L | 0 | | 24 | 0 | 0 40 |
| Ethoprop | ng/L | 0 | | 24 | 0 | 0 40 |
| Famphur | ng/L | 0 | 0 | 24 | 0 | 0 40 |
| Fenchlorphos | ng/L | 0 | | 24 | 0 | 0 40 |
| Fenitrothion | ng/L | 0 | | 24 | 0 | 0 40 |
| Fensulfothion | ng/L | 0 | | 24 | 0 | 0 40 |
| Fenthion | ng/L | 0 | | 24 | 0 | 0 40 |
| Fonofos | ng/L | 0 | | 24 | 0 | 0 40 |
| gamma-BHC (Lindane) | ng/L | Ő | U | 1 | 0 | 1 |
| HCH, alpha | ng/L | 0 | 0 | 24 | 0.33 | 0.88 40 |
| HCH, beta | ng/L | 0 | | | 0.46 | 2.54 40 |
| HCH, delta | ng/L | 0 | | 24 | 0.40 | 0.43 40 |
| HCH, gamma | ng/L | 0 | | | 0.13 | 0.45 40 |
| Heptachlor | ng/L | 0 | | 25 | 0.00 | 0.33 40 |
| Heptachlor epoxide | ng/L | 0 | | 25 | 0 | 0 41 |
| Hexachlorobenzene | _ | 0 | 0 | 24 | 0.12 | 0.27 40 |
| | ng/L | 0 | | 24 | 0.12 | 0.27 40 |
| Leptophos Malathion | ng/L | 0 | | 24 | 0.93 | 5.85 40 |
| | ng/L | 0 | | 24 | | |
| Merphos | ng/L | | | | 0 | 0 40 |
| Methidathion | ng/L | 0 | | 24 | 1.00 | 6.32 40 |
| Methoxychlor | ng/L | 0 | | 25 24 | 0.02 | 0.16 41 |
| Mevinphos | ng/L | 0 | 0 | | 5.00 | 13.40 40 |
| Malianta | ng/L | 0 | | 24 | 0 | 0 40 |
| Molinate | ng/L | 0 | | 24 | 0 | 0 40 |
| Naled | ng/L | 0 | | | 5.00 | 13.40 40 |
| Nonachlor, cis- | ng/L | 0 | | 24 | 0 | 0 40 |
| Nonachlor, trans- | ng/L | 0 | | 24 | 0.05 | 0.22 40 |
| Oxadiazon | ng/L | 5.08 | 8.87 | | 64.74 | 292.81 40 |
| Oxychlordane | ng/L | 0 | 0 | 24 | 0 | 0 40 |
| Parathion, Ethyl | ng/L | 0 | 0 | 24 | 0 | 0 40 |
| Parathion, Methyl | ng/L | 0 | - | 24 | 0.75 | 4.74 40 |
| Phorate | ng/L | 0 | | 24 | 0 | 0 40 |
| Phosmet | ng/L | 0 | | 24 | 0 | 0 40 |
| Phosphamidon | ng/L | 0 | | 24 | 0 | 0 40 |
| Prometon | ng/L | 0 | | 24 | 7.18 | 20.81 40 |
| Prometryn | ng/L | 0 | | 24 | 0 | 0 40 |
| Propazine | ng/L | 0 | | 24 | 7.00 | 14.18 40 |
| Secbumeton | ng/L | 4.50 | 12.50 | | 85.00 | 153.57 40 |
| Simazine | ng/L | | 60.61 | | 0 | 0 40 |
| Simetryn | ng/L | 0 | - | 24 | 0 | 0 40 |
| Sulfotep | ng/L | 0 | | 24 | 0 | 0 40 |
| Tedion | ng/L | 0 | 0 | 24 | 0 | 0 40 |

| | | San Luis Rey | | Ca | rlsbad |
|-------------------|--------------|--------------|----------|--------|-----------|
| Pesticides | Symbol Units | Mean S | SD n | Mean | SD n |
| Terbufos | ng/L | 0 | 0 24 | 0 | 0 40 |
| Terbuthylazine | ng/L | 7.29 | 15.43 24 | 242.40 | 488.92 40 |
| Terbutryn | ng/L | 0 | 0 24 | 0 | 0 40 |
| Tetrachlorvinphos | ng/L | 0 | 0 24 | 1.00 | 6.32 40 |
| Thiobencarb | ng/L | 0 | 0 24 | 0 | 0 40 |
| Thionazin | ng/L | 0 | 0 24 | 0 | 0 40 |
| Tokuthion | ng/L | 0 | 0 24 | 0 | 0 40 |
| Toxaphene | ng/L | 0 | 1 | 0 | 1 |
| Trichlorfon | ng/L | 0 | 0 24 | 0 | 0 40 |
| Trichloronate | ng/L | 0 | 0 24 | 0 | 0 40 |

| Appendix, continued. | | | San D | ieguito | | Los | Peñasqui | tos |
|---------------------------------------|--------------------|------------|-------|---------|----|------|----------|-----|
| Physical water quality and inorganics | Symbol | Units | | SD | n | | SD | n |
| Alkalinity as CaCO3 | Alk | mg/l | 243 | 129 | | 221 | | 20 |
| Chloride | | mg/l | | | 0 | | | 0 |
| Fine-ASTM | | % | | | 0 | 33.3 | 31.8 | 13 |
| Fine-ASTM, Passing No. 200 Sieve | | % | 18.1 | 7.1 | 7 | 32.2 | 53.0 | 3 |
| Oxygen, Dissolved | DO | mg/L | | | 0 | | | 0 |
| Oxygen, Saturation | | % | 86 | 19 | 17 | 107 | 34 | 19 |
| pH | | рН | 8.0 | 0.4 | 17 | 7.9 | 0.4 | 19 |
| Salinity | Sal | ppt | 2.53 | 5.47 | 17 | | | 0 |
| Specific conductivity | Cond | μS/cm | 4316 | 8781 | 17 | 2981 | 1256 | 19 |
| Sulfate | SO_4 | mg/l | 358 | 367 | 18 | 674 | 407 | 20 |
| Suspended Sediment Concentration | | % | | | 0 | | | 0 |
| Temperature | | °C | 15.1 | 6.0 | 17 | 19.4 | 4.4 | 19 |
| Total Organic Carbon | | mg/L | | | 0 | | | 0 |
| Total Suspended Solids | | mg/L | 18 | 23 | 18 | 0 | | 1 |
| Turbidity | | NTU | 5.9 | 5.3 | 17 | 14.5 | 26.2 | 19 |
| Velocity | | ft/s | 0.3 | 0.5 | 18 | 1.1 | 1.5 | 20 |
| Nutrients | | | | | | | | |
| Ammonia as N | NH ₃ -N | mg/l | 0.10 | 0.12 | 18 | 0.09 | 0.06 | 20 |
| Nitrate + Nitrite as N | | mg/l | 0.44 | 0.60 | 18 | 0.52 | 0.70 | 20 |
| Nitrate as N | | mg/l | 0.42 | 0.58 | 18 | 0 | | 1 |
| Nitrate as NO3 | NO_3 | mg/l | 1.96 | 2.58 | 17 | 0 | | 1 |
| Nitrite as N | | mg/l | 0.02 | 0.03 | 18 | 0 | | 1 |
| Nitrogen, Total Kjeldahl | | mg/l | 1.08 | 0.81 | | 0.58 | 0.55 | 20 |
| OrthoPhosphate as P | | mg/l | 0.16 | 0.15 | 18 | 0.04 | 0.04 | 20 |
| Phosphorus as P,Total | TP | mg/l | 0.24 | 0.35 | 18 | 0.10 | 0.16 | 20 |
| Metals | | | | | | | | |
| Aluminum | | μg/L | 3.9 | 5.4 | 18 | 5.7 | 7.5 | 20 |
| Arsenic | As | μg/L | 2.1 | 1.7 | 18 | 3.4 | 0.8 | 20 |
| Cadmium | Cd | μg/L | 0.03 | 0.03 | 18 | 0.02 | 0.01 | 20 |
| Chromium | Cr | μg/L | 0.18 | 0.17 | 18 | 0.89 | 0.97 | 20 |
| Copper | Cu | μg/L | 2.41 | 1.76 | | 4.03 | 1.58 | |
| Lead | Pb | μg/L | 0.03 | 0.03 | 18 | 0.05 | 0.08 | |
| Manganese | Mn | μg/L | 135 | 135 | | 141 | 156 | |
| Nickel | Ni | μg/L | 0.70 | 0.79 | 18 | 3.38 | 3.55 | 20 |
| Selenium | Se | μg/L | 3.7 | 5.4 | | 7.8 | | 19 |
| Silver | Ag | μg/L | 0.00 | 0.00 | | 0.06 | 0.25 | |
| Zinc | Zi | μg/L | 2.2 | 1.7 | 17 | 8.4 | 8.7 | 19 |
| Bacteria | | | | | | | | |
| Enterococcus | | MPN/100 ml | | | 1 | 11 | | 1 |
| Fecal Coliform | | MPN/100 ml | | | 1 | 500 | | 1 |
| Total Coliform | | MPN/100 ml | 1600 | | 1 | 1600 | | 1 |

| | | | San Dieguito | | Los Pe | ñasquitos |
|-------------------------------|--------------|-------|--------------|----|---------|-----------|
| PAHs | Symbol Units | | | | Mean SI | |
| Acenaphthene | ng/L | 0 | | 18 | 0 | 0 20 |
| Acenaphthylene | ng/L | 0 | 0 | 18 | 0 | 0 20 |
| Anthracene | ng/L | 0 | 0 | 18 | 0 | 0 20 |
| Benz(a)anthracene | ng/L | 0 | 0 | 18 | 0 | 0 20 |
| Benzo(a)pyrene | ng/L | 1.0 | 4.4 | 18 | 0 | 0 20 |
| Benzo(b)fluoranthene | ng/L | 3.0 | 5.9 | 18 | 0 | 0 20 |
| Benzo(e)pyrene | ng/L | 0.9 | 4.0 | 18 | 0 | 0 20 |
| Benzo(g,h,i)perylene | ng/L | 10.1 | 42.9 | 18 | 0 | 0 20 |
| Benzo(k)fluoranthene | ng/L | 0 | 0 | 17 | 0 | 0 19 |
| Biphenyl | ng/L | 0 | 0 | 17 | 0 | 0 19 |
| Chrysene | ng/L | 0 | 0 | 18 | 0 | 0 20 |
| Chrysenes, C1 - | ng/L | 0 | 0 | 17 | 0 | 0 4 |
| Chrysenes, C2 - | ng/L | 3.9 | 16.2 | 17 | 0 | 0 4 |
| Chrysenes, C3 - | ng/L | 0.7 | 3.0 | 17 | 0 | 0 4 |
| Dibenz(a,h)anthracene | ng/L | 0 | | 18 | 0 | 0 20 |
| Dibenzothiophene | ng/L | 0 | 0 | 17 | 0 | 0 4 |
| Dibenzothiophenes, C1 - | ng/L | 6.6 | 8.4 | | 0 | 0 4 |
| Dibenzothiophenes, C2 - | ng/L | 13.6 | 13.0 | | 0 | 0 |
| Dibenzothiophenes, C3 - | ng/L | 4.5 | 9.9 | | 0 | 0 |
| Dimethylnaphthalene, 2,6- | ng/L | 2 | | 17 | 0 | 0 1 |
| Dimethylphenanthrene, 3,6- | ng/L | _ | J | 0 | · · | 0 1 |
| Fluoranthene | ng/L | 6.99 | 24.36 | | 0 | 0 2 |
| Fluoranthene/Pyrenes, C1 - | ng/L | 0.63 | 2.60 | | 0 | 0 4 |
| Fluorene | ng/L | 0.00 | | 18 | 0 | 0 2 |
| Fluorenes, C1 - | ng/L | 1.29 | 3.65 | | 0 | 0 2 |
| Fluorenes, C2 - | ng/L | 0 | | 17 | 27.93 | 2.53 |
| Fluorenes, C3 - | ng/L | 2.93 | 5.49 | | 8.28 | 16.55 |
| | _ | 2.93 | 12.42 | | 0.20 | 0 2 |
| Indeno(1,2,3-c,d)pyrene | ng/L | 2.93 | 12.42 | 0 | U | |
| Methyldibenzothiophene, 4- | ng/L | | | | | |
| Methylfluoranthene, 2- | ng/L | | | 0 | | |
| Methylfluorene, 1- | ng/L | 0 | 0 | | ^ | |
| Methylnaphthalene, 1- | ng/L | 0 | | 17 | 0 | 0 1 |
| Methylnaphthalene, 2- | ng/L | 0 | | 17 | 0 | 0 1 |
| Methylphenanthrene, 1- | ng/L | 1 | | 17 | 0 | 0 1 |
| Naphthalene | ng/L | 0 | | 18 | 1.75 | 7.83 2 |
| Naphthalenes, C1 - | ng/L | 0 | | 17 | 0 | 0 |
| Naphthalenes, C2 - | ng/L | 2.81 | 11.57 | | 0 | 0 |
| Naphthalenes, C3 - | ng/L | 2.06 | 5.82 | | 5.45 | 10.90 |
| Naphthalenes, C4 - | ng/L | 1.25 | 3.52 | | 0 | 0 4 |
| Perylene | ng/L | 0 | | 17 | 0 | 0 1 |
| Phenanthrene | ng/L | 2 | 11 | | 0 | 0 2 |
| Phenanthrene/Anthracene, C1 - | ng/L | 7.96 | 15.35 | | 0 | 0 |
| Phenanthrene/Anthracene, C2 - | ng/L | 3.74 | 7.48 | | 0 | 0 |
| Phenanthrene/Anthracene, C3 - | ng/L | 3.28 | 10.12 | | 0 | 0 |
| Phenanthrene/Anthracene, C4 - | ng/L | 0.64 | 2.64 | 17 | 0 | 0 - |
| Pyrene | ng/L | 24.61 | 82.69 | | 0 | 0 2 |
| Trimethylnaphthalene, 2,3,5- | ng/L | 0 | 0 | 17 | 0 | 0 19 |

SWAMP Synthesis Report on Stream Assessments in the San Diego Region

| | | San Diegu | ito | Los Peñasquitos | | |
|--------------------|--------------|-----------|------|-----------------|------|--|
| PCBs | Symbol Units | Mean SD | n | Mean SD | n | |
| PCB 005 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 008 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 015 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 018 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 027 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 028 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 029 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 031 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 033 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 044 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 049 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 052 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 056 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 060 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 066 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 070 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 074 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 087 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 095 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 097 | ng/L | 0 | 0 17 | 0 | 0 19 | |
| PCB 099 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 101 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 105 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 110 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 114 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 118 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 128 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 137 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 138 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 141 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 149 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 151 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 153 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 156 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 157 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 158 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 170 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 174 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 177 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 180 | ng/L | Ō | 0 17 | | 0 19 | |
| PCB 183 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 187 | ng/L | Ö | 0 17 | | 0 19 | |
| PCB 189 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 194 | ng/L | Ö | 0 17 | | 0 19 | |
| PCB 195 | ng/L | Ö | 0 17 | | 0 19 | |
| PCB 200 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 201 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 203 | ng/L | 0 | 0 17 | | 0 19 | |
| PCB 203 PCB 206 | ng/L | 0 | 0 17 | | 0 19 | |
| 1 00 200 | ⊓g/∟ | <u> </u> | 0 17 | U | 0 19 | |

| | | San Di | | | Peñasquitos |
|---------------------|--------------|---------|---------|---------|-------------|
| PCBs | Symbol Units | Mean SI | | | SD n |
| PCB 209 | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| PCB-1016 | ng/L | 0 | | 1 0 | 1 |
| PCB-1221 | ng/L | 0 | | 1 0 | 1 |
| PCB-1232 | ng/L | 0 | | 1 0 | 1 |
| PCB-1242 | ng/L | 0 | | 1 0 | 1 |
| PCB-1248 | ng/L | 0 | | 1 0 | 1 |
| PCB-1254 | ng/L | 0 | | 1 0 | 1 |
| PCB-1260 | ng/L | 0 | | 1 0 | 1 |
| PCBs | ng/L | 0 | 0 1 | 8 0 | 0 20 |
| Pesticides | _ | _ | | | |
| Aldrin | ng/L | 0 | 0 1 | | 0.67 20 |
| alpha-BHC | ng/L | 0 | | 1 0 | 1 |
| Ametryn | ng/L | 0 | 0 1 | | 0 19 |
| Aspon | ng/L | 0 | 0 1 | | 0 19 |
| Atraton | ng/L | 0 | 0 1 | | 29.82 19 |
| Atrazine | ng/L | 0 | 0 1 | | 35.70 19 |
| Azinphos ethyl | ng/L | 0 | | 7 0 | 0 19 |
| Azinphos methyl | ng/L | 0 | 0 1 | | 12.61 19 |
| beta-BHC | ng/L | 0 | | 1 0 | 1 |
| Bolstar | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| Carbophenothion | ng/L | 0 | 0 1 | 7 9.37 | 19.00 19 |
| Chlordane (tech) | ng/L | 0 | | 1 0 | 1 |
| Chlordane, cis- | ng/L | 0.06 | 0.24 1 | 7 0 | 0 19 |
| Chlordane, trans- | ng/L | 0 | 0 1 | | 0 19 |
| Chlordene, alpha- | ng/L | 0 | | 7 0.63 | 2.75 19 |
| Chlordene, gamma- | ng/L | 0 | 0 1 | 7 0.37 | 1.16 19 |
| Chlorfenvinphos | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| Chlorpyrifos | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| Chlorpyrifos methyl | ng/L | 0 | 0 1 | | 0 19 |
| Ciodrin | ng/L | 0 | 0 1 | | 0 19 |
| Coumaphos | ng/L | 0 | 0 1 | 7 2.63 | 11.47 19 |
| Dacthal | ng/L | 0 | 0 1 | 7 0.16 | 0.37 19 |
| DDD(o,p') | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| DDD(p,p') | ng/L | 0 | 0 1 | 8 0.05 | 0.22 20 |
| DDE(o,p') | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| DDE(p,p') | ng/L | 0.17 | 0.51 1 | 8 5.30 | 12.71 20 |
| DDMU(p,p') | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| DDT(o,p') | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| DDT(p,p') | ng/L | 0.11 | 0.47 1 | 8 0.30 | 0.92 20 |
| DDTs | ng/L | 0.28 | 0.96 1 | 8 5.65 | 12.75 20 |
| delta-BHC | ng/L | 0 | | 1 0 | 1 |
| Demeton-s | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| Diazinon | ng/L | 12.64 | 14.90 1 | 7 50.81 | 51.65 19 |
| Dichlofenthion | ng/L | 0 | 0 1 | 7 0 | 0 19 |
| Dichlorvos | ng/L | 0 | | 7 0 | 0 19 |
| Dicrotophos | ng/L | 0 | 0 1 | | 17.90 19 |
| Dieldrin | ng/L | 0.11 | 0.32 1 | | 0 20 |
| Dimethoate | ng/L | 0 | 0 1 | | 35.77 19 |
| Dioxathion | ng/L | 0 | 0 1 | | 18.10 19 |

| Appendix, continued. | | San Di | | | | Peñasquitos |
|----------------------|--------------|---------|-------|----|--------|-------------|
| Pesticides | Symbol Units | Mean SI | | | | SD n |
| Disulfoton | ng/L | 0 | 0 | 17 | 52.92 | 64.68 19 |
| Endosulfan I | ng/L | 0.06 | 0.24 | | 0.05 | 0.22 20 |
| Endosulfan II | ng/L | 0 | | 18 | 0.54 | 1.54 20 |
| Endosulfan sulfate | ng/L | 0.06 | 0.24 | | 0.05 | 0.22 20 |
| Endrin | ng/L | 0 | | 18 | 0.05 | 0.22 20 |
| Endrin Aldehyde | ng/L | 0 | | 18 | 0.15 | 0.67 20 |
| Endrin Ketone | ng/L | 0 | | 17 | 0 | 0 19 |
| Ethion | ng/L | 0 | 0 | 17 | 0 | 0 19 |
| Ethoprop | ng/L | 0 | | 17 | 2.11 | 9.18 19 |
| Famphur | ng/L | 0 | | 17 | 0 | 0 19 |
| Fenchlorphos | ng/L | 0 | 0 | 17 | 0 | 0 19 |
| Fenitrothion | ng/L | 0 | 0 | 17 | 0 | 0 19 |
| Fensulfothion | ng/L | 0 | 0 | 17 | 0 | 0 19 |
| Fenthion | ng/L | 0 | 0 | 17 | 4.21 | 12.61 19 |
| Fonofos | ng/L | 1.57 | 6.48 | 17 | 0 | 0 19 |
| gamma-BHC (Lindane) | ng/L | 0 | | 1 | 0 | 1 |
| HCH, alpha | ng/L | 0 | 0 | 17 | 2.05 | 8.71 19 |
| HCH, beta | ng/L | 0 | 0 | 17 | 0.42 | 1.39 19 |
| HCH, delta | ng/L | 0.06 | 0.24 | 17 | 0.05 | 0.23 19 |
| HCH, gamma | ng/L | 0.06 | 0.24 | 17 | 0.05 | 0.23 19 |
| Heptachlor | ng/L | 0 | 0 | 18 | 0 | 0 20 |
| Heptachlor epoxide | ng/L | 0.11 | 0.32 | 18 | 0 | 0 20 |
| Hexachlorobenzene | ng/L | 0 | | 17 | 0.15 | 0.29 19 |
| Leptophos | ng/L | 0 | | 17 | 0 | 0 19 |
| Malathion | ng/L | 0 | | 17 | 18.95 | 82.59 19 |
| Merphos | ng/L | 0 | | 17 | 0 | 0 19 |
| Methidathion | ng/L | 0 | | 17 | 0 | 0 19 |
| Methoxychlor | ng/L | 0 | | 18 | 0.05 | 0.22 20 |
| Mevinphos | ng/L | 0 | | 17 | 8.97 | 17.97 19 |
| Mirex | ng/L | 0.06 | 0.24 | | 0 | 0 19 |
| Molinate | ng/L | 0 | | 17 | 15.79 | 37.46 19 |
| Naled | ng/L | 0 | 0 | 17 | 10.53 | 18.10 19 |
| Nonachlor, cis- | ng/L | 0 | | 17 | 0 | 0 19 |
| Nonachlor, trans- | ng/L | 0 | | 17 | 0 | 0 19 |
| Oxadiazon | ng/L | 8.65 | 15.04 | | 47.01 | 44.26 19 |
| Oxychlordane | ng/L | 0 | | 17 | 0 | 0 19 |
| Parathion, Ethyl | ng/L | 0 | 0 | 17 | 2.11 | 9.18 19 |
| Parathion, Methyl | ng/L | 0 | | 17 | 8.26 | 23.47 19 |
| Phorate | ng/L | 0 | | 17 | 0.20 | 0 19 |
| Phosmet | ng/L | 0 | | 17 | 0 | 0 19 |
| Phosphamidon | ng/L | 0 | | 17 | 0 | 0 19 |
| Prometon | ng/L | 0 | | 17 | 0 | 0 19 |
| Prometryn | | 0 | | 17 | 0 | 0 19 |
| Propazine | ng/L ng/L | 0 | | 17 | 17.11 | 35.25 19 |
| Secbumeton | | | | 17 | 131.16 | 147.98 19 |
| | ng/L | 0 | | | | |
| Simazine | ng/L | 0 | | 17 | 0 | 0 19 |
| Simetryn | ng/L | 0 | | 17 | 0 | 0 19 |
| Sulfotep | ng/L | 0 | | 17 | 0 | 0 19 |
| Tedion | ng/L | 0.12 | 0.33 | 17 | 0 | 0 19 |

| | | San Diegu | uito | Los Peñasquitos | | |
|-------------------|--------------|-----------|------|-----------------|-----------|--|
| Pesticides | Symbol Units | Mean SD | n | Mean | SD n | |
| Terbufos | ng/L | 0 | 0 17 | 0 | 0 19 | |
| Terbuthylazine | ng/L | 0 | 0 17 | 380.37 | 373.84 19 | |
| Terbutryn | ng/L | 0 | 0 17 | 0 | 0 19 | |
| Tetrachlorvinphos | ng/L | 0 | 0 17 | 0 | 0 19 | |
| Thiobencarb | ng/L | 0 | 0 17 | 69.74 | 218.99 19 | |
| Thionazin | ng/L | 0 | 0 17 | 0 | 0 19 | |
| Tokuthion | ng/L | 0 | 0 17 | 2.11 | 9.18 19 | |
| Toxaphene | ng/L | 0 | 1 | 0 | 1 | |
| Trichlorfon | ng/L | 0 | 0 17 | 0 | 0 19 | |
| Trichloronate | ng/L | 0 | 0 17 | 2.11 | 9.18 19 | |

| Appendix, continued. | | | Sar | Diego | | Pueblo | San Diego |
|---------------------------------------|----------|------------|-------|-------|----|--------|-----------|
| Physical water quality and inorganics | Symbol | Units | | SD | n | Mean | SD n |
| Alkalinity as CaCO3 | Alk | mg/l | 242 | 85 | 27 | 191 | 15 5 |
| Chloride | | mg/l | 330 | | 1 | | 0 |
| Fine-ASTM | | % | | | 0 | | 0 |
| Fine-ASTM, Passing No. 200 Sieve | | % | | | 0 | | 0 |
| Oxygen, Dissolved | DO | mg/L | 9.5 | 3.0 | 26 | 113.3 | 198.9 4 |
| Oxygen, Saturation | | % | 102 | 32 | 26 | 175 | 98 4 |
| pH | | pН | 8.0 | 0.3 | 26 | 8.8 | 0.8 4 |
| Salinity | Sal | ppt | 1.00 | 0.45 | 26 | | 5.66 4 |
| Specific conductivity | Cond | μS/cm | 1872 | 830 | 26 | 6925 | 9619 4 |
| Sulfate | SO_4 | mg/l | 283 | 159 | 27 | 437 | 394 5 |
| Suspended Sediment Concentration | | % | | | 0 | | 0 |
| Temperature | | °C | 19.3 | 4.1 | 26 | 25.7 | 5.4 4 |
| Total Organic Carbon | | mg/L | | | 0 | | 0 |
| Total Suspended Solids | | mg/L | 34 | 104 | 27 | 38 | 42 5 |
| Turbidity | | NTU | 7.2 | 8.9 | 26 | 6.2 | 7.4 4 |
| Velocity | | ft/s | 0.8 | 1.1 | 27 | 4.0 | 8.9 5 |
| Nutrients | | | | | | | |
| Ammonia as N | NH_3-N | mg/l | 0.05 | 0.05 | 27 | 0.05 | 0.05 5 |
| Nitrate + Nitrite as N | | mg/l | 15.64 | 57.07 | 27 | 0.41 | 0.47 5 |
| Nitrate as N | | mg/l | 15.62 | 57.08 | 27 | 0.39 | 0.45 5 |
| Nitrate as NO3 | NO_3 | mg/l | | | 0 | 0 | 1 |
| Nitrite as N | | mg/l | 0.03 | 0.04 | 27 | 0.02 | 0.02 5 |
| Nitrogen, Total Kjeldahl | | mg/l | 0.79 | 0.36 | 26 | 1.81 | 0.45 5 |
| OrthoPhosphate as P | | mg/l | 0.10 | | 1 | 32.00 | 1 |
| Phosphorus as P,Total | TP | mg/l | 0.12 | 0.09 | 27 | 0.25 | 0.07 5 |
| Metals | | | | | | | |
| Aluminum | | μg/L | 7.9 | 9.0 | 27 | 7.4 | 9.4 5 |
| Arsenic | As | μg/L | 3.2 | 2.4 | 27 | 2.5 | 1.5 5 |
| Cadmium | Cd | μg/L | 0.03 | 0.02 | | | 0.05 5 |
| Chromium | Cr | μg/L | 0.64 | 0.74 | | 1.22 | 0.79 5 |
| Copper | Cu | μg/L | 4.16 | 2.06 | | 8.23 | 3.94 5 |
| Lead | Pb | μg/L | 0.09 | 0.08 | 27 | 0.51 | 0.28 5 |
| Manganese | Mn | μg/L | 60 | 116 | | | 67 5 |
| Nickel | Ni | μg/L | 1.15 | 1.92 | | | 2.81 5 |
| Selenium | Se | μg/L | 8.1 | 6.8 | | | 115.8 4 |
| Silver | Ag | μg/L | 0.00 | 0.00 | | | 0.27 5 |
| Zinc | Zi | μg/L | 3.9 | 2.3 | 26 | 13.6 | 4.1 4 |
| Bacteria | | | | | | | |
| Enterococcus | | MPN/100 ml | | | 1 | 11 | 1 |
| Fecal Coliform | | MPN/100 ml | | | 1 | 900 | 1 |
| Total Coliform | | MPN/100 ml | 1600 | | 1 | 1600 | 1 |

| | | San Di | | | | San Diego |
|-------------------------------|--------------|---------|------|--------|-------|-----------|
| PAHs | Symbol Units | Mean SD | | | | SD n |
| Acenaphthene | ng/L | 1 | | 27 | 0 | 0 5 |
| Acenaphthylene | ng/L | 0 | | 27 | 0 | 0 5 |
| Anthracene | ng/L | 0 | | 27 | | 0 5 |
| Benz(a)anthracene | ng/L | 0 | | 27 | | 7 5 |
| Benzo(a)pyrene | ng/L | 0.6 | 3.0 | 27 | 4.6 | 10.3 5 |
| Benzo(b)fluoranthene | ng/L | 0.8 | 4.0 | 27 | 5.8 | 13.0 5 |
| Benzo(e)pyrene | ng/L | 0.4 | 2.3 | 27 | 0 | 0.5 |
| Benzo(g,h,i)perylene | ng/L | 0.8 | 4.3 | 27 | 5.9 | 13.2 5 |
| Benzo(k)fluoranthene | ng/L | 0 | 0 | 26 | 7.3 | 14.5 |
| Biphenyl | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Chrysene | ng/L | 0.3 | 1.5 | 27 | 4.3 | 9.6 5 |
| Chrysenes, C1 - | ng/L | 0 | | 26 | | 0 4 |
| Chrysenes, C2 - | ng/L | 0 | | 26 | | 0 4 |
| Chrysenes, C3 - | ng/L | 0 | | 26 | | 0 4 |
| Dibenz(a,h)anthracene | ng/L | 0 | | 27 | | 14.4 5 |
| Dibenzothiophene | ng/L | 0 | | 26 | 3.7 | 7.5 |
| Dibenzothiophenes, C1 - | ng/L | 0 | | 26 | 18.4 | 32.9 |
| Dibenzothiophenes, C2 - | ng/L | 0 | | 26 | 35.6 | 65.5 |
| Dibenzothiophenes, C3 - | ng/L | 0 | | 26 | 1.4 | 2.8 |
| Dimethylnaphthalene, 2,6- | ng/L | 0 | | 26 | 4 | 7 4 |
| Dimethylphenanthrene, 3,6- | ng/L | 0 | | 19 | 0 | 0 4 |
| Fluoranthene | _ | 0.68 | 3.54 | | | 6.69 |
| | ng/L | 0.60 | 3.04 | | | 0.09 3 |
| Fluoranthene/Pyrenes, C1 - | ng/L | | | | 0 | |
| Fluorene | ng/L | 0 | | 27 | | 0 ! |
| Fluorenes, C1 - | ng/L | 0 | | 26 | 0 | 0 4 |
| Fluorenes, C2 - | ng/L | 0 | | 26 | | 0 4 |
| Fluorenes, C3 - | ng/L | 0 | | 26 | | 34.75 |
| Indeno(1,2,3-c,d)pyrene | ng/L | 0.83 | 4.33 | | | 15.97 |
| Methyldibenzothiophene, 4- | ng/L | 0 | | 19 | 2.38 | 4.76 |
| Methylfluoranthene, 2- | ng/L | 0 | | 19 | 0 | 0 4 |
| Methylfluorene, 1- | ng/L | 0 | | 19 | 0 | 0 4 |
| Methylnaphthalene, 1- | ng/L | 0 | | 26 | 0 | 0 4 |
| Methylnaphthalene, 2- | ng/L | 0 | | 26 | 0 | 0 4 |
| Methylphenanthrene, 1- | ng/L | 0 | | 26 | 0 | 0 4 |
| Naphthalene | ng/L | 0.19 | 0.97 | | 1.23 | 2.76 |
| Naphthalenes, C1 - | ng/L | 0 | | 26 | 0 | 0 4 |
| Naphthalenes, C2 - | ng/L | 0.51 | 1.80 | | 4.93 | 9.85 |
| Naphthalenes, C3 - | ng/L | 1.15 | 3.31 | 26 | 7.43 | 8.64 4 |
| Naphthalenes, C4 - | ng/L | 0 | 0 | 26 | 2.90 | 5.80 4 |
| Perylene | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Phenanthrene | ng/L | 0 | 1 | 27 | 5 | 7 5 |
| Phenanthrene/Anthracene, C1 - | ng/L | 0.77 | 2.72 | 26 | 6.90 | 8.29 |
| Phenanthrene/Anthracene, C2 - | ng/L | 0.63 | 2.23 | | 11.00 | 14.07 |
| Phenanthrene/Anthracene, C3 - | ng/L | 0 | | 26 | 2.55 | 5.10 4 |
| Phenanthrene/Anthracene, C4 - | ng/L | 0 | | 26 | 0 | 0 4 |
| Pyrene | ng/L | 0.77 | 3.98 | | 1.45 | 3.23 5 |
| Trimethylnaphthalene, 2,3,5- | ng/L | 0 | | 26 | 0 | 0 4 |

| | | San Dieg | go P | ueblo San | Diego |
|--------------------|--------------|----------|------|-----------|-------|
| PCBs | Symbol Units | Mean SD | | ean SD | n |
| PCB 005 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 008 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 015 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 018 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 027 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 028 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 029 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 031 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 033 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 044 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 049 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 052 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 056 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 060 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 066 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 070 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 074 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 087 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 095 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 097 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 099 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 101 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 105 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 110 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 114 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 118 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 128 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 137 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 138 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 141 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 149 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 151 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 153 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 156 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 157 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 158 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 170 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 174 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 177 | ng/L | Ö | 0 26 | 0 | 0 4 |
| PCB 180 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 183 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 187 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 189 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 194 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 195 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 200 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 200 PCB 201 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 201 PCB 203 | ng/L | 0 | 0 26 | 0 | 0 4 |
| PCB 203 | ng/L | 0 | 0 26 | 0 | 0 4 |

| | | | Diego | Pueblo San Diego | | |
|---------------------|--------------|--------|-------|------------------|---------|------------|
| PCBs | Symbol Units | Mean S | | | Mean SE | |
| PCB 209 | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| PCB-1016 | ng/L | 0 | | 1 | 0 | 1 |
| PCB-1221 | ng/L | 0 | | 1 | 0 | 1 |
| PCB-1232 | ng/L | 0 | | 1 | 0 | 1 |
| PCB-1242 | ng/L | 0 | | 1 | 0 | 1 |
| PCB-1248 | ng/L | 0 | | 1 | 0 | 1 |
| PCB-1254 | ng/L | 0 | | 1 | 0 | 1 |
| PCB-1260 | ng/L | 0 | | 1 | 0 | 1 |
| PCBs | ng/L | 0 | 0 | 27 | 0 | 0 5 |
| Pesticides | | | | | | |
| Aldrin | ng/L | 0 | 0 | 27 | 0 | 0 5 |
| alpha-BHC | ng/L | 0 | | 1 | 0 | 1 |
| Ametryn | ng/L | 0 | 0 | 26 | | 0 |
| Aspon | ng/L | 0 | | 26 | 0 | 0 4 |
| Atraton | ng/L | 0 | | 26 | _ | 0 |
| Atrazine | ng/L | 3.27 | 16.67 | | | 0 |
| Azinphos ethyl | ng/L | 0 | | 26 | 0 | 0 4 |
| Azinphos methyl | ng/L | 0 | | 26 | 0 | 0 4 |
| beta-BHC | ng/L | 0 | Ū | 1 | 0 | 1 |
| Bolstar | ng/L | 0 | Λ | 26 | 0 | 0 4 |
| Carbophenothion | ng/L | 0 | | 26 | 0 | 0 4 |
| Chlordane (tech) | | 0 | U | 1 | 0 | 1 |
| ` , | ng/L | | 0 | 26 | 0 | |
| Chlordane, cis- | ng/L | 0 0 | | 26 26 | 0 | 0 4 0 4 |
| Chlordane, trans- | ng/L | | | 26 26 | | |
| Chlordene, alpha- | ng/L | 0 | | | 0 | 0 4 |
| Chlordene, gamma- | ng/L | 0 | | 26 | 0 | 0 4 |
| Chlorfenvinphos | ng/L | 0 | | 26 | 0 | 0 4 |
| Chlorpyrifos | ng/L | 0 | | 26 | 0 | 0 4 |
| Chlorpyrifos methyl | ng/L | 0 | | 26 | 0 | 0 4 |
| Ciodrin | ng/L | 0 | | 26 | 0 | 0 4 |
| Coumaphos | ng/L | 0 | | 26 | 0 | 0 4 |
| Dacthal | ng/L | 0 | | 26 | 0 | 0 4 |
| DDD(o,p') | ng/L | 0 | | 26 | 0 | 0 4 |
| DDD(p,p') | ng/L | 0 | | 27 | 0 | 0 5 |
| DDE(o,p') | ng/L | 0 | | 26 | 0 | 0 4 |
| DDE(p,p') | ng/L | 0 | | 27 | 0 | 0 5 |
| DDMU(p,p') | ng/L | 0 | | 26 | 0 | 0 4 |
| DDT(o,p') | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| DDT(p,p') | ng/L | 0 | | 27 | 0 | 0 5 |
| DDTs | ng/L | 0 | 0 | 27 | 0 | 0 5 |
| delta-BHC | ng/L | 0 | | 1 | 0 | 1 |
| Demeton-s | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Diazinon | ng/L | 1.73 | 5.50 | | 26.25 | 29.17 4 |
| Dichlofenthion | ng/L | 0 | | 26 | 0 | 0 4 |
| Dichlorvos | ng/L | 0 | | 26 | 0 | 0 4 |
| Dicrotophos | ng/L | 0 | | 26 | 0 | 0 4 |
| Dieldrin | ng/L | 0 | | 27 | 0 | 0 5 |
| Dimethoate | ng/L | 0 | | 26 | 0 | 0 3 |
| Dinethoate | ng/L | 0 | | 26 26 | 73.50 | 94.18 4 |

| Appendix, continued. | | San Diego | | | Pueblo S | San Diego | |
|----------------------|--------------|-----------|-------|----|----------|-----------|--|
| Pesticides | Symbol Units | Mean S | D | n | | SD n | |
| Disulfoton | ng/L | 0 | | 26 | | 26.50 4 | |
| Endosulfan I | ng/L | 0 | | 27 | | 0 5 | |
| Endosulfan II | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Endosulfan sulfate | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Endrin | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Endrin Aldehyde | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Endrin Ketone | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Ethion | ng/L | 0 | | 26 | | 0 4 | |
| Ethoprop | ng/L | 0 | | 26 | | 0 4 | |
| Famphur | ng/L | 0 | | 26 | | 0 4 | |
| Fenchlorphos | ng/L | 0 | | 26 | | 0 4 | |
| Fenitrothion | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Fensulfothion | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Fenthion | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Fonofos | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| gamma-BHC (Lindane) | ng/L | 0 | | 1 | 0 | 1 | |
| HCH, alpha | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| HCH, beta | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| HCH, delta | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| HCH, gamma | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Heptachlor | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Heptachlor epoxide | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Hexachlorobenzene | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Leptophos | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Malathion | ng/L | 1.27 | 6.47 | 26 | 0 | 0 4 | |
| Merphos | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Methidathion | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Methoxychlor | ng/L | 0 | 0 | 27 | 0 | 0 5 | |
| Mevinphos | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Mirex | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Molinate | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Naled | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Nonachlor, cis- | ng/L | 0 | | 26 | | 0 4 | |
| Nonachlor, trans- | ng/L | 0 | | 26 | | 0 4 | |
| Oxadiazon | ng/L | 7.23 | 9.52 | 26 | 19.00 | 15.53 4 | |
| Oxychlordane | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Parathion, Ethyl | ng/L | 0 | | 26 | | 0 4 | |
| Parathion, Methyl | ng/L | 0 | 0 | 26 | 0 | 0 4 | |
| Phorate | ng/L | 0 | | 26 | | 0 4 | |
| Phosmet | ng/L | 0 | | 26 | | 0 4 | |
| Phosphamidon | ng/L | 0 | | 26 | | 0 4 | |
| Prometon | ng/L | 0 | | 26 | | 0 | |
| Prometryn | ng/L | 0 | | 26 | | 0 | |
| Propazine | ng/L | 8.23 | 26.98 | | | 0 | |
| Secbumeton | ng/L | 12.92 | 48.18 | | | 0 | |
| Simazine | ng/L | 10.46 | 25.68 | | | 0 | |
| Simetryn | ng/L | 0 | | 26 | | 0 | |
| Sulfotep | ng/L | 0 | | 26 | | 0 4 | |
| Tedion | ng/L | 0 | | 26 | | 0 4 | |

| | | San Diego | | | | |
|-------------------|--------------|-----------|-------|----|------|------|
| Pesticides | Symbol Units | Mean | SD | n | Mean | SD n |
| Terbufos | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Terbuthylazine | ng/L | 16.42 | 64.61 | 26 | | 0 |
| Terbutryn | ng/L | 0 | 0 | 26 | | 0 |
| Tetrachlorvinphos | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Thiobencarb | ng/L | 4.46 | 22.75 | 26 | 0 | 0 4 |
| Thionazin | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Tokuthion | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Toxaphene | ng/L | 0 | | 1 | 0 | 1 |
| Trichlorfon | ng/L | 0 | 0 | 26 | 0 | 0 4 |
| Trichloronate | ng/L | 0 | 0 | 26 | 0 | 0 4 |

| Appendix, continued. | | | Sweetwater | | Otay | | Tijuana | | | |
|---------------------------------------|--------------------|------------|------------|------|------|------|---------|-------|-------|----|
| Physical water quality and inorganics | Symbol | Units | Mean | | | Mean | | | SD | n |
| Alkalinity as CaCO3 | Alk | mg/l | 191 | 92 | 15 | 233 | 21 7 | 400 | 164 | 12 |
| Chloride | | mg/l | 1740 | 622 | 2 | | 0 | | | 0 |
| Fine-ASTM | | % | | | 0 | | 0 | | | 0 |
| Fine-ASTM, Passing No. 200 Sieve | | % | | | 0 | 2.3 | 1 | | | 0 |
| Oxygen, Dissolved | DO | mg/L | 9.2 | 2.0 | 14 | | 0 | 8.3 | 4.3 | 12 |
| Oxygen, Saturation | | % | 99 | 25 | 14 | 123 | 31 5 | 90 | 51 | 12 |
| pH | | pН | 8.0 | 0.4 | 14 | 7.8 | 0.3 5 | 7.9 | 0.7 | 12 |
| Salinity | Sal | ppt | 1.85 | 1.57 | 14 | 1.30 | 0.85 5 | 0.76 | 0.46 | 12 |
| Specific conductivity | Cond | μS/cm | 3930 | 3649 | 14 | 2478 | 1539 5 | 1482 | 887 | 12 |
| Sulfate | SO_4 | mg/l | 276 | 190 | 15 | 217 | 40 7 | 201 | 133 | 12 |
| Suspended Sediment Concentration | | % | | | 0 | | 0 | | | 0 |
| Temperature | | °C | 17.9 | 6.7 | 14 | 15.5 | 2.1 5 | 19.4 | 7.0 | 12 |
| Total Organic Carbon | | mg/L | | | 0 | | 0 | | | 0 |
| Total Suspended Solids | | mg/L | 7 | 6 | 14 | 22 | 37 7 | 36 | 51 | 12 |
| Turbidity | | NTU | 4.1 | 5.0 | 14 | 2.5 | 1.4 5 | 12.3 | 13.3 | 12 |
| Velocity | | ft/s | 1.1 | 1.2 | 15 | 1.2 | 1.8 6 | 0.8 | 1.2 | 12 |
| Nutrients | | | | | | | | | | |
| Ammonia as N | NH ₃ -N | mg/l | 0.06 | 0.05 | 15 | 0.08 | 0.10 7 | 11.33 | 16.63 | 12 |
| Nitrate + Nitrite as N | | mg/l | 5.21 | 8.04 | 15 | 2.11 | 2.11 7 | 0.23 | 0.32 | 12 |
| Nitrate as N | | mg/l | 5.19 | 8.01 | 15 | 2.08 | 2.07 7 | 0.18 | 0.33 | 12 |
| Nitrate as NO3 | NO_3 | mg/l | | | 0 | 9.21 | 9.14 7 | | | 0 |
| Nitrite as N | | mg/l | 0.02 | 0.03 | 15 | 0.03 | 0.04 7 | 0.05 | 0.11 | 12 |
| Nitrogen, Total Kjeldahl | | mg/l | 0.72 | 0.35 | | 0.73 | 0.37 7 | 13.00 | 15.53 | 12 |
| OrthoPhosphate as P | | mg/l | 0.02 | | 1 | 2.01 | 5.29 7 | | | 0 |
| Phosphorus as P,Total | TP | mg/l | 0.06 | 0.04 | 15 | 0.01 | 0.02 7 | 3.31 | 3.70 | 12 |
| Metals | | Ü | | | | | | | | |
| Aluminum | | μg/L | 5.5 | 8.2 | 15 | 1.3 | 2.0 7 | 5.7 | 5.1 | 12 |
| Arsenic | As | μg/L | 11.5 | 16.8 | 15 | 7.7 | 6.0 7 | 3.7 | 2.6 | 12 |
| Cadmium | Cd | μg/L | 0.02 | 0.02 | 15 | 0.02 | 0.01 7 | 0.06 | 0.03 | 12 |
| Chromium | Cr | μg/L | 0.96 | 1.06 | 15 | 0.37 | 0.30 7 | 2.67 | 2.94 | 12 |
| Copper | Cu | μg/L | 4.25 | 3.05 | 15 | 2.91 | 1.16 7 | 5.14 | 6.25 | 12 |
| Lead | Pb | μg/L | 0.07 | 0.13 | 15 | 0.02 | 0.02 7 | 0.25 | 0.27 | 12 |
| Manganese | Mn | μg/L | 54 | 71 | 15 | 41 | 67 7 | 238 | 228 | 12 |
| Nickel | Ni | μg/L | 0.78 | 0.93 | 15 | 1.80 | 3.22 7 | 9.16 | 11.13 | 12 |
| Selenium | Se | μg/L | 26.6 | 27.8 | 14 | 9.2 | 7.3 6 | 7.2 | 4.6 | 12 |
| Silver | Ag | μg/L | 0.00 | 0.00 | 15 | 0.39 | 1.02 7 | 0.02 | 0.04 | 12 |
| Zinc | Zi | μg/L | 2.9 | 2.0 | 14 | 2.0 | 0.8 6 | 4.5 | 6.6 | 12 |
| Bacteria | | | | | | | | | | |
| Enterococcus | | MPN/100 ml | 2400 | | 1 | 210 | 1 | | | 0 |
| Fecal Coliform | | MPN/100 ml | 900 | | 1 | 500 | 1 | | | 0 |
| Total Coliform | | MPN/100 ml | 1600 | | 1 | 1600 | 1 | | | 0 |

| Appendix, continued. | | Swe | etwate | r | 0 | tay | | Ti | juana | |
|-------------------------------|--------------|------|--------|----|------|-----|---|--------|--------|------|
| PAHs | Symbol Units | Mean | SD | n | Mean | SĎ | n | | SD | n |
| Acenaphthene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 1 | 4 | 12 |
| Acenaphthylene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 1 | 4 | - 12 |
| Anthracene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 1 | 3 | 12 |
| Benz(a)anthracene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 3 | 12 | 12 |
| Benzo(a)pyrene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 6.6 | 23.0 | 12 |
| Benzo(b)fluoranthene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 8.8 | 30.3 | 12 |
| Benzo(e)pyrene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 12.2 | 42.1 | 12 |
| Benzo(g,h,i)perylene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 13.8 | 47.6 | 12 |
| Benzo(k)fluoranthene | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 2.8 | 9.6 | 12 |
| Biphenyl | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 7 | 16 | 12 |
| Chrysene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 9.7 | 33.5 | 12 |
| Chrysenes, C1 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 18.9 | 65.5 | 12 |
| Chrysenes, C2 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 24.3 | 78.5 | 12 |
| Chrysenes, C3 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 23.3 | 80.8 | 12 |
| Dibenz(a,h)anthracene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 1.7 | 5.9 | 12 |
| Dibenzothiophene | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 3.5 | 9.8 | 12 |
| Dibenzothiophenes, C1 - | ng/L | 0 | 0 | 14 | 3.6 | 5.6 | 6 | 35.1 | 82.8 | 12 |
| Dibenzothiophenes, C2 - | ng/L | 0 | 0 | 14 | 7.4 | 8.6 | 6 | 95.5 | 271.1 | 12 |
| Dibenzothiophenes, C3 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 89.0 | 264.3 | 12 |
| Dimethylnaphthalene, 2,6- | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 31 | 80 | 12 |
| Dimethylphenanthrene, 3,6- | ng/L | 0 | 0 | 14 | | | 0 | 14.02 | 43.42 | 12 |
| Fluoranthene | ng/L | 0.53 | 2.06 | 15 | 0 | 0 | 7 | 26.98 | 71.43 | 12 |
| Fluoranthene/Pyrenes, C1 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 58.74 | 177.90 | 12 |
| Fluorene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 7 | 22 | 12 |
| Fluorenes, C1 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 43.72 | 121.07 | 12 |
| Fluorenes, C2 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 | 12 |
| Fluorenes, C3 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 | 12 |
| Indeno(1,2,3-c,d)pyrene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 4.53 | 15.70 | 12 |
| Methyldibenzothiophene, 4- | ng/L | 0 | 0 | 14 | | | 0 | 12.58 | 31.73 | 12 |
| Methylfluoranthene, 2- | ng/L | 0 | 0 | 14 | | | 0 | 5.89 | 20.41 | 12 |
| Methylfluorene, 1- | ng/L | 0 | 0 | 14 | | | 0 | 16.57 | 44.06 | 12 |
| Methylnaphthalene, 1- | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 10 | 32 | 12 |
| Methylnaphthalene, 2- | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 15 | 51 | 12 |
| Methylphenanthrene, 1- | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 11 | 29 | 12 |
| Naphthalene | ng/L | 0.47 | 1.82 | 15 | 0 | 0 | 7 | 9.83 | 34.06 | 12 |
| Naphthalenes, C1 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 26.53 | 84.37 | 12 |
| Naphthalenes, C2 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 83.41 | 240.10 | 12 |
| Naphthalenes, C3 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 155.11 | 427.76 | 12 |
| Naphthalenes, C4 - | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 48.74 | 120.60 | 12 |
| Perylene | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 2.58 | 8.95 | 12 |
| Phenanthrene | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 26 | 57 | 12 |
| Phenanthrene/Anthracene, C1 - | ng/L | 0.36 | 1.34 | 14 | 0 | 0 | 6 | 69.21 | 157.09 | 12 |
| Phenanthrene/Anthracene, C2 - | ng/L | 0.41 | 1.53 | 14 | 0 | 0 | 6 | 157.18 | 438.58 | 12 |
| Phenanthrene/Anthracene, C3 - | ng/L | 0 | | 14 | Ö | | | 109.16 | | |
| Phenanthrene/Anthracene, C4 - | ng/L | 0 | | 14 | 0 | 0 | | 27.80 | 90.90 | |
| Pyrene | ng/L | 0.44 | 1.70 | | 0 | 0 | | 26.23 | 70.99 | |
| Trimethylnaphthalene, 2,3,5- | ng/L | 0 | | 14 | 0 | 0 | | 19 | | 12 |

| | | Sweetw | ater | Otay | | Tijuana | а |
|--------------------|--------------|---------|------|---------|-----|---------|------|
| PCBs | Symbol Units | Mean SD | n | Mean SD | | 1ean SD | n |
| PCB 005 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 008 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 015 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 018 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 027 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 028 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 029 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 031 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 033 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 044 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 049 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 052 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 056 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 060 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 066 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 070 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 074 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 087 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 095 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 097 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 099 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| PCB 101 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 105 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 110 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 114 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 118 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 128 | ng/L | 0 | 0 14 | | 0 6 | Ö | 0 12 |
| PCB 137 | ng/L | 0 | 0 14 | | 0 6 | Ö | 0 12 |
| PCB 138 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 141 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 149 | ng/L | Ö | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 151 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 153 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 156 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 157 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 158 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 170 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 174 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 174 PCB 177 | | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 177 PCB 180 | ng/L | 0 | 0 14 | | 0 6 | | 0 12 |
| PCB 183 | ng/L | | | | | 0 | 0 12 |
| | ng/L | 0 | 0 14 | _ | 0 6 | 0 | |
| PCB 187 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 189 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 194 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 195 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 200 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 201 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 203 | ng/L | 0 | 0 14 | | 0 6 | 0 | 0 12 |
| PCB 206 | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |

| Appendix, continued. | | Swe | etwate | r | (| Otay | Т | juana |
|----------------------|--------------|-------|--------|----|------|--------|-------|-----------|
| PCBs | Symbol Units | Mean | SD | n | Mean | SD n | Mean | SD n |
| PCB 209 | ng/L | 0 | 0 | 14 | 0 | 0 6 | 6 0 | 0 12 |
| PCB-1016 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCB-1221 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCB-1232 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCB-1242 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCB-1248 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCB-1254 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCB-1260 | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| PCBs | ng/L | 0 | 0 | 15 | 0 | 0 7 | ' 0 | 0 12 |
| Pesticides | G | | | | | | | |
| Aldrin | ng/L | 0 | 0 | 15 | 0 | 0 7 | ' 0 | 0 12 |
| alpha-BHC | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| Ametryn | ng/L | | | 0 | 0 | 0.6 | | 0 |
| Aspon | ng/L | 0 | 0 | 14 | 0 | 0.6 | | 0 12 |
| Atraton | ng/L | · | · | 0 | 0 | 0 6 | | 0 |
| Atrazine | ng/L | | | 0 | 6.80 | | | 0 |
| Azinphos ethyl | ng/L | 0 | 0 | 14 | 0.00 | 0 6 | | 0 12 |
| Azinphos methyl | ng/L | 0 | | 14 | 0 | 0.6 | | 0 12 |
| beta-BHC | ng/L | 0 | U | 1 | 0 | 1 | | 0 12 |
| Bolstar | _ | 0 | 0 | 14 | 0 | 0 6 | | 0 12 |
| Carbophenothion | ng/L ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| • | | 0 | U | 14 | | 1 | | 0 12 |
| Chlordane (tech) | ng/L | | 0 | | 0 | | | |
| Chlordane, cis- | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Chlordane, trans- | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Chlordene, alpha- | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Chlordene, gamma- | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Chlorfenvinphos | ng/L | 0 | | 14 | 0 | 0.6 | | 0 12 |
| Chlorpyrifos | ng/L | 0 | | 14 | 0 | 0.6 | | 6.93 12 |
| Chlorpyrifos methyl | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Ciodrin | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Coumaphos | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Dacthal | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| DDD(o,p') | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| DDD(p,p') | ng/L | 0 | | 15 | 0 | 0 7 | | 0 12 |
| DDE(o,p') | ng/L | 0 | 0 | 14 | 0 | 0 6 | | 0 12 |
| DDE(p,p') | ng/L | 0 | 0 | 15 | 1.43 | 2.51 7 | ' 0 | 0 12 |
| DDMU(p,p') | ng/L | 0 | 0 | 14 | 0 | 0 6 | 0 | 0 12 |
| DDT(o,p') | ng/L | 0 | 0 | 14 | 0 | 0.6 | 0 | 0 12 |
| DDT(p,p') | ng/L | 0 | 0 | 15 | 0.57 | 0.98 7 | ' 0 | 0 12 |
| DDTs | ng/L | 0 | 0 | 15 | 2.00 | 2.83 7 | ' 0 | 0 12 |
| delta-BHC | ng/L | 0 | | 1 | 0 | 1 | | 0 |
| Demeton-s | ng/L | 0 | 0 | 14 | | 0.6 | 0 | 0 12 |
| Diazinon | ng/L | | | | | 26.47 | | 20.93 12 |
| Dichlofenthion | ng/L | 0 | | 14 | 0 | 0.6 | | 0 12 |
| Dichlorvos | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Dicrotophos | ng/L | 0 | | 14 | 0 | 0 6 | | 0 12 |
| Dieldrin | ng/L | 0 | | 15 | 0 | 0 7 | | 0 12 |
| Dimethoate | | 0 | | 14 | 0 | 0 6 | | 0 12 |
| | ng/L | | | | | | | |
| Dioxathion | ng/L | 34.43 | 52.78 | 14 | 0 | 0.6 | 90.07 | 231.08 12 |

| | | | etwate | r | | ay | _ | | juana |
|---------------------|--------------|------|--------|----|--------|-------|---|-------|-------|
| Pesticides | Symbol Units | Mean | | n | Mean S | | | | SD r |
| Disulfoton | ng/L | | 15.50 | | 0 | | 6 | 16.17 | 40.76 |
| Endosulfan I | ng/L | 0 | | 15 | 0 | | 7 | 0 | 0 |
| Endosulfan II | ng/L | 0 | | 15 | 0 | | 7 | 0 | 0 |
| Endosulfan sulfate | ng/L | 0 | | 15 | 0 | | 7 | 0 | 0 |
| Endrin | ng/L | 0 | 0 | 15 | 0 | | 7 | 0 | 0 |
| Endrin Aldehyde | ng/L | 0 | 0 | 15 | 0 | 0 | 7 | 0 | 0 |
| Endrin Ketone | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Ethion | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Ethoprop | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Famphur | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Fenchlorphos | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Fenitrothion | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Fensulfothion | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Fenthion | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Fonofos | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 4.50 | 15.59 |
| gamma-BHC (Lindane) | ng/L | 0 | | 1 | 0 | | 1 | | |
| HCH, alpha | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 . |
| HCH, beta | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| HCH, delta | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| HCH, gamma | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |
| Heptachlor | ng/L | 0 | 0 | 15 | 0 | | 7 | 0 | 0 - |
| Heptachlor epoxide | ng/L | 0 | 0 | 15 | 0 | | 7 | 0 | 0 |
| Hexachlorobenzene | ng/L | 0 | 0 | 14 | 0 | | | 0 | 0 |
| Leptophos | ng/L | 0 | | 14 | 0 | 0 | | 0 | 0 |
| Malathion | ng/L | 0 | 0 | 14 | 0 | 0 | | 0 | 0 |
| Merphos | ng/L | 0 | 0 | 14 | 0 | 0 | | 0 | 0 |
| Methidathion | ng/L | 0 | | 14 | 0 | | 6 | 0 | 0 - |
| Methoxychlor | ng/L | 0 | | 15 | 0 | | 7 | 0 | 0 |
| Mevinphos | ng/L | 0 | 0 | 14 | 0 | 0 | | 0 | 0 |
| Mirex | ng/L | 0 | | 14 | 0 | 0 | | 0 | 0 |
| Molinate | ng/L | 0 | 0 | 14 | 0 | 0 | | 0 | 0 |
| | _ | 0 | 0 | 14 | 0 | 0 | | 0 | 0 |
| Naled | ng/L | | 0 | 14 | 0 | 0 | | | |
| Nonachlor, cis- | ng/L | 0 | | 14 | 0 | | 6 | 0 | 0 - |
| Nonachlor, trans- | ng/L | | | | | | | 0 | |
| Oxadiazon | ng/L | | | | | 38.13 | | 0 | 0 |
| Oxychlordane | ng/L | 0 | 0 | 14 | 0 | 0 | | 0 | 0 |
| Parathion, Ethyl | ng/L | 0 | | 14 | 0 | 0 | | 0 | 0 |
| Parathion, Methyl | ng/L | 0 | | 14 | 0 | | 6 | 0 | 0 |
| Phorate | ng/L | 0 | 0 | 14 | 0 | | 6 | 0 | 0 |
| Phosmet | ng/L | 0 | | 14 | 0 | | 6 | 0 | 0 |
| Phosphamidon | ng/L | 0 | 0 | 14 | 0 | | 6 | 0 | 0 |
| Prometon | ng/L | | | 0 | 0 | | 6 | | |
| Prometryn | ng/L | | | 0 | 0 | | 6 | | |
| Propazine | ng/L | | | 0 | 0 | | 6 | | |
| Secbumeton | ng/L | | | 0 | 0 | | 6 | | |
| Simazine | ng/L | | | 0 | 0 | | 6 | | |
| Simetryn | ng/L | | | 0 | 0 | | 6 | | |
| Sulfotep | ng/L | 0 | | 14 | | 0 | 6 | 0 | 0 |
| Tedion | ng/L | 0 | 0 | 14 | 0 | 0 | 6 | 0 | 0 |

| | | Sweetwa | iter | Otay | | Tijuan | а |
|-------------------|--------------|---------|------|---------|-----|---------|------|
| Pesticides | Symbol Units | Mean SD | n | Mean SD | n | Mean SD | n |
| Terbufos | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| Terbuthylazine | ng/L | | 0 | 0 | 0 6 | | 0 |
| Terbutryn | ng/L | | 0 | 0 | 0 6 | | 0 |
| Tetrachlorvinphos | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| Thiobencarb | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| Thionazin | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| Tokuthion | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| Toxaphene | ng/L | 0 | 1 | 0 | 1 | | 0 |
| Trichlorfon | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |
| Trichloronate | ng/L | 0 | 0 14 | 0 | 0 6 | 0 | 0 12 |

HYDROLOGIC UNIT REPORTS

Carlsbad Hydrologic Unit Report

ftp://ftp.sccwrp.org/pub/download/DOCUMENTS/TechnicalReports/527 CarlsbadHU Report.pdf

Otay Hydrologic Unit Report

ftp://ftp.sccwrp.org/pub/download/DOCUMENTS/TechnicalReports/527 OtayHU Report.pdf

Peñasquitos Hydrologic Unit Report

<u>ftp://ftp.sccwrp.org/pub/download/DOCUMENTS/TechnicalReports/527_PenasquitosHU_Report.pdf</u>

Pueblo San Diego Hydrologic Unit Report

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San Diego Hydrologic Unit Report

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San Dieguito Hydrologic Unit Report

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San Juan Hydrologic Unit Report

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San Luis Rev Hydrologic Unit Report

<u>ftp://ftp.sccwrp.org/pub/download/DOCUMENTS/TechnicalReports/527_SanLuisReyHU_Report.pdf</u>

Santa Margarita Hydrologic Unit Report

ftp://ftp.sccwrp.org/pub/download/DOCUMENTS/TechnicalReports/527_SantaMargaritaHU_Report.pdf

Sweetwater Hydrologic Unit Report

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Tijuana Hydrologic Unit Report

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