Review of flow duration methods and indicators of flow duration in the scientific literature: Western Mountains







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Southern California Coastal Water Research Project Technical Report 1222

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#### **STATEMENT OF THE PURPOSE**

This review documents methods and indicators to determine the flow status of streams in the Western Mountains (WM), with an emphasis on field-based methods that distinguish ephemeral from perennial and intermittent streams. WM, within the context of this review, is considered to be wetter, higher-elevation portions of the western US that typically receive over 15" rainfall per year (in contrast to the drier Arid Southwest [ASW]; US Army Corps of Engineers 2008; Omernik and Griffith 2014). States within this region include California, Utah, Colorado, Wyoming, Arizona, Montana, New Mexico, and small portions of South Dakota and Nevada. Within many of these states, non-arid regions occur, such as the North Coast of California, or interior mountains (Figure 1).

Although direct measures of flow duration (e.g., long-term records from stream gauges) are usually preferred, indirect measures of flow duration indicators have two major strengths that make them effective tools for watershed managers. First, they are substantially less expensive to measure, typically requiring little more than a single site-visit, whereas stream gauges require substantial installation and maintenance costs. Second, many indirect indicators reflect long-term hydrologic characteristics, integrating over space and time; thus, they provide better information about flow duration than instantaneous observations of aquatic state. They may even be better than direct observations with short periods of records, which may be influenced by short-term changes that do not reflect typical reach conditions.

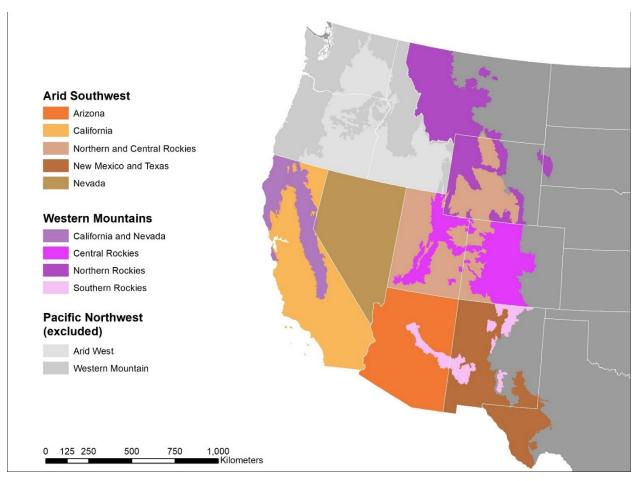


Figure 1. Ecoregions of the Western Mountains.

### **M**ETHODS

#### **General approach**

To identify potential indicators of flow duration in the WM, we first identified a set of indicators used in established flow duration methods (Figure 2). These indicators were characterized by type (e.g., plants, benthic macroinvertebrates) and endpoint used to assess the indicator (e.g., presence of indicator taxa, abundance). We then supplemented this set with additional indicators whose use was supported by scientific literature and other appropriate sources, but not incorporated into established methods. This full list of potential indicators was then evaluated for key criteria:

*Consistency*: Does it work? Is there evidence from appropriate sources (see below) that the indicator can discriminate flow classes across different environmental settings, seasons, etc.? Indicators were consistent if it was used in at least 2 methods or showed support as a discriminatory tool in the scientific literature.

*Repeatability*: Can different practitioners take similar measurements, with sufficient training and standardization? Is the indicator robust to sampling conditions (e.g., time of day)? Repeatability was assessed based on personal knowledge of the field methods.

*Defensibility*: Does the indicator have a rational or mechanistic relationship with flow duration in the WM? This was assessed based on personal knowledge of ephemeral and intermittent stream systems in the region. For example, hydric soils develop in the anoxic conditions created during prolonged inundation and therefore are unlikely to be found in ephemeral streams (Cowardin et al. 1979). In contrast, substrate sorting reflects the magnitude of flow (Hassan et al. 2006), and sorting is evident in ephemeral, as well as perennial and intermittent streams.

*Rapidness*: Can the indicator be measured during a one-day site-visit (even if subsequent lab analyses are required)? Methods requiring multi-day visits or continuous measurement are outside the goals of the present study.

*Objectivity*: Does the indicator rely on objective (often quantitative) measures? Or does it require extensive subjective interpretation by the practitioner?

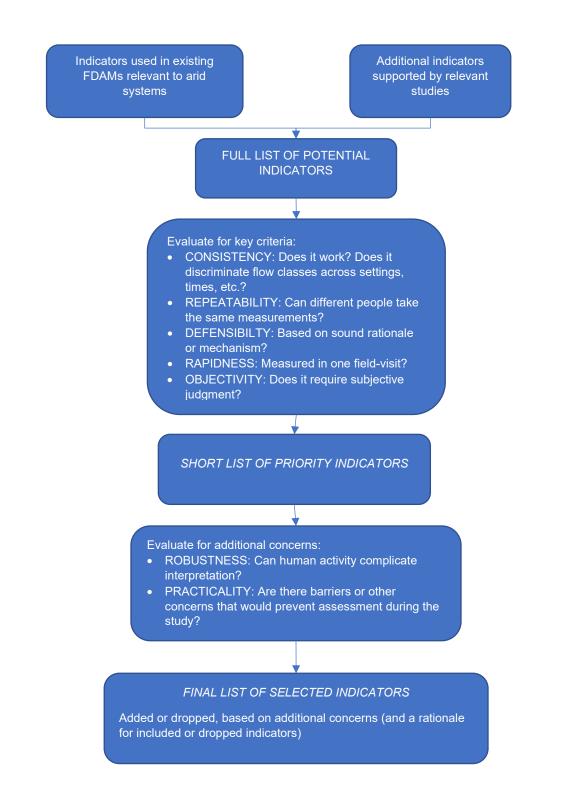
For each indicator, we also noted if there were studies demonstrating efficacy of the indicator in determining flow-duration classes.

The list of potential indicators is shortened to a list of priority indicators for further evaluation if they met most of these criteria. This list is further evaluated for two additional desirable (but not essential) criteria:

*Robustness*: Does human activity complicate interpretation of the indicator in highly disturbed or managed settings? For example, aquatic vegetation may be purposefully eliminated from streams managed as flood control channels, limiting the value of vegetation indicators in certain environments. Although many indicators can be influenced by human activity, they may still provide value in determining flow class (particularly in undisturbed streams). Therefore, this was considered an important, but non-essential, criterion for selecting indicators for exploration.

*Practicality*: Can the technical team realistically sample the indicator in the present study? For example, if special permits are required for assessment, an indicator may be inappropriate for further investigation.

Based on these criteria, a final list of potential indicators of flow duration will serve as the basis for potential field data collection in the WM. The objectives here is to identify indicators that can be combined and evaluated as a FDAM for the WM region. A subsequent objective would be to see how well that preliminary FDAM works compared to Nadeau (2015) and the method developed by the New Mexico Environment Department (NMED 2011).





#### **Search methods**

We built upon previous efforts to conduct a literature review for flow duration indicators in the Arid Southwest (McCune and Mazor 2019) by evaluating sources discovered in that process for relevance to the Western Mountains. We then added to this collection of sources with additional

searches of reference libraries (Table 1). Search terms in Table 1 were used as singular search terms, in combination with western mountains associated regions (e.g., Sierra Nevada mountains), and combined into one "OR" search in Web of Science, Google, and Google Scholar. The first titles or abstracts of the 50 search results were reviewed to determine applicability to the western mountains; relevant results (see next section) were then added to a compiled reference library (<u>https://paperpile.com/shared/3iHwBc</u>), although some sources were later excluded following a more thorough review. This compiled library was supplemented by appropriate sources from the personal libraries of the technical team.

Search Source	Search Date	Key Terms	Hits
Google	11/9/2018	"western mountains"	548,000
Google	11/9/2018	"western mountains" AND "flow duration"	475
Google	11/9/2018	"western mountains" AND "flow duration" AND "indicators"	98
Google Scholar	11/9/2018	"western mountains"	12,000
Google Scholar	11/9/2018	"western mountains" AND "flow duration"	63
Google Scholar	11/9/2018	"western mountains" AND "flow duration" AND "indicator"	43
Google Scholar	11/9/2018	mountains stream indicator "sierra nevada" OR "southern cascade" OR arizona OR "new mexico" OR "black hills" OR montana OR wyoming OR klamath "hydrologic regime"	2,830
Google Scholar	11/9/2018	mountains stream indicator "sierra nevada" OR "southern cascade" OR arizona OR "new mexico" OR "black hills" OR montana OR wyoming OR klamath "flow duration"	1,560
Google Scholar	11/9/2018	"western mountains" AND "hydrologic regime"	115
Google Scholar	11/9/2018	"western mountains AND "intermittent stream"	77
Google Scholar	11/9/2018	"western mountains" AND "perennial stream"	80
Google Scholar	11/9/2018	"western mountains" AND "ephemeral stream"	44
Web of Science	11/9/2018	"western mountains" AND ("streamflow duration" OR "flow assessment" OR "intermittent" OR "ephemeral" OR "biological Indicators" OR "clean water act jurisdiction" OR "hydrologic regime" OR "merovoltine" OR "semivoltine" OR "univoltine")	410,510

# Table 1. Search parameters and dates used to assemble literature on indicators of flow duration in western mountains.

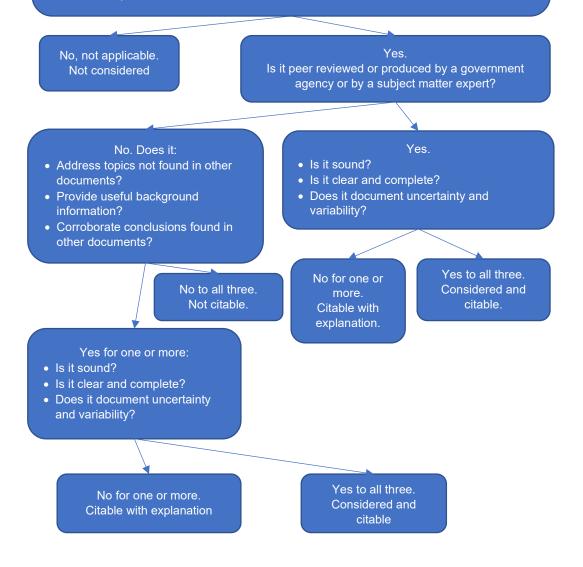
#### Analysis of sources

Including sites in the review

Sources with available articles were reviewed and annotated to assess the applicability, soundness, clarity, and uncertainty. Annotations were focused on synthesizing the scientific merit of each goal, procedure, result and conclusion reported by the authors, after relevance to the ASW or flow duration classes was evaluated. Sources were reviewed for five elements to warrant inclusion in the library, following the decision tree in Figure 3.

#### Is the source applicable?

- Does it pertain to flow duration of streams?
- Does it include information about biological, physical, or hydrologic characteristics of streams associated with flow duration that can be rapidly measured in the field?
- Does it provide information about streams in the WM, or mountainous regions elsewhere OR does it provide other relevant information about flow duration?



#### Figure 3. Decision tree for reviewing sources.

*Applicability/Utility:* Sources that provide information about the biological, physical, or hydrologic characteristics of streams along a flow duration gradient in the WM were considered applicable. Sources in regions outside the WM may also be considered applicable if other elements of the reference were relevant to the study. Several sources found during searches did not meet this criterion. Factors that limited the applicability of a citation include reliance on intensive hydrologic data (e.g., continuous flow gauge data), or reliance on other data types that could not be rapidly measured in the field (e.g., remote sensing data).

*Review:* Sources needed to undergo peer-review, be published by a government agency, or come from a subject-matter expert. All sources met this criterion.

*Soundness:* Sources needed to rely on sound scientific principles, and conclusions had to be consistent with data presented. All sources met this criterion.

*Clarity/Completeness:* Sources needed to provide underlying data, assumptions, or model parameters, as well as author sponsorship or author affiliations. Several sources did not provide a clear basis for determining flow-duration classes for study sites. Where possible, we applied the most appropriate flow-duration class based on available data, sometimes applying ambiguous classifications (e.g., "perennial or intermittent", or "intermittent or ephemeral"). If data were insufficient to support these designations, the source was excluded from the review.

*Uncertainty/Variability:* Sources needed to identify variability, uncertainties, sources of error, or bias, reflecting them in any conclusions drawn. We looked for reported ranges or measures of variability and uncertainty (e.g., standard deviation, statistical significance) associated with each indicator and flow-duration class. No sources were excluded for this criterion.

#### Evaluating information about indicators

Each source was reviewed to identify information about indicators of flow duration. First, the classes represented in the study were determined. Classes were either reported by the authors, or determined from other data presented in the study. For example, sites were classified as perennial if year-round flow was reported. Where appropriate, ambiguous classes were applied; for example, if a study reported that a stream dried, but the duration of the dry period was unclear, the site was classified as "ephemeral or intermittent." Results, including manuscript text, figures, and tables, were reviewed for information about indicators associated with different site classes. Typical levels (e.g., means) and associated measures of variability (e.g., ranges, standard deviations) were recorded for each indicator.

### RESULTS

#### Literature review

All literature (including PDF copies, where available) are included in this endnote library:

https://ftp.sccwrp.org/pub/download/TMP/RaphaelMazor/USEPA WM Bibliography.zip

#### Flow duration assessment methods

Seven methods were appropriate for evaluating flow-duration classes in the WM (Table 2). An additional three methods were found (Kennard et al. 2010; Trubilowicz et al. 2013; and Berhanu et al. 2015), but were excluded because they lacked a rapid field component, focusing instead on long-term records of measured or modeled flow. Table 3 provides a summary of which indicators were used with which method.

Table 4 provides a summary of the evaluation criteria for each indicator. Indicators that met all criteria were designated as priority indicators. With some exceptions, all priority indicators were

proposed for inclusion in the pilot study in the WM; rationale for excluding priority indicators, or for including non-priority indicators, is provided in the table.

Table 2. Methods for assessing flow duration and their associated indicators. Asterisks indicate that the protocol includes portions of the WM.

		Indicators							
Source	Geographic location	Represented classes	Biological	Geomorphological	Other				
Nadeau (2015a)	Pacific Northwest, USA*	Ephemeral, perennial and intermittent	Benthic macroinvertebrate, wetland plants, riparian corridor, fish, amphibians/snakes	Slope, evidence of erosion/deposition, floodplain connectivity					
Topping et al. (2009)	Oregon, USA*	Ephemeral, perennial and Intermittent	Wetland plants, fibrous roots and rooted plants, streamer mosses or algal mats, iron-oxidizing bacteria, fungi, flocculent material, benthic macroinvertebrates, amphibians/snakes, fish, lichen line, riparian vegetation corridor	Continuous bed and bank, in-channel structure, soil texture or stream substrate sorting, erosional features, depositional features, sinuosity, headcuts and grade controls, groundwater/hyporheic saturation, springs and seeps					
Fritz et al. (2006)	Temperate USA (Indiana, Kentucky, Ohio, Illinois, New Hampshire, New York, Vermont, West Virginia, and Washington)*	Ephemeral, perennial and intermittent	Benthic macroinvertebrates, amphibians, algal cover, algal assemblage, bryophyte assemblage, riparian canopy cover	Sinuosity, slope, depth, wetted width, depth to bedrock/groundwater table, streambed sediment moisture/size distribution	water chemistry, habitat unit designation, water velocity, continuous hydrologic monitoring				
NC Division of Water Quality (2010)	North Carolina	Intermittent and perennial	Iron oxidizing bacteria, leaf litter, organic debris drift accumulation, fibrous roots, rooted upland plants, benthic macroinvertebrates, aquatic mollusks, fish, crayfish, amphibians, algae, wetland plants in streambed	Presence of modification/ditches, channel and bank continuity, sinuosity, channel structure, streambed particle size, active/relict floodplain, depositional bars/benches, recent alluvial deposits, headcuts, grade control (natural), natural valley, 2nd or > order channel,	Baseflow presence, sediment on plants/debris, soil chroma				
Surface Water Quality Bureau, NM Environme nt Departme nt (2011)	New Mexico, USA*	Ephemeral, perennial and intermittent	Fish, benthic macroinvertebrates, filamentous algae and periphyton, riparian vegetation, rooted upland plants in streambed, iron oxidizing bacteria/fungi, bivalves, amphibians	Sinuosity, floodplain and channel dimensions, channel structure, particle size or stream substrate sorting, seeps/springs	Water in channel, hydric soils, sediment on plants or debris, hyporheic zone/groundw ater table				
Gallart et al. (2017)	Mediterranean Europe	Intermittent-pools, intermittent-dry,			Hydrologic metrics (based				

		episodic-ephemeral, perennial; Hyperrheic, eurheic, oligorheic, arheic, hyporheic/dry			on modeled or recorded flow), citizen observations
Svec et al. (2005)	Eastern Kentucky	Ephemeral, intermittent, perennial		Bankfull width & depth, entrenchment ratio, slope, watershed area, estimated flood plain area	
Ohio EPA (2012)	Ohio	Ephemeral, intermittent/perennial (warm water), perennial (cold water)	Fish, benthic macroinvertebrates, amphibians, exposed plant roots on banks, riparian/in channel vegetation, organic matter	Bankfull width & depth, bed & bank presence, erosional/depositional channel features, sinuosity, estimated channel gradient, substrate sorting, groundwater presence, leaf litter, seeps & springs	Water quality measurements , flowing water in channel
Straka et al. (2019)	Czech Republic	Intermittent, near- perennial, and perennial	Benthic macroinvertebrates		
McCleary et al. (2012)	Alberta, Canada	Upland, ephemeral or water source areas, intermittent, intermittent/transition al/small permanent, small permanent or large permanent	In channel vegetation presence	Bankfull width, undercut width, substrate sorting, erosional/depositional channel features, bed and bank presence, moisture regime	Water in channel
Savage and Rabe (1979)	Idaho	Ephemeral, "spring streams" and perennial		Gradient, substrate	

Indicator	Temperate USA	Oregon	North Carolina	New Mexico (Phase 1)	New Mexico (Phase 2)	Pacific Northwest	Mediterranean	Kentucky	Ohio	Czech Republic
Geomorphology										
Bankfull width and depth	Х							Х	Х	
Continuous bed and banks presence		Х	Х						v	
Depositional or erosional features in the channel		Х	Х						Х	
Depositional or erosional features on the floodplain Distinct substrate composition in streambed from adjacent uplands	х	х	x x	х						
Entrenchment ratio	Х			Х				Х		
Evidence of active floodplain										
Evidence of relict floodplain			Х							
Natural valley presence			Х							
Presence of headcuts In-channel sequences of erosional and depositional features	x x	x x	x x	х						
Stream order	~	~	X	~						
Sinuosity	х	х	X	Х					х	
Slope	X	~	~	~		Х		Х	Х	
Hydrology	~					~				
Continuous logged data	х				Х					
Groundwater observation	X	х	х	Х					х	
Distribution of leaf litter or debris	X	Λ	X	χ					х	
Hydric soils or redoximorphic features	Λ	Х	Λ	Х						
Modeled hydrology		Λ		χ			Х			
Observed aquatic state	х			Х			Х		Х	
Reported aquatic state from interviews							Х			
Observed or reported soil saturation	х	х		Х			Х			
Observation of baseflow	Λ	~	х	~	Х		X			
Presence of wrack or drift lines		х	X				~			
Sediment deposition on plants or debris		Χ	X	Х						
Soil-based evidence of a high water-table			x	~						
Presence of seeps and springs		Х	~	Х					х	
Velocity	х	Λ		Λ						
Biology	~									
Iron-oxidizing bacteria or fungi		х	х	Х						
Algae	х	X	X	x	Х					
Lichens	~	X	~	~						

#### Table 3. Summary of indicators included in flow-duration assessment methods.

Bryophytes	Х	Х						
Wetland vegetation		Х	Х			Х	Х	
Upland vegetation in channel		Х	Х	Х			Х	
Riparian vegetation		Х		Х		Х	Х	
Aquatic macroinvertebrates - Presence	х	Х		Х		Х	Х	
Aquatic macroinvertebrates - Abundance	Х		Х		Х	Х	Х	
Aquatic macroinvertebrates - Indicator taxa		Х	Х		Х	Х	Х	
Aquatic macroinvertebrates – Traits								
Amphibians – Presence	х	Х			Х	Х	Х	
Amphibians - Abundance and diversity	Х		х				Х	
Reptiles – Presence		Х				Х		
Fish – Abundance			х		Х		Х	
Fish – Presence		Х		Х		Х		

# Table 4. Evaluation criteria for indicators identified in the literature review. 1: Non-priority indicator proposed for inclusion because it is required by the New Mexico protocol (NMED 2011).

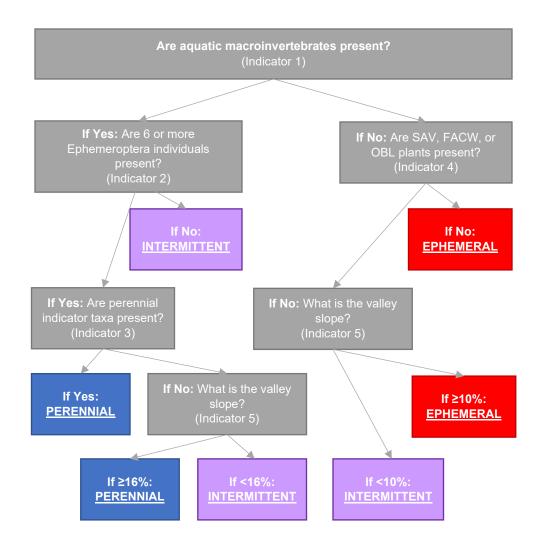
Indicator	Consistency	Repeatability	Defensibility	Rapidness	Objectivity	Priority Indicator	Robustness	Practicality	Proposed
Geomorphology									
Bankfull width and depth	Х	х		х	Х	No	Х	Х	No
Continuous bed and banks presence	Х	Х		Х		No	х	Х	No
Depositional or erosional features in the channel	Х	х		х		No		х	No
Depositional or erosional features on the floodplain	х	х		х		No		х	No
Distinct substrate composition in streambed from adjacent uplands	х	х		х		No	х	х	Yes
Entrenchment ratio	Х	Х		Х	Х	No		Х	Yes
Evidence of active floodplain		Х		Х		No	Х	Х	No
Evidence of relict floodplain		Х		Х		No	Х	Х	No
Natural valley presence		Х		Х		No		Х	No
Presence of headcuts	Х	Х		Х	Х	No	Х	Х	Yes <sup>1</sup>
In-channel sequences of erosional and depositional features	х	х		х		No	х	х	Yes <sup>1</sup>
Stream order		Х		Х	Х	No		Х	No
Sinuosity	Х	Х		Х	Х	No	Х	Х	Yes <sup>1</sup>
Slope	Х	Х	Х	Х	Х	Yes	Х	Х	Yes

Hydrology									
Continuous logged data	Х	Х	Х		Х	No	Х		No
Groundwater observation	Х	Х	Х		Х	No	Х		No
Distribution of leaf litter or debris	Х	Х		Х		No		Х	No
Hydric soils or redoximorphic features	Х	Х	Х	Х	Х	Yes	Х	Х	Yes
Modeled hydrology	Х	Х	Х		Х	No	Х		No
Observed aquatic state	Х	Х	Х	Х	Х	Yes		Х	Yes
Reported aquatic state from interviews		Х	Х		Х	No	Х		No
Observed or reported soil saturation		Х	Х	Х	Х	No		Х	No
Observation of baseflow	Х	Х	Х	Х		No	Х		No
Presence of wrack or drift lines	Х	Х		Х		No		Х	No
Sediment deposition on plants or debris	Х	Х		Х	Х	Yes	Х	Х	Yes <sup>1</sup>
Soil-based evidence of a high water table	Х	Х	Х	Х		No	Х	Х	No
Presence of seeps and springs	Х	Х	Х	Х	Х	Yes	Х	Х	Yes
Velocity		Х		Х	Х	No	Х	Х	No
Biology									
Iron-oxidizing bacteria or fungi	Х	Х	Х	Х	Х	Yes	Х	Х	Yes
Algae	Х	Х	Х	Х	х	Yes		Х	Yes
Lichens		Х	Х	Х	х	No		Х	No
Bryophytes	Х	Х	Х	Х	Х	Yes		Х	Yes
Wetland vegetation (FACW, OBL, SAV)	Х	Х	Х	Х	Х	Yes		Х	Yes
Upland vegetation in channel	Х	Х	Х	Х	Х	Yes		Х	Yes
Riparian vegetation	Х	Х	Х	Х	х	Yes		Х	Yes
Aquatic macroinvertebrates - Presence	Х	Х	Х	Х	х	Yes	х	Х	Yes
Aquatic macroinvertebrates - Abundance	х	Х	Х	Х	х	Yes	х	Х	Yes
Aquatic macroinvertebrates - Indicator taxa	Х	Х	Х	Х	Х	Yes		Х	Yes
Amphibians - Presence	Х	Х	Х	Х	х	Yes		Х	Yes
Amphibians - Abundance and diversity	х	Х	Х		х	No			No
Reptiles - Presence	х	Х	Х	Х	х	Yes		Х	Yes
Fish - Abundance	х	х	Х		х	No			No
Fish - Presence	х	х	х	х	х	Yes		Х	Yes
Amphibians - Abundance and diversity Reptiles - Presence Fish - Abundance Fish - Presence	X X X	X X X	X X X		х	X X X X	X No X X Yes X No	X No X X Yes X No	X No X X Yes X X No
Additional indicators from primary literature Geomorphology								-	
Max pool depth		Х		Х	Х	No	Х	Х	No
Hydrology									
Dissolved O <sub>2</sub>		х		х	х	No		Х	No
Woody jams		х	х	х	х	No	х	Х	No

#### **Pacific Northwest**

For purposes of classifying perennial, intermittent, and ephemeral streams in the Pacific Northwest (including the portion of that region within the WM), Nadeau (2015) developed a method that uses five biological and physical habitat indicators: 1) presence of aquatic macroinvertebrates; 2) number of mayflies (order Ephemeroptera); 3) presence of perennial indicator taxa from Mazzacano and Black (2008) or Blackburn (2012); 4) presence of wetland indicator plants (specifically, SAV, FACW, or OBL) from the US Army Corps of Engineers (2013); and 5) valley slope. Additional indicators, such as the presence of fish, aquatic stages of amphibians, and evidence of sediment erosion or deposition, are also considered. These five indicators will serve as the foundation for evaluation of flow duration assessment methods in the WM. Indicators are measured in an objective fashion, without requiring subjective or qualitative visual assessments by practitioners.

Indicators are evaluated with a simple branching flow-chart (Figure 4), and not all indicators are needed to make a determination at every site. Consequently, it is among the simplest tools to implement. This method strongly emphasizes biological indicators, including only one geomorphological indicator (i.e., slope), and no hydrological indicators.



# Figure 4. Flowchart used to determine flow class in the Pacific Northwest method (adapted from Nadeau 2015).

#### Oregon Interim Method

Prior to the development of the method of Nadeau (2015) for the PNW, Topping et al. (2009) developed a flow duration assessment tool for Oregon that evaluates a series of geomorphological, hydrological, and biological indicators as absent, weak, moderate, or strong at the reach. This method was developed for a region that includes the WM. In general, the strength of the indicator is considered evidence of longer flow durations. Each indicator is scored and summed; if the total score is below 13, the stream is considered ephemeral, and if it's above 25, it is considered perennial. Single indicators (e.g., presence of fish, amphibians, or aquatic macroinvertebrates) may trump an "ephemeral" score. In contrast to Nadeau (2015), assessing the strength of the indicators requires subjective visual assessments by practitioners.

#### New Mexico

The New Mexico Environment Department developed a two-phase method for assessing flow duration (NM Environment Department 2011) in both arid and mountainous regions of that state.

The first phase is more rapid, and it is sometimes sufficient to classify a stream as perennial, intermittent, or ephemeral. This first phase relies on qualitative sampling of benthic macroinvertebrates, fish, filamentous algae, and other organisms, plus field observation of channel morphology and soils. In some cases, a second phase consisting of quantitative fish and benthic macroinvertebrate samples may be necessary. This second phase also requires the use of continuous loggers or stream gauges to measure water presence. In this method, 14 indicators of flow duration ("attributes") are scored, yielding a quantitative index that forms the basis of the classification (Table 5). Notably, this method may result in ambiguous situations (gray rows in Table 5), which may be resolved by more intensive "level 2" analysis, and by investigation of adjacent reaches. Certain indicators (specifically, fish and aquatic macroinvertebrates) may result in a perennial designation, even if scores are low. Like Nadeau (2015), this method was designed for application in arid regions. Like Topping et al. (2009), many indicators require subjective visual assessment by practitioners.

Waterbody type	Level 1 total score	Determination
Ephemeral	Less than 9.0	Stream is ephemeral
	≥ 9.0 and < 12.0	Stream is recognized as intermittent until further analysis indicates that the stream is ephemeral.
Intermittent	≥ 12 and ≤ 19.0 or score is lower but aquatic macroinvertebrates and/or fish are present	Stream is intermittent
	> 19.0 and ≤ 22.0	Stream is recognized as perennial until further analysis indicates that the stream is intermittent
Perennial	Greater than 22.0	Stream is perennial

#### Table 5. Score interpretation for the New Mexico method.

#### Mediterranean Europe

Prat et al. (2014) developed an assessment framework known as Mediterranean Intermittent River ManAGEment (MIRAGE) to identify the flow status of streams in order to guide selection of appropriate condition assessment tools based on biology, water chemistry, habitat, or other condition indicators. The first step in analysis is determining the flow duration of a stream using the Temporary Stream Regime Tool (TRS-Tool; Gallart et al. 2012; Gallart et al. 2017). The TRS-Tool uses three potential sources of flow estimation/observation to determine stream flow classification: 1) interviews, 2) interpretation of high-resolution aerial photographs and rapid field observation, and 3) outputs from hydrologic rainfall-runoff models.

In contrast with other methods, this assessment method classifies streams into four classes, reflecting the predominant aquatic states: intermittent-pool, intermittent-dry, episodic-ephemeral, or perennial.

Methodology for interviews is documented in Gallart et al. (2016). Interviews target people encountered in the vicinity of a stream in question, as well as with regional experts with a "professional or leisure" relationship with the river. The core interview consists of five key questions:

1. How often does flow cease?

- 2. During non-flowing months, are there pools and for how long?
- 3. When there is no surface water, is there water in the alluvium?
- 4. How frequently are flow/pools/dry riverbeds observed during each season?
- 5. Have any changes in flow regime been observed recently?

Rapid field observations and photographic interpretation focuses strictly on hydrologic indicators, such as presence of pools, riffles, or dry streambed.

Hydrologic rainfall-runoff models are interpreted by calculating a set of flow metrics that are associated with different aquatic states (Gallart et al. 2012). These metrics relate to flow permanence (Mf), pool permanence (Mp), and dry-period permanence (Md; Figure 5). Similar models have been developed for use in the WM region (e.g., runoff models developed for the Sawtooth Mtns., Arp et al. 2006), but require intensive investments for model development and are thus not considered an appropriate methodology in this review.

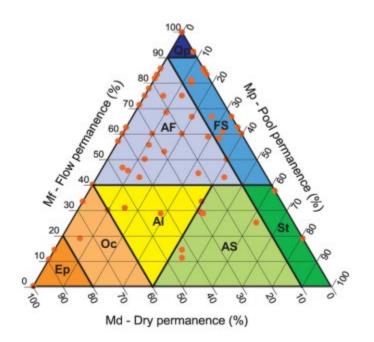


Figure 5. Plot used for classifying flow duration based on three metrics calculated from hydrologic model outputs.

#### North Carolina

This method, developed by the North Carolina Division of Water Quality (2010), includes 9 biological, 11 geomorphic, and 6 hydrologic indicators to classify a stream as perennial, intermittent, or ephemeral, as well as to designate locations in the landscape as origins of streamflow or sinks where flow ceases. As with the New Mexico method, indicators are individually scored and then summed to yield an overall index score. Some indicators (or more robustly evident indicators) are weighted higher than others and the presence of specific taxa (fish, crayfish, amphibians, or clams) can result in a perennial designation, even if the overall index score is low. Scores required for perennial or intermittent designations are somewhat

higher for the North Carolina method than the New Mexico method, perhaps due to the higher number of indicators (26 vs. 14). This method was developed for a region that receives considerably more rainfall than the WM.

#### Temperate US (IN, KY, OH, IL, NH, NY, VT, WV, and WA)

Fritz et al. (2006) described a comprehensive suite of protocols for measuring potential flow permanence indicators. The suite of methods described is more comprehensive than the other listed methods, but no conclusive flow duration classification is drawn upon at the end of the methods. Publications following this report (Fritz et al. 2009; Johnson et al. 2009; Fritz et al. 2009; Roy et al. 2009; Datry et al. 2014) assess the effectiveness of each indicator separately. These methods have been applied widely throughout the USA, mostly for forested headwater streams outside the WM (the exception being Washington state).

#### Czech Republic

Straka et al. (2019) recently developed a "Biodrought" index to classify streams as perennial, or intermittent based strictly on the composition of benthic macroinvertebrate communities (Figure 6). Based on a data set of 23 streams in the Czech Republic (mostly in the Carpathian Mountains and Central Highlands) consisting mostly of paired perennial and non-perennial sites (both "intermittent" and "near perennial"), they identified indicator species associated with different flow regimes, and developed a seasonally-adjusted index consisting of three metrics that could discriminate between the three flow-regime classes (Table 6).

#### Table 6. Metrics in the Biodrought index developed by Straka et al. (2019).

Metric	Flow state indicated by high values
Proportion of indicator taxa (perennial indicators/ perennial + intermittent indicators)	Perennial
Proportion of taxa with high body flexibility	Intermittent
Preference for organic sustarte (Autumn samples only)	Intermittent
Total abundance (Spring samples only)	Perennial

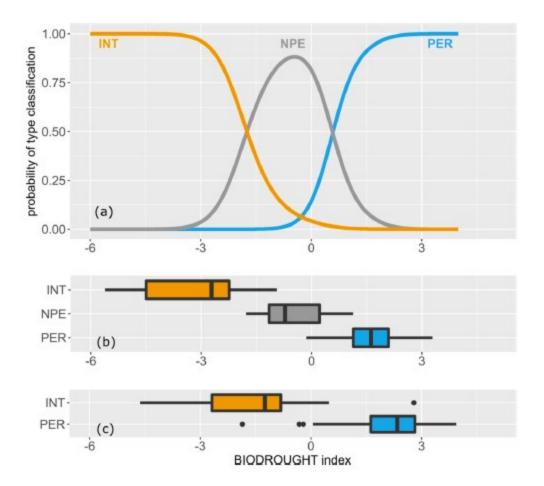


Figure 6. Relationship between Biodrought index scores and flow classes, from Straka et al. (2019). Top panel shows the probability of classification as the index score increases. The second panel shows scores associated with calibration data. The bottom panel shows scores associated with independent validation data. INT: Intermittent. NPE: Near-perennial. PER: Perennial.

As with Nadeau (2015), the index of Straka et al. (2019) uses aquatic invertebrates to discriminate between perennial and intermittent streams, but not to discriminate ephemeral streams. But the two indices differ in a few important aspects. First, indicator taxa for Straka et al. (2019) were identified at the species or genus level, which reduces the rapidness of this method if lab-based identifications are required. Second, indicator taxa were identified through an empirical method (i.e., indicator species analysis), whereas the indicators of Nadeau (2015) were derived from life history information and experience of stream ecologists in the Pacific Northwest (Blackburn 2012). Third, the Biodrought index takes into account the presence of intermittent indicator taxa, whereas the method of Nadeau (2015) found superior performance when only perennial indicator taxa are considered. This index has not been validated.

#### INDICATORS IN THE WESTERN MOUNTAINS

A review of literature describing indicators in the WM shows general support for indicators used in current flow duration assessment methods, particularly biological indicators. A number of potential new indicators not used in any of the methods in Table 2 are also discussed, and summarized in Table 7. We discuss each class of indicators and determine whether specific indicators should be included in the evaluation of flow duration assessment methods in the WM, with particular attention to the indicators included in Nadeau (2015). Relevant information from personal experience or communication with regional experts is included as well.

#### **Geomorphological Indicators**

Geomorphic indicators in the WM are defined primarily by Mersel and Lichvar (2014), while many of the specific relationships between hydrology and channel geometry are initially defined for streams in the Western US by Hedman and Osterkamp (1982). Changes in geomorphological indicators over time can be indicative of a change to ephemerality from a perennial state (e.g., bed and bank destabilization), but one-time observations of indicators may be more related to storm intensity, stream power, and substrate composition (Friedman and Lee 2002). Outside of the temporary stream indicators defined by Mersel and Lichvar (2014), there were no studies found that defined differences in stream geomorphology based on flow duration classification; rather, there were studies that characterized geomorphology for WM streams based on channel types – e.g., cascade, step pool, plane bed and pool riffle steam morphologies typified for the region by Wohl and Merritt (2008).

#### Tufa deposits

In alkaline waters rich in carbonate, tufa deposits may form under certain conditions. Tufa deposition processes are highly dependent on physiochemical and biological factors not directly related to flow duration (Ford and Pedley 1996). For example, Ford and Pedley (1996) described areas throughout the US (including sites in the WM) in which tufa formations occur, including fossil tufa sites, where historical conditions allowed for the formation of tufa but are no longer actively forming – tufa presence is not representative of present-day hydrologic conditions. No studies were found to support the use of tufa deposits as an indicator of flow duration, as the presence of such formations is not an indicator of present-day streamflow. Observations of tufa formations in an ephemeral stream by Wright (2000) showed that minimal flow is needed for such formations, whereas flow obstructions can be the major factor affecting tufa formation in ephemeral streams. Other than Wright (2000), there were no other studies found that focused on describing connections between flow duration and tufa formation; rather, most research found aimed at understanding the physiochemical or biological processes that affect tufa formations.

#### Hydrologic Indicators

Several studies supported the use of certain hydrologic indicators of flow duration, particularly direct observation of flow, groundwater, and soil conditions (Turner and Richter 2011; Gallart et al. 2016). In dry channels, several methods distinguish intermittent from ephemeral streams by evaluating the distribution of leaf litter or looking for the presence of wrack lines (e.g., Topping et al. 2009; NMED 2011). As with many geomorphological indicators, the distribution of organic material in the stream channel is more of an indication of stream power than flow duration, and thus has limited utility outside of low-power headwater systems. An important exception is the development of hydric soils, which are produced by the anoxic conditions associated with prolonged inundation (Cowardin et al. 1979). A couple studies showed that water chemistry is distinct in intermittent and perennial streams (e.g., Bonada et al. 2006; Bogan 2017). Notably, solute concentrations tend to be higher in intermittent streams, particularly towards the end of the

drying period. However, values overlapped considerably among flow duration classes, suggesting that this would not be a consistent indicator for flow duration assessment.

#### Woody jams

Mersel and Lichvar (2014) identified large woody jams (also called "debris jams") as an important component of streams in the WM, and several studies in our review investigated the impacts of large woody jams on stream ecology, stream channel morphology, water velocity, and to a lesser extent, flow duration. There were conflicting reports of effect (Gippel 1995; Mason Jr. et al. 1990; Faustini & Jones 2003; Shields & Gippel 1995) versus no effect (Matheson et al. 2017; Lester & Wright 2009) on the influence of organic jams (flow obstructing large woody debris) on flow duration in our review, but several studies did consistently support the direct influence of jams on modifications of other stream flow duration indicators - e.g., Abbe and Montgomery (1996), Faustini and Jones (2003), and Smith et al. (1993) showed significant differences in gradient, bank morphology and pool frequency along stream reaches pre- and postjam removal. In a review of hydrologic effects of large woody jams, Gippel (1995) shows that their presence can have an indirect slowing effect on flow conveyance via increases in channel roughness and increases in channel stage height. Several studies have documented the prevalence large woody jams in WM streams (Mersel and Lichvar 2014), and although there is little evidence to support direct hydrologic influence of debris jams, the other processes they affect support its evaluation as a potential indicator of flow duration.

#### **Biological Indicators**

In contrast to the many of the other indicators mentioned above, biological indicators are often directly related to flow duration. Consequently, many studies corroborated relationships between these indicators and flow duration, particularly aquatic macroinvertebrates and plants.

#### Aquatic macroinvertebrates

Several methods presume that ephemeral streams are unable to support aquatic macroinvertebrates, with the exception of short-lived taxa like Culicidae. Although there are numerous studies of aquatic invertebrates in the WM, no studies we are aware of examine ephemeral streams in this region; studies that explore the role of streamflow duration typically located perennial sites in the WM, and intermittent or ephemeral sites in adjacent portions of the ASW (e.g., Bogan et al. 2013). De Jong et al. (2015) sampled ephemeral streams in arid (i.e., non-mountainous) portions of New Mexico, collecting 86 different taxa of aquatic macroinvertebrates within a few days of the onset of flow. Many of these taxa had aerially dispersing adult life-stages, and most were found only in ephemeral reaches that were connected to perennial reaches. Newly hatched larvae of the mayfly *Callibaetis* were found within 1 day of the onset of flow, as were adults of taxa that can aerially disperse (typically beetles). Tadpole shrimp (*Triops*) were also frequently observed, as partially terrestrial taxa (e.g., annelids). However, these observations are likely to have only a small impact on the ability to use the presence of aquatic invertebrates as an indicator of intermittent or ephemeral flow. First, assessments should be timed to avoid the first few days of the onset of flow, after which ephemeral streams will no longer support aquatic macroinvertebrates. Second, additional

exclusions can be made, such as ignoring early instars (which are unlikely to be detected in rapid field methods), aerially dispersing adults, and partially terrestrial fauna.

In general, studies provide strong support for the use of aquatic invertebrates as indicators of flow duration. Although training is required, field-based family level identifications are practical for aquatic macroinvertebrates, further underscoring their suitability as indicators. While many studies demonstrate consistent compositional differences between perennial and intermittent streams (e.g., Bramblett and Fausch 1991; Rader and Belish 1999), only some presented data in a way to ascertain the value of specific taxa to indicate flow status. Typically, results are presented at species or genus level, when field indicators may require identifications at family level or lower.

Nadeau (2015) makes use of studies by the Xerces society (i.e., Mazzacano and Black 2008; Blackburn 2012) to identify perennial indicator taxa in the PNW, and it is likely that most of these taxa have similar indicator value in the rest of the WM. However, a few studies show that some taxa exhibit different habitat affinities in the WM. Many studies noted that intermittent reaches adjacent to perennial waters may support perennial indicator taxa, suggesting that this indicator may reflect hydrologic patterns at larger spatial scales than needed of flow-duration assessment methods.

In a study of drought-impacted streams of the Sierra Nevada, California, Herbst et al. (2019) noted that intermittency was a large driver of change in invertebrate communities. Sites that experienced intermittency were associated with midges, as well as the alderfly *Sialis*, whereas streams that experienced continuous flow throughout the drought had higher densities of mayflies, stoneflies, and caddisflies, including several perennial indicators identified by Mazzacano and Black (2008): Rhyacophilidae, Hydropsychidae, and Perlidae. They also noted that filter feeders had lower densities at intermittent streams, compared to continuously-flowing streams.

Straka et al. (2019) identified numerous taxa indicative of either perennial or intermittent flows in their study of Czech streams, and their list of taxa diverges from those of Blackburn (2012) in several aspects, at least partly due to the different taxonomic resolution of the two studies.

#### Mollusks

In general, support was strong for the perennial indicator status of mollusks (e.g., Lusardi et al. 2016), particularly for the New Zealand mudsnail (*Potamopyrgus antipodarum*), a non-native invader in streams throughout the West (e.g., Herbst et al. 2008; Bogan et al. 2013). However, Straka et al. (2019) identified this taxon as an indicator of intermittent or nearly perennial Czech streams, along with numerous taxa in Physidae, Planorbiidae, and Lymnaeidae. A number of Lymnaeid taxa were also indicators of perennial flow, along with the Ancylid snail *Ancylus fluviatilis*. Although they are less widespread than many gastropods, freshwater mussels are also likely to be good indicators of perennial flow, some species have been observed in perennial pools within intermittent streams (e.g., Clark 2010). Fingernail clams (Sphaeriidae) are not treated as a perennial indicator taxon, but some support for this classification is found in Lusardi et al. (2016). However, Straka et al (2019) identified *Pisidium* as an indicator of intermittent flow.

#### Mayflies

No mayfly families are considered to be an indicator of perennial flow in Blackburn (2012), although some studies suggest that some taxa show a preference for perennial flow (e.g., Isonychidae, King et al. 2015). Some studies support Baetidae as a perennial indicator (e.g., Bonada et al. 2006; Bramblett and Fausch 1991), while others suggest they prefer intermittent flow (e.g., Miller and Brasher 2011). Straka et al. (2019) found numerous mayfly indicator taxa of both intermittent/nearly perennial streams (e.g., *Cloeon dipterm*) and perennial streams (e.g., *Baetis rhodani*).

#### Stoneflies

Several studies supported the use of perlid stoneflies as indicators of perennial flow (e.g., Bonada et al. 2006; Lusardi et al. 2016; Bogan 2017), but a few studies report them at very low abundance in intermittent streams (e.g., del Rosario and Resh 2000). Few studies indicated if Pteronarcyidae were collected, suggesting that this taxon may be too rare to be a useful indicator in the WM.

Although Capniidae are listed as an indicator of intermittent flow in Blackburn (2012), and this family is known to contain intermittent stream specialist taxa (e.g., *Mesocapnia arizonensis*, Bogan 2017), intermittent indicators are not used in Nadeau (2015), and many taxa in this family are found in perennial streams as well as intermittent (Bogan 2017).

In Czech streams, Straka et al. (2019) identified four indicators of intermittent flows (in Taeniopterygidae, Capniidae, Perlodidae, and Nemouridae), and numerous indicators of perennial flow (species in Nemouridae, Perlidae, Perlodidae, Chloroperlidae, and Leuctridae). One *Isoperla* species (i.e., *I. tripartita*) was an indicator of intermittent flows, whereas two species (i.e., *I. oxlepis* and *I. rivularum*) were indicators of perennial flows, suggesting that even genus-level identifications may be too coarse to provide meaningful indication of flow duration.

#### Caddisflies

Several studies support the use of Hydropsychidae, and to a lesser extent, the other three families (i.e., Philopotamidae, Rhyacophilidae, and Glossosomatidae) as indicators of perennial flow (Bonada et al. 2006; Miller and Brasher 2011; Erman and Erman 1995). Several studies suggested that additional families, such as Brachycentridae or Calamoceratidae, may also be a good indicator of perennial flow in parts of the WM (Bonada et al. 2006; Miller and Brasher 2011). Staka et al. (2019) identified a handful of indicator species for intermittent flows in Czech streams (Beraeidae, Phryganeidae, and numerous species in Limnephilidae), and numerous indicators of perennial flows in several families (including Glossosomatidae, Hydropsychidae, Limnephilidae, Phryganeidae, Polycentropidae, and Rhyacophilidae).

#### Beetles

Several studies showed that elmid beetles showed a strong preference for perennial streams, but that they are occasionally found in intermittent reaches as well, particularly if they are close to perennial waterbodies. De Jong et al. (2013) note that *Optioservus quadrimaculatus* and *Zaitzevia parvula* are comparatively well-adapted to colonize intermittent streams shortly after rewetting in the ASW. Psephenidae were supported as an indicator of perennial flow in Bonada

et al. (2006) and King et al. (2015). Several aquatic beetle families could be indicators of intermittent (e.g., Hydrophilidae: Bonada et al. 2006; Bogan and Lytle 2007), and some are documented ephemeral streams (De Jong et al. 2015). Straka et al. (2019) identified several indicators of intermittent flow in Czech streams (mostly Dytiscidae, Hydrophilidae, Helophoridae, and Hydraenidae), as well perennial streams (several Elimdae, as well as Dytiscidae, Gryinidae, Hydraenidae, and Scirtidae).

#### Odonata

Several studies support the use of Gomphidae and Cordulegastridae as indicators of perennial flow (e.g., Bonada et al. 2006; King et al. 2015). Straka et al. (2019) identified a Coenagrionidae species to be indicative of intermittent flows in Czech streams; while they found no taxa to be indicative of perennial flows, Cordulegastrid taxa were excluded from intermittent streams (in agreement with Blackburn 2012), whereas Calopterygidae were more widespread (in disagreement with Blackburn 2012).

#### Megaloptera

Corydalidae are listed as an indicator of perennial streams in Blackburn (2012), but some reports from montane regions in the arid southwest (e.g., Bogan and Lytle 2007) considered them to be indicative of intermittent conditions. Cover et al. (2015) describes two genus-groups within this family: The *Neohermes-Protochauliodes* group, which is well adapted to intermittency by building hyporheic aestivation chambers to survive the dry period (Figure 7), and the *Orohermes-Dysmicohermes* group, which does not burrow and is therefore restricted to perennial streams. Distinguishing the two genus-groups in the field may be possible, as the *Neohermes-Protochauliodes* group has distinctive head patterns in late instars (M. Cover, personal communication).



Figure 7. Neohermes aestivation chamber in a dry streambed in Arizona (courtesy M.T. Bogan).

#### Diptera

Cañedo-Argüelles et al. (2016) suggest that the diverse genera within Chironomidae may have strong preferences for certain flow duration conditions, which is supported by several other studies (e.g., Bonada et al. 2006; Miller and Brasher 2011). Herbst et al. (2019) found numerous midge taxa associated with perennial flows, while other taxa were associated with intermittent flows. Challenges with identifying this group in the field may make them impractical for use in a field-based flow duration assessment method.

#### Other aquatic invertebrates

In their study of Czech streams, Straka et al. (2019) identified several non-insect indicators of intermittent streams, including the flatworm *Mesastoma*, the nematomorph *Gordius*, several oligochaetes and leeches, and the isopod *Asllus aquaticus*. They also found numerous non-insect indicators of perennial flows, such as several flatworm species (e.g., *Dugesia, Polycelis*), several oligochaetes and leeches, the Hydracarina mites, and the amphipod *Gammarus fossarum*.

#### Algae

Algal biofilm, mats and other macroalgal forms are evident in most streams within a week of the onset of flow (even 1 day, in the case of biofilms), and thus their presence may not always be a good indicator of perennial or intermittent flow (Benenati et al. 1998, Robson et al. 2008, Corcoll et al. 2015). However, most studies suggest that macroalgal growth in the first two weeks may be limited, particularly in hydrologically isolated systems without access to perennial refugia (Robson et al. 2008). Thus, the abundance, rather than the presence of macroalgae may be an effective indicator of flow duration.

Taxonomic identity for most algal species is difficult to ascertain in the field, and they are therefore ill suited for use as a field-based flow duration indicator. However, several studies suggest that there are flow-duration affinities for several groups. For example, Benenati et al. (1998) showed that the macroalga *Cladophora* tend to dominate in perennial streams, while diatoms and the filamentous cyanobacterium *Oscillatoria* dominate in intermittent streams. Certain macroalgae groups are readily identifiable in the field (Entwisle et al. 1997), potentially providing sufficient information to inform flow duration assessment.

Dormant algal propagules may accumulate in the dry streambed and be resuscitated in lab conditions. This approach has been proposed as a way to assess ecological conditions of dry lakes and streambeds (Carvalho et al. 2002; Robson et al. 2008), and could be used to assess flow duration. But because of the intensive nature of this approach, it is not well suited for a rapid flow duration assessment method.

#### **Bryophytes**

The presence of "streamer mosses" is an indicator of intermittent or perennial flow duration in Topping et al. (2009). Several studies support this use (Fritz et al. 2009; Cole et al. 2010), and a number of taxa have been designated in terms of moisture preferences (e.g., Appendix A in Fritz et al. 2009). Vieira et al. (2012a, 2016) identified bryophyte community types characteristic of intermittent and perennial rivers in Mediterranean Europe. They found that intermittent rivers

were dominated by drought tolerant taxa (e.g., *Scorpiurium*), and upright acrocarpous annual forms, while perennial streams had more prostrate pleurocarpic perennial mats.

#### Riparian and wetland vascular plants

The presence of wetland indicator plants is an important indicator of flow duration in several methods, especially in Nadeau (2015), where it may the most important indicator in a dry stream reach. An advantage of riparian plants over other biological indicators of flow duration is that they are non-motile organisms, some of which have very long life-spans (i.e., decades). Therefore, they are well suited to reflect local, long-term conditions in a way that fish or invertebrates cannot.

Several studies show a very strong relationship between flow duration and plant communities (e.g., Caskey et al. 2015; Stromberg et al. 2007). Caskey et al. (2015) showed a decrease in wetland plant occurrence after diversion of perennial flow along stream reaches in the Routt National Forest, CO. Reynolds and Shafroth (2017) noted a number of plant species indicative of perennial versus intermittent flow regimes in high and low elevation streams in the Colorado Basin. Although that study did not identify ephemeral streams, the authors report that the driest streams in their study were dominated by upland plants, such as sagebrush and juniper (Lindsay Reynolds, personal communication). Thus, the taxonomic composition of riparian and wetland plants may be an effective indicator of flow duration.

#### Vertebrates

Several flow duration assessment methods use the presence of vertebrates as indicators of perennial or intermittent flow. The list of species used in Nadeau (2015) should be updated to include taxa found in the WM through consultation with regional experts. Habitat preferences of taxa specific to the WM will need to be developed if they are to be used as an indicator of flow duration classes.

### **PROPOSED INDICATORS**

For the present study, we will evaluate indicators for the Pacific Northwest (Nadeau 2015) and New Mexico (NMED 2011):

#### **Geomorphological indicators**

- Slope (Nadeau 2015)
- Sinuosity (NMED 2011)
- Floodplain and channel dimensions (aka, entrenchment ratio; NMED 2011)
- In-channel structure (NMED 2011)
- Substrate sorting (NMED 2011)

#### Hydrologic indicators

- Water in channel (NMED 2011)
- Hydric soils (NMED 2011)
- Sediment on plants and debris (NMED 2011)
- Seeps and springs (NMED 2011)

• Number of woody jams within 10 m of the reach

#### **Biological indicators**

#### Aquatic macroinvertebrates

- Presence of aquatic macroinvertebrates (Nadeau 2015). Early instars, partial terrestrial taxa, and aerially dispersing life stages will be noted separately, if encountered.
- Abundance of mayflies (Nadeau 2015). Again, early instars will be ignored.
- Presence of perennial indicator taxa (Nadeau 2015). Additional taxa recommended by the Southwest Association of Freshwater Invertebrate Taxonomists (SAFIT) will also be noted. Other taxa detected in the field will be reported at the family level, if possible. If lab-analyzed data are available from samples collected at the same site, these will be used as well.
- Available data should be evaluated to ascertain if indicator taxa can be identified in labanalyzed samples.

#### Terrestrial arthropods

- No indicators proposed for evaluation in the current study.
- Available data should be evaluated to ascertain if indicator taxa can be identified in labanalyzed samples.

#### Algae

- Presence of filamentous algae (NMED 2011)
- Presence of live or dead algal mats
- Available data should be evaluated to ascertain if indicator taxa can be identified in labanalyzed samples.

#### Bryophytes

- Presence of streamer mosses (Topping 2009)
- Presence of liverworts
- Presence of pleurocarp and acrocarp bryophytes in the channel and banks.
- Available data should be evaluated to ascertain if indicator taxa can be identified in labanalyzed samples.

#### Wetland and riparian plants

- Presence of FACW, OBL, and SAV plants, following Nadeau (2015). The regional plant list for the Arid West shall be used (Lichvar et al. 2016).
- Absence of rooted vegetation in thalweg (NMED 2011)
- Differences of vegetation between riparian zone and adjacent uplands (NMED 2011)

#### Vertebrates

• Presence of fish, reptiles, and amphibians (Nadeau 2015)

• Presence of fish (NMED 2011)

Source	Region	Notes	Indicator Class	Perennial	Intermittent	Ephemeral
Caskey et al. (2015)	Colorado Rocky Mtns. – Routt NF	Flow diversion experiment, summarizing vegetation changes above and below diversions	Vegetation taxa	Associated (labeled as obligate wetland species): <i>Carex</i> <i>utriculata,</i> <i>Mertensia ciliate,</i> <i>Salix planifolia,</i> <i>Salix wolfii,</i> <i>Veronica</i> <i>americana</i>		
Rader and Belish (1999)	Colorado Rocky Mtns. – St. Louis Creek	Flow diversion experiment, summarizing vegetation changes above and below diversions	Macroinvertebr -ate density	Associated taxa: Ephemeroptera, Baetis bicaudatus, Drunella coloradensis, Cinygmula spp.		
Bramblett and Fausch (1991)	Southeastern Colorado – Purgatoire River	Habitat and biota descriptions	Macroinvertebr -ate taxa	Associated taxa: Choroterpes mexicanus, Microcylloepus sp., Cheumatopsyche, Hydropsyche, Simuliidae		Distinctly missing functional feeding groups present in perennial rivers (scrapers and predators) Dominated by collector- gatherers
Reynolds and Shafroth (2017)	Upper Colorado River Basin	Riparian plant traits for high and low elevation streams	Vegetation taxa – at both high and low elevation	Associated taxa: Equisetum arvense, Rosa woodsii	Associated taxa: Sporobolus cryptandrus, Gutierrezia sarothrae	
Straka et al. (2019)	Central Europe *Not in WM	Species density and frequency of occurrence assessed for perennial, near- perennial and intermittent streams	Macroinvertebr -ate taxa	Strongly Associated Taxa: Dugesia gonocephala, Baetis muticus, Baetis rhodani s.1., Leuctra sp., Hydropsyche sp.	Strongly Associated Taxa: Eiseniella tetraedra, Brachyptera risi, Parametriocne- mus stylatus, Paraphaenoclad -ius sp., Marionina sp.	

#### Table 7. Select examples of ASW indicators and levels associated with flow-duration classes.

#### **BIBLIOGRAPHY**

#### Flow duration assessment methods

Here we present the sources reviewed for methodology or methodology validation in determining flow duration classes. Methods that were excluded based on the criteria in Figure 3 are presented separately; rationale for exclusion is provided in "*Notes*" for each entry. Where applicable, validation studies or studies associated with the development of a method are listed under "*Related Sources*".

#### Review

McCune, R.D. and Mazor, K. 2019. Review of flow duration methods and indicators of flow duration in the scientific literature: Arid Southwest. Technical Report #1063. Southern California Coastal Water Research Project. Costa Mesa, CA. <u>https://ftp.sccwrp.org/pub/download/DOCUMENTS/TechnicalReports/1063\_FlowMethodsReview.p</u> df

#### Methods

#### **Mediterranean Europe**

Gallart, F., Cid, N., Latron, J., Llorens, P., Bonada, N., Jeuffroy, J., ... Prat, N. (2017). TREHS: An open-access software tool for investigating and evaluating temporary river regimes as a first step for their ecological status assessment. *The Science of the Total Environment*, 607-608, 519–540. <u>https://doi.org/10.1016/j.scitotenv.2017.06.209</u>

#### **Related Sources**

Belmar, O., Velasco, J., & Martinez-Capel, F. (2011). Hydrological classification of natural flow regimes to support environmental flow assessments in intensively regulated Mediterranean rivers, Segura River Basin (Spain). *Environmental Management*, 47(5), 992–1004. https://doi.org/10.1007/s00267-011-9661-0

Gallart, F., Llorens, P., Latron, J., Cid, N., Rieradevall, M., & Prat, N. (2016). Validating alternative methodologies to estimate the regime of temporary rivers when flow data are unavailable. *The Science of the Total Environment*, *565*, 1001–1010. https://doi.org/10.1016/j.scitotenv.2016.05.116

Gallart, F., Prat, N., Garca-Roger, E. M., Latron, J., Rieradevall, M., Llorens, P., ... Froebrich, J. (2012). A novel approach to analysing the regimes of temporary streams in relation to their controls on the composition and structure of aquatic biota. *Hydrology and Earth System Sciences*, *16*(9), 3165–3182. https://doi.org/10.5194/hess-16-3165-2012

Prat, N., Gallart, F., Von Schiller, D., Polesello, S., García-Roger, E. M., Latron, J., ... Froebrich, J. (2014). THE MIRAGE TOOLBOX: AN INTEGRATED ASSESSMENT TOOL FOR TEMPORARY STREAMS. *River Research and Applications*, *30*(10), 1318– 1334. <u>https://doi.org/10.1002/rra.2757</u>

#### **New Mexico**

Surface Water Quality Bureau, New Mexico Environment Department. (2011). *HYDROLOGY PROTOCOL FOR THE DETERMINATION OF USES SUPPORTED BY EPHEMERAL, INTERMITTENT, AND PERENNIAL WATERS.* Retrieved from <u>https://www.env.nm.gov/swqb/documents/swqbdocs/MAS/Hydrology/HydrologyProtocolAPPR</u> <u>OVED05-2011.pdf</u>

#### **Related Sources**

Surface Water Quality Bureau, New Mexico Environment Department. (2011). HYDROLOGY PROTOCOL FOR THE DETERMINATION OF USES SUPPORTED BY EPHEMERAL, INTERMITTENT, AND PERENNIAL WATERS (Appendix 1). Retrieved from

https://www.env.nm.gov/swqb/documents/swqbdocs/MAS/Hydrology/Appendix1.pdf

N C division of water quality, New Mexico Environment Department. (2012). *Hydrology Protocol Use Attainability Analysis for an Ephemeral Stream (Cover Sheet, Dec 2012).* Retrieved from

https://www.env.nm.gov/swqb/documents/swqbdocs/MAS/Hydrology/HydrologyCoverS heetREVDec2012.docx

Division of water quality, New Mexico Environment Department. (2011). *Hydrology Determination Field Sheets (May 2011)*. Retrieved from <u>https://www.env.nm.gov/swqb/documents/swqbdocs/MAS/Hydrology/HydrologyFieldSh</u> <u>eetsREVMay2011.docx</u>

#### North Carolina

N C Division of Water Quality. (2010). *Methodology for Identification of Intermittent and Perennial Streams and their Origins* (Version 4.11). North Carolina Department of Environment and Natural Resources. Retrieved from http://portal.ncdenr.org/web/wq/swp/ws/401/waterresources/streamdeterminations

#### **Related Sources**

Fritz, K. M., Wenerick, W. R., & Kostich, M. S. (2013). A validation study of a rapid field-based rating system for discriminating among flow permanence classes of headwater streams in South Carolina. *Environmental Management*, *52*(5), 1286–1298. https://doi.org/10.1007/s00267-013-0158-x

#### **Pacific Northwest (Including Oregon)**

Nadeau, T.-L. (2015). *Streamflow Duration Assessment Method for the Pacific Northwest* (No. EPA 910-K-14-001). U.S. Environmental Protection Agency, Region 10, Seattle, WA.

#### **Related Sources**

Nadeau, T.-L. (2011). 2011 Streamflow Duration Assessment Method for Oregon (No. EPA 910-R-11-002). U.S. Environmental Protection Agency, Region 10.

Nadeau, T.-L., Leibowitz, S. G., Wigington, P. J., Jr, Ebersole, J. L., Fritz, K. M.,

Coulombe, R. A., ... Blocksom, K. A. (2015). Validation of rapid assessment methods to determine streamflow duration classes in the Pacific Northwest, USA. *Environmental Management*, *56*(1), 34–53. https://doi.org/10.1007/s00267-015-0466-4

Topping, B. J. D., Nadeau, T.-L., & Turaski, M. R. (2009). Oregon Streamflow Duration Assessment Method Interim Version - Interim version (March 2009).

## **Temperate US**

Fritz, K. M., Johnson, B. R., & Walters, D. M. (2006). *Field Operations Manual for Assessing the Hydrologic Permanence and Ecological Condition of Headwater Streams* (No. EPA/600/ R-06/126). U.S. Environmental Protection Agency, Office of Research and Development, Washington DC.

## **Excluded Methods**

Berhanu, B., Seleshi, Y., Demisse, S. S., & Melesse, A. M. (2015). Flow Regime Classification and Hydrological Characterization: A Case Study of Ethiopian Rivers. *WATER*, 7(6), 3149–3165. <u>https://doi.org/10.3390/w7063149</u>

Rational for Exclusion: Low utility. Analysis of flow duration class requires flow metrics extracted from observed or modeled daily flow data, and is therefore not a rapid method.

Berkowitz, J., Casper, A. F., & Noble, C. (2011). A multiple watershed field test of hydrogeomorphic functional assessment of headwater streams—Variability in field measurements between independent teams. *Ecological Indicators*, *11*(5), 1472–1475. https://doi.org/10.1016/j.ecolind.2011.01.004

Rational for Exclusion: Low applicability and utility. This study may be useful in estimating expected errors in field measured hydrogeomorphic flow duration indicators, but there is no direct application for these indicators in direct assessment of flow duration.

Kennard, M. J., Pusey, B. J., Olden, J. D., Mackay, S. J., Stein, J. L., & Marsh, N. (2010). Classification of natural flow regimes in Australia to support environmental flow management. *Freshwater Biology*, *55*(1), 171–193. <u>https://doi.org/10.1111/j.1365-2427.2009.02307.x</u>

Rational for Exclusion: Low utility. Methods of determining hydrologic regime here rely upon mean daily discharge data.

Noble, C. V., Berkowitz, J., & Spence, J. (2010). Operational Draft Regional Guidebook for the Functional Assessment of High-gradient Ephemeral and Intermittent Headwater Streams in Western West Virginia and Eastern Kentucky. US Army Corps of Engineers, Vicksburg. Retrieved from

http://www.lrh.usace.army.mil/Portals/38/docs/Operational%20Draft%20Regional%20Guideboo k1.pdf Rational for Exclusion: Low applicability. This study assesses indicators of habitat capacity and function at streams with flow duration known from intensive hydrologic data.

Ohio EPA. (2012). *Field Evaluation Manual for Ohio's Primary Headwater Streams (Version 3.0)*. Ohio EPA, Division of Surface Water. Retrieved from <a href="http://epa.ohio.gov/portals/35/wqs/headwaters/PHWHManual\_2012.pdf">http://epa.ohio.gov/portals/35/wqs/headwaters/PHWHManual\_2012.pdf</a>

Rational for Exclusion: Meets criteria. This citation was discovered after the initial literature search; it may be added in a future revision of the literature review. However, it was developed for a region with a different climate from the ASW.

Porras, A., & Scoggins, M. (2013). *The Flow Permanence Index: A Statistical Assessment of Flow Regime in Austin Streams*. City of Austin, Watershed Protection Department. Retrieved from <a href="http://www.austintexas.gov/watershed">http://www.austintexas.gov/watershed</a> protection/publications/document.cfm?id=213560

Rational for Exclusion: Low utility. The methodology relies on regionally specific hydrologic models, rather than field indicators.

Svec, J. R., Kolka, R. K., & Stringer, J. W. (2005). Defining perennial, intermittent, and ephemeral channels in Eastern Kentucky: Application to forestry best management practices. *Forest Ecology and Management*, *214*(1), 170–182. <u>https://doi.org/10.1016/j.foreco.2005.04.008</u>

Rational for Exclusion: Meets criteria. This citation was discovered after the initial literature search; it may be added in a future revision of the literature review. However, it was developed for a region with a different climate from the ASW.

Trubilowicz, J. W., Moore, R. D., & Buttle, J. M. (2013). Prediction of stream-flow regime using ecological classification zones. *Hydrological Processes*, *27*(13), 1935–1944. <u>https://doi.org/10.1002/hyp.9874</u>

Rational for Exclusion: Low utility. The methodology relies on regionally specific hydrologic models, rather than field indicators.

McCleary, R., Haslett, S., & Christie, K. (2012). Field Manual for Erosion-Based Channel Classification – Version 7.0. Fish and Watershed Program, Foothills Research Institute, Alberta, Canada.

Rational for Exclusion: Meets criteria. This methodology was discovered after the initial literature search and describes previously identified field indicators of streamflow duration, but the emphasis of this methodology is defining less specific flow duration classes and is specific for a climate different from that of the WM.

## Indicators

Biology

# Algae

Bechtold, H. A., Marcarelli, A. M., Baxter, C. V., & Inouye, R. S. (2012). Effects of N, P, and organic carbon on stream biofilm nutrient limitation and uptake in a semi-arid watershed. *Limnology and Oceanography*, *57*(5), 1544–1554.

Benenati, P. L., Shannon, J. P., & Blinn, D. W. (1998). Desiccation and recolonization of phytobenthos in a regulated desert river: Colorado River at Lees Ferry, Arizona, USA. *Regulated Rivers: Research & Management*, *14*(6), 519–532.

Carvalho, L., Bennion, H., Dawson, H., Furse, M., Gunn, I., Hughes, R., ... Others. (2002). Nutrient conditions for different levels of ecological status and biological quality in surface waters (Phase I). *Environment*, 2002. Retrieved from <u>http://www.academia.edu/download/44662450/Nutrient\_Conditions\_for\_Different\_Levels20160</u> <u>412-18347-xm5rby.pdf</u>

Corcoll, N., Casellas, M., Huerta, B., Guasch, H., Acuña, V., Rodríguez-Mozaz, S., ... Sabater, S. (2015). Effects of flow intermittency and pharmaceutical exposure on the structure and metabolism of stream biofilms. *The Science of the Total Environment*, *503-504*, 159–170.

Entwisle, T. J., Sonneman, J. A., & Lewis, S. H. (1997). *Freshwater algae in Australia*. Sainty & Associates.

Gillett, N. D., Pan, Y., Manoylov, K. M., Stancheva, R., & Weilhoefer, C. L. (2011). THE POTENTIAL INDICATOR VALUE OF RARE TAXA RICHNESS IN DIATOM-BASED STREAM BIOASSESSMENT 1. *Journal of Phycology*, *47*(3), 471–482.

Gray, D. W., Lewis, L. A., & Cardon, Z. G. (2007). Photosynthetic recovery following desiccation of desert green algae (Chlorophyta) and their aquatic relatives. *Plant, Cell & Environment*, *30*(10), 1240–1255.

Johansen, J. R., Rushforth, S. R., Orbendorfer, R., Fungladda, N., & Grimes, J. A. (1983). The algal flora of selected wet walls in Zion National Park, Utah, USA. *Nova Hedwigia. Lehre*, *38*, 765–774.

Robson, B. J., Matthews, T. Y. G., Lind, P. R., & Thomas, N. A. (2008). Pathways for algal recolonization in seasonally-flowing streams. *Freshwater Biology*, *53*(12), 2385–2401.

Shaver, M. L., Shannon, J. P., Wilson, K. P., Benenati, P. L., & Blinn, D. W. (1997). Effects of suspended sediment and desiccation on the benthic tailwater community in the Colorado River, USA. *Hydrobiologia*, *357*(1), 63–72.

Spindler, P. H., & Sommerfeld, M. R. (1996). Distribution of Algae in Pinal Creek, Gila County, Arizona. *Journal of the Arizona-Nevada Academy of Science*, *29*(2), 108–117.

Stanley, E. H., Fisher, S. G., Jones, Jr., & B., J. (2004). Effects of water loss on primary production: A landscape-scale model. *Aquatic Sciences*, *66*(1), 130–138.

Tornés, E., Acuña, V., Dahm, C. N., & Sabater, S. (2015). Flood disturbance effects on benthic diatom assemblage structure in a semiarid river network. *Journal of Phycology*, *51*(1), 133–143.

Usher, H. D., & Blinn, D. W. (1990). INFLUENCE OF VARIOUS EXPOSURE PERIODS ON THE BIOMASS AND CHLOROPHYLL A OF CLADOPHORA GLOMERATA (CHLOROPHYTA). *Journal of Phycology*, *26*(2), 244–249.

Vavilova, V. V., & Lewis, W. M., Jr. (1998). Temporal and altitudinal variations in the attached algae of mountain streams in Colorado. *Hydrobiologia*, *390*(1), 99–106.

## Amphibians & Snakes

Bateman, H. L., Harner, M. J., & Chung-MacCoubrey, A. (2008). Abundance and reproduction of toads (Bufo) along a regulated river in the southwestern United States: Importance of flooding in riparian ecosystems. *Journal of Arid Environments*, 72(9), 1613–1619.

Johnson, B. R., Fritz, K. M., Blocksom, K. A., & Walters, D. M. (2009). Larval salamanders and channel geomorphology are indicators of hydrologic permanence in forested headwater streams. *Ecological Indicators*, *9*(1), 150–159.

## **Benthic Macroinvertebrates**

Allen, G. H., Pavelsky, T. M., Barefoot, E. A., Lamb, M. P., Butman, D., Tashie, A., & Gleason, C. J. (2018). Similarity of stream width distributions across headwater systems. *Nature Communications*, *9*(1), 610.

Alyakrinskaya, I. O. (2004). Resistance to drying in aquatic mollusks. *Biology Bulletin*, 31(3).

Baron, J. S., LaFrancois, T., & Kondratieff, B. C. (1998). CHEMICAL AND BIOLOGICAL CHARACTERISTICS OF DESERT ROCK POOLS IN INTERMITTENT STREAMS OF CAPITOL REEF NATIONAL PARK, UTAH. *The Great Basin Naturalist*, *58*(3), 250–264.

Beche, L. A., Mcelravy, E. P., & Resh, V. H. (2006). Long-term seasonal variation in the biological traits of benthic-macroinvertebrates in two Mediterranean-climate streams in California, USA. *Freshwater Biology*, 51(1), 56–75.

Beche, L. A., Connors, P. G., Resh, V. H., & Merenlender, A. M. (2009). Resilience of fishes and invertebrates to prolonged drought in two California streams. *Ecography*, *32*(5), 778–788.

Bell, L. N. (1972). Notes on dry-season survival in two species of Elmidae (Coleoptera). *Pan Pacific Entomol*. Retrieved from <u>http://agris.fao.org/agris-</u>search/search.do?recordID=US201303215139

Blackburn, M., & Mazzacano, C. (2012). *Using aquatic macroinvertebrates as indicators of streamflow duration* (Washington and Idaho Indicators). Prepared for the U.S. Environmental Protection Agency, Region 10.

Blevins, E., Jepsen, S., Box, J. B., Nez, D., Howard, J., Maine, A., & O'Brien, C. (2017). Extinction risk of western north American freshwater mussels: Anodonta nuttalliana, the Anodonta oregonensis/kennerlyi clade, Gonidea angulata, and Margaritifera falcata. *Freshwater Mollusk Biology and Conservation*, 20, 71–88.

Bogan, M. T. (2017). Hurry up and wait: life cycle and distribution of an intermittent stream specialist (Mesocapnia arizonensis). *Freshwater Science*. https://doi.org/10.1086/694746

Bogan, M. T. and Lytle, D. A. (2007). Seasonal flow variation allows time-sharing by disparate aquatic insect communities in montane desert streams. *Freshwater Biology*, *52*(2), 290-304.

Bogan, M. T., Boersma, K. S., & Lytle, D. A. (2013). Flow intermittency alters longitudinal patterns of invertebrate diversity and assemblage composition in an arid-land stream network. *Freshwater Biology*, *58*(5), 1016–1028.

Bonada, N., Rieradevall, M., Prat, N., & Resh, V. H. (2006). Benthic macroinvertebrate assemblages and macrohabitat connectivity in Mediterranean-climate streams of northern California. *Journal of the North American Benthological Society*, *25*(1), 32–43.

Bottorff, R. L., & Knight, A. W. (1987). Functional organization of macroinvertebrate communities in two first-order California streams: Comparison of perennial and intermittent flow conditions. *SIL Proceedings*, *1922-2010*, *23*(2), 1147–1152.

Bramblett, R. G., & Fausch, K. D. (1991). Fishes, macroinvertebrates, and aquatic habitats of the Purgatoire River in Pinon Canyon, Colorado. *The Southwestern Naturalist*, 281–294.

Brasher, A. M. D., Albano, C. M., Close, R. N., Cannon, Q. H., & Miller, M. P. (2010). *Macroinvertebrate Communities and Habitat Characteristics in the Northern and Southern Colorado Plateau Networks: Pilot Protocol Implementation*. National Park Service, Fort Collins, Colorado.

Bryce, S. A., Lomnicky, G. A., & Kaufmann, P. R. (2010). Protecting sediment-sensitive aquatic species in mountain streams through the application of biologically based streambed sediment criteria. *Journal of the North American Benthological Society*, *29*(2), 657–672.

Byrne, R. A., & Mcmahon, R. F. (1994). Behavioral and Physiological Responses to Emersion in Freshwater Bivalves. *Integrative and Comparative Biology*, *34*(2), 194–204.

Cañedo-Argüelles, M., Bogan, M. T., Lytle, D. A., & Prat, N. (2016). Are Chironomidae (Diptera) good indicators of water scarcity? Dryland streams as a case study. *Ecological Indicators*, *71*, 155–162.

Canton, S. P., Cline, L. D., Short, R., & Ward, J. V. (1984). The macroinvertebrates and fish of a Colorado stream during a period of fluctuating discharge. *Freshwater Biology*, *14*(3), 311–316.

Clarke, A., Mac Nally, R., Bond, N., & Lake, P. S. (2010). Flow permanence affects aquatic

macroinvertebrate diversity and community structure in three headwater streams in a forested catchment. *Canadian Journal of Fisheries and Aquatic Sciences. Journal Canadien Des Sciences Halieutiques et Aquatiques*, 67(10), 1649–1657.

Clements, W. H., Carlisle, D. M., Lazorchak, J. M., & Johnson, P. C. (2000). HEAVY METALS STRUCTURE BENTHIC COMMUNITIES IN COLORADO MOUNTAIN STREAMS. *Ecological Applications: A Publication of the Ecological Society of America*, *10*(2), 626–638.

Cover, M. R., Seo, J. H., & Resh, V. H. (2015). Life History, Burrowing Behavior, and Distribution of Neohermes filicornis (Megaloptera: Corydalidae), a Long-Lived Aquatic Insect in Intermittent Streams. *Western North American Naturalist / Brigham Young University*, 75(4), 474–490.

De Jong, G. D., Canton, S. P., Lynch, J. S., & Murphy, M. (2015). Aquatic Invertebrate and Vertebrate Communities of Ephemeral Stream Ecosystems In the Arid Southwestern United States. *The Southwestern Naturalist*, *60*(4), 349–359.

De Jong, G. D., Smith, E. R., & Conklin, D. J., JR. (2013). RIFFLE BEETLE COMMUNITIES OF PERENNIAL AND INTERMITTENT STREAMS IN NORTHERN NEVADA, USA, WITH A NEW STATE RECORD FOR OPTIOSERVUS CASTANEIPENNIS (FALL) (COLEOPTERA: ELMIDAE). *The Coleopterists' Bulletin*, 67(3), 1–9.

del Rosario, R. B., & Resh, V. H. (2000). Invertebrates in intermittent and perennial streams: is the hyporheic zone a refuge from drying? *Journal of the North American Benthological Society*, *19*(4), 680–696.

Engelke, M. J., Jr. (1980). Aestivation of a water scavenger beetle in southwestern Wyoming (Coleoptera: Hydrophilidae). *The Coleopterists' Bulletin*, 176–176.

Erman, N. A. (1998). Invertebrate richness and Trichoptera phenology in Sierra Nevada (California, USA) cold springs: sources of variation. *L. Botosaneanu. Studies in Crenobiology. The Biology of Springs and Springbrooks. Backhuys Publishers, Leiden, The Netherlands*, 95–108.

Erman, N. A., & Erman, D. C. (1995). Spring Permanence, Trichoptera Species Richness, and the Role of Drought. *Journal of the Kansas Entomological Society*, *68*(2), 50–64.

Finn, D. S., & Leroy Poff, N. (2005). Variability and convergence in benthic communities along the longitudinal gradients of four physically similar Rocky Mountain streams. *Freshwater Biology*, *50*(2), 243–261.

Griffith, M. B., Hill, B. H., McCormick, F. H., Kaufmann, P. R., Herlihy, A. T., & Selle, A. R. (2005). Comparative application of indices of biotic integrity based on periphyton, macroinvertebrates, and fish to southern Rocky Mountain streams. *Ecological Indicators*, *5*(2), 117–136.

Hall, L. W., Jr, & Killen, W. D. (2005). Temporal and spatial assessment of water quality, physical habitat, and benthic communities in an impaired agricultural stream in California's San Joaquin Valley. *Journal of Environmental Science and Health. Part A, Toxic/hazardous Substances & Environmental Engineering*, 40(5), 959–989.

Harper, M. P., & Peckarsky, B. L. (2006). Emergence cues of a mayfly in a high-altitude stream ecosystem: potential response to climate change. *Ecological Applications: A Publication of the Ecological Society of America*, *16*(2), 612–621.

Herbst, D. B., Bogan, M. T., & Lusardi, R. A. (2008). Low specific conductivity limits growth and survival of the New Zealand mud snail from the Upper Owens River, California. *Western North American Naturalist / Brigham Young University*, 68(3), 324–333.

Herbst, D. B., Cooper, S. D., Medhurst, R. B., Wiseman, S. W., & Hunsaker, C. T. (2019). Drought ecohydrology alters the structure and function of benthic invertebrate communities in mountain streams. *Freshwater Biology*, *64*(5), 886–902.

Holmquist, J. G., Jones, J. R., Schmidt-Gengenbach, J., Pierotti, L. F., & Love, J. P. (2011). Terrestrial and Aquatic Macroinvertebrate Assemblages as a Function of Wetland Type across a Mountain Landscape. *Arctic, Antarctic, and Alpine Research*, 43(4), 568–584.

King, R. S., Scoggins, M., & Porras, A. (2015). Stream biodiversity is disproportionately lost to urbanization when flow permanence declines: evidence from southwestern North America. *Freshwater Science*. <u>https://doi.org/10.1086/684943</u>

Knight, A. W., & Gaufin, A. R. (1967). Stream Type Selection and Associations of Stoneflies (Plecoptera) in a Colorado River Drainage System. *Journal of the Kansas Entomological Society*, 40(3), 347–352.

Konrad, C. P., Brasher, A. M. D., & May, J. T. (2008). Assessing streamflow characteristics as limiting factors on benthic invertebrate assemblages in streams across the western United States. *Freshwater Biology*, *53*(10), 1983–1998.

Lind, O. T. (1982). Biogeographic affinities of benthic communities in isolated desert aquatic ecosystems. *Proceedings of Symposium on Recent Benthological Investigations in Texas and Adjacent States (JR Davis Ed.), Texas Acad. Sci*, 69–78.

Lusardi, R. A., Bogan, M. T., Moyle, P. B., & Dahlgren, R. A. (2016). Environment shapes invertebrate assemblage structure differences between volcanic spring-fed and runoff rivers in northern California. *Freshwater Science*. <u>https://doi.org/10.1086/687114</u>

McArthur, J. V., & Barnes, J. R. (1985). PATTERNS OF MACROINVERTEBRATE COLONIZATION IN AN INTERMITTENT ROCKY MOUNTAIN STREAM IN UTAH. *The Great Basin Naturalist*, *45*(1), 117–123.

Meyerhoff, R. D., & Lind, O. T. (1987). Factors Affecting the Benthic Community Structure of a Discontinuous Stream in Guadelupe Mountains National Park, Texas. *Internationale Revue Der Gesamten Hydrobiologie Und Hydrographie*, 72(3), 283–296.

Miller, M. P., Blinn, D. W., & Keim, P. (2002). Correlations between observed dispersal capabilities and patterns of genetic differentiation in populations of four aquatic insect species from the Arizona White Mountains, USA. *Freshwater Biology*, *47*(9), 1660–1673.

Miller, M. P., & Brasher, A. (2011). Differences in macroinvertebrate community structure in streams and rivers with different hydrologic regimes in the semi-arid Colorado Plateau. *River Systems*, *19*(3), 225–238.

Monk, W. A., Wood, P. J., Hannah, D. M., Wilson, D. A., Extence, C. A., & Chadd, R. P. (2006). Flow variability and macroinvertebrate community response within riverine systems. *River Research and Applications*, 22(5), 595–615.

Rader, R. B., & Belish, T. A. (1999). Influence of mild to severe flow alterations on invertebrates in three mountain streams. *Regulated Rivers: Research & Management: An International Journal Devoted to River Research and Management*, 15(4), 353–363.

Resh, V. H., & Rosenberg, D. M. (2010). Recent trends in life-history research on benthic macroinvertebrates. *Journal of the North American Benthological Society*, 29(1), 207–219.

Sada, D. W., Fleishman, E., & Murphy, D. D. (2005). Associations among spring-dependent aquatic assemblages and environmental and land use gradients in a Mojave Desert mountain range. *Diversity and Distributions*, 11(1), 91–99.

Schröder, M., Kiesel, J., Schattmann, A., Jähnig, S. C., Lorenz, A. W., Kramm, S., ... Hering, D. (2013). Substratum associations of benthic invertebrates in lowland and mountain streams. *Ecological Indicators*, *30*, 178–189.

Shaver, M. L., Shannon, J. P., Wilson, K. P., Benenati, P. L., & Blinn, D. W. (1997). Effects of suspended sediment and desiccation on the benthic tailwater community in the Colorado River, USA. *Hydrobiologia*, *357*(1), 63–72.

Shepard, W. D. (2011). Survival of stream dewatering by Postelichus immsi (Hinton)(Coleoptera Byrrhoidea: Dryopidae). *The Pan-Pacific Entomologist*.

Short, R. A., & Ward, J. V. (1980). Macroinvertebrates of a Colorado High Mountain Stream. *The Southwestern Naturalist*, 25(1), 23–32.

Straka, M., Polášek, M., Syrovátka, V., Stubbington, R., Zahrádková, S., Němejcová, D., ... Pařil, P. (2019). Recognition of stream drying based on benthic macroinvertebrates: A new tool in Central Europe. Ecological Indicators, 106, 105486.

Stribling, J. B., Jessup, B. K., & Feldman, D. L. (2008). Precision of benthic macroinvertebrate indicators of stream condition in Montana. *Journal of the North American Benthological Society*, 27(1), 58–67.

Tronstad, L. M., Brown, K. M., & Andersen, M. D. (2018). Using species distribution models to guide field surveys for an apparently rare aquatic beetle. <u>https://doi.org/10.3996/112016-JFWM-085</u>

URS, & GEI Consultants. (2006). Arid West Water Quality Research Project: Aquatic Communities of Ephemeral Stream Ecosystems (No. Final report).

Voelz, N. J., Zuellig, R. E., Shieh, S.-H., & Ward, J. V. (2005). The effects of urban areas on benthic macroinvertebrates in two Colorado Plains rivers. *Environmental Monitoring and Assessment*, *101*(1-3), 175–202.

Ward, J. V., & Short, R. A. (1978). Macroinvertebrate community structure of four special lotic habitats in Colorado, USA. *Verhandlungen Der Internationalen Vereinigung Fur Theoretische Und Angewandte Limnologie. International Association of Theoretical and Applied Limnology*, 20(2), 1382–1387.

Wu, D., & Legg, D. (2007). Structures of benthic insect communities in two southeastern Wyoming (USA) streams: similarities and differences among spatial units at different local scales. *Hydrobiologia*, *579*(1), 279–289.

## **Bryophytes**

Brusven, M. A., Meehan, W. R., & Biggam, R. C. (1990). The role of aquatic moss on community composition and drift of fish-food organisms. *Hydrobiologia*, 196(1), 39–50.

Cole, C., Stark, L. R., Bonine, M. L., & McLetchie, D. N. (2010). Transplant Survivorship of Bryophyte Soil Crusts in the Mojave Desert. *Restoration Ecology*, *18*(2), 198–205.

During, H. J. (1979). Life Strategies of Bryophytes: A Preliminary Review. *Lindbergia*, 5(1), 2–18.

Eldridge, D. J., & Rosentreter, R. (1999). Morphological groups: a framework for monitoring microphytic crusts in arid landscapes. *Journal of Arid Environments*, *41*(1), 11–25.

Fritz, K. M., Glime, J. M., Hribljan, J., & Greenwood, J. L. (2009). Can bryophytes be used to characterize hydrologic permanence in forested headwater streams? *Ecological Indicators*, *9*(4), 681–692.

Gillrich, J. J., & Bowman, K. C. (2010). *The use of bryophytes as indicators of hydric soils and wetland hydrology during wetland delineations in the United States.* 

Gimingham, C. H., & Birse, E. M. (1957). Ecological Studies on Growth-Form in Bryophytes: I. Correlations Between Growth-Form and Habitat. *The Journal of Ecology*, *45*(2), 533–545.

Glime, J. M. (1971). Response of Two Species of Fontinalis to Field Isolation from Stream Water. *The Bryologist*, 74(3), 383–386.

Glime, J. M., & Vitt, D. H. (1984). The physiological adaptations of aquatic Musci. *Lindbergia*, *10*(1), 41–52.

Maurer, M. A., & Brusven, M. A. (1983). Insect abundance and colonization rate in Fontinalis neo-mexicana (Bryophyta) in an Idaho Batholith stream, USA. *Hydrobiologia*, *98*(1), 9–15.

Rosentreter, R. (1984). The Zonalion of Mosses and Lichens Along the Salmon River in Idaho.

Slack, N. G. (1990). Bryophytes and ecological niche theory. *Botanical Journal of the Linnean Society. Linnean Society of London*, *104*(1-3), 187–213.

Vieira, C., Aguiar, F. C., Portela, A. P., Monteiro, J., Raven, P. J., Holmes, N. T. H., ... Ferreira, M. T. (2016). Bryophyte communities of Mediterranean Europe: a first approach to model their potential distribution in highly seasonal rivers. *Hydrobiologia*, 1–17.

Vieira, C., Séneca, A., Sérgio, C., & Ferreira, M. T. (2012a). Bryophyte taxonomic and functional groups as indicators of fine scale ecological gradients in mountain streams. *Ecological Indicators*, *18*(Supplement C), 98–107.

Vieira, C., Séneca, A., Sérgio, C., & Ferreira, M. T. (2012b). Bryophyte taxonomic and functional groups as indicators of fine scale ecological gradients in mountain streams. *Ecological Indicators*, *18*, 98–107.

Vitt, D. H., & Glime, J. M. (1984). The Structural Adaptations of Aquatic Musci. *Lindbergia*, *10*(2), 95–110.

## Vertebrates

Bliesner, A. K., & Robison, E. G. (2007). Detecting the Upstream Extent of Fish in the Redwood Region of Northern California. *In: Standiford, Richard B.; Giusti, Gregory A.; Valachovic, Yana; Zielinski, William J.; Furniss, Michael J., Technical Editors. 2007. Proceedings of the Redwood Region Forest Science Symposium: What Does the Future Hold? Gen. Tech. Rep. PSW-GTR-194. Albany, CA: Pacific Southwest Research Station, Forest Service, US Department of Agriculture; P. 135-146, 194.* 

Bondi, C. A., & Marks, S. B. (2013). Differences in flow regime influence the seasonal migrations, body size, and body condition of western pond turtles (Actinemys marmorata) that inhabit perennial and intermittent riverine sites in northern California. *Copeia*, 2013(1), 142–153.

Boughton, D. A., Fish, H., Pope, J., & Holt, G. (2009). Spatial patterning of habitat for Oncorhynchus mykiss in a system of intermittent and perennial streams. *Ecology of Freshwater Fish*, *18*(1), 92–105.

Bowen, H. L., & Marchetti, M. P. (2016). Ecomorphological plasticity of juvenile fall-run chinook salmon (Oncorhynchus tshawytscha) in perennial and ephemeral streams. *Environmental Biology of Fishes*, *99*(1), 67–78.

Bryce, S. A., Lomnicky, G. A., & Kaufmann, P. R. (2010). Protecting sediment-sensitive aquatic species in mountain streams through the application of biologically based streambed sediment criteria. *Journal of the North American Benthological Society*, *29*(2), 657–672.

Canton, S. P., Cline, L. D., Short, R., & Ward, J. V. (1984). The macroinvertebrates and fish of a Colorado stream during a period of fluctuating discharge. *Freshwater Biology*, *14*(3), 311–316.

De Jong, G. D., Canton, S. P., Lynch, J. S., & Murphy, M. (2015). Aquatic Invertebrate and Vertebrate Communities of Ephemeral Stream Ecosystems In the Arid Southwestern United States. *The Southwestern Naturalist*, *60*(4), 349–359.

Diller, L. V., & Wallace, R. L. (1999). Distribution and Habitat of Ascaphus truei in Streams on Managed, Young Growth Forests in North Coastal California. *Journal of Herpetology*, *33*(1), 71–79.

Feral, D., Camann, M. A., & Welsh, H. H., Jr. (2005). Dicamptodon tenebrosus larvae within hyporheic zones of intermittent streams in California. *Herpetological Review, Vol. 36 (1): 26-27.* 

Fox, J. T., & Magoulick, D. D. (2019). Predicting hydrologic disturbance of streams using species occurrence data. *The Science of the Total Environment*, 686, 254–263.

Griffith, M. B., Hill, B. H., McCormick, F. H., Kaufmann, P. R., Herlihy, A. T., & Selle, A. R. (2005). Comparative application of indices of biotic integrity based on periphyton, macroinvertebrates, and fish to southern Rocky Mountain streams. *Ecological Indicators*, *5*(2), 117–136.

Harvey, B. C., Nakamoto, R. J., & White, J. L. (2006). Reduced streamflow lowers dry-season growth of rainbow trout in a small stream. *Transactions of the American Fisheries Society*, *135*(4), 998–1005.

Hooley-Underwood, Z. E., Stevens, S. B., Salinas, N. R., & Thompson, K. G. (2019). An Intermittent Stream Supports Extensive Spawning of Large-River Native Fishes. *Transactions of the American Fisheries Society*, *148*(2), 426–441.

Hughes, R. M., Herlihy, A. T., & Sifneos, J. C. (2015). Predicting aquatic vertebrate assemblages from environmental variables at three multistate geographic extents of the western USA. *Ecological Indicators*, *57*, 546–556.

Jeffres, C. A., Opperman, J. J., & Moyle, P. B. (2008). Ephemeral floodplain habitats provide best growth conditions for juvenile Chinook salmon in a California river. *Environmental Biology of Fishes*, *83*(4), 449–458.

Leidy, R. A., Bogan, M. T., Neuhaus, L., Rosetti, L., & Carlson, S. M. (2016). Summer die-off of western pond turtle (Actinemys marmorata) along an intermittent coast range stream in central California. *The Southwestern Naturalist*, *61*(1), 71–75.

Lorig, R. C., Marchetti, M. P., & Kopp, G. (2013). Spatial and temporal distribution of native fish larvae in seasonal and perennial tributaries of the Sacramento River, CA, USA. *Journal of Freshwater Ecology*, *28*(2), 179–197.

Miller, D. A. W., Brehme, C. S., Hines, J. E., Nichols, J. D., & Fisher, R. N. (2012). Joint estimation of habitat dynamics and species interactions: disturbance reduces co-occurrence of

non-native predators with an endangered toad. *The Journal of Animal Ecology*, 81(6), 1288–1297.

Osmundson, D. B., Ryel, R. J., Lamarra, V. L., & Pitlick, J. (2002). FLOW–SEDIMENT– BIOTA RELATIONS: IMPLICATIONS FOR RIVER REGULATION EFFECTS ON NATIVE FISH ABUNDANCE. *Ecological Applications: A Publication of the Ecological Society of America*, 12(6), 1719–1739.

Power, M. E., Bouma-Gregson, K., Higgins, P., & Carlson, S. M. (2015). The Thirsty Eel: Summer and Winter Flow Thresholds that Tilt the Eel River of Northwestern California from Salmon-Supporting to Cyanobacterially Degraded States. *Copeia*, *103*(1), 200–211.

Ruso, G. E., Meyer, E., & Das, A. J. (2017). Seasonal and Diel Environmental Conditions Predict Western Pond Turtle (Emys marmorata) Behavior at a Perennial and an Ephemeral Stream in Sequoia National Park, California. *Chelonian Conservation and Biology: Journal of the IUCN/SSC Tortoise and Freshwater Turtle Specialist Group and International Bulletin of Chelonian Research*, 16(1), 20–28.

Snyder, G. K., & Hammerson, G. A. (1993). Interrelationships between water economy and thermo-regulation in the Canyon tree-frog Hyla arenicolor. *Journal of Arid Environments*, 25(3), 321–329.

Stoddard, M. A., & Hayes, J. P. (2005). THE INFLUENCE OF FOREST MANAGEMENT ON HEADWATER STREAM AMPHIBIANS AT MULTIPLE SPATIAL SCALES. *Ecological Applications: A Publication of the Ecological Society of America*, *15*(3), 811–823.

Tate, K. W., Lancaster, D. L., & Lile, D. F. (2007). Assessment of thermal stratification within stream pools as a mechanism to provide refugia for native trout in hot, arid rangelands. *Environmental Monitoring and Assessment*, *124*(1-3), 289–300.

Vinson, M., & Levesque, S. (1994). REDBAND TROUT RESPONSE TO HYPOXIA IN A NATURAL ENVIRONMENT. *The Great Basin Naturalist*, *54*(2), 150–155.

Waters, J. R., Zabel, C. J., McKelvey, K. S., & Welsh, H. H. (2001). Vegetation patterns and abundances of amphibians and small mammals along small streams in a northwestern California watershed. *Northwest Science*. *75 (1): 37-52.*, *75*(1), 37–52.

Welsh, H. H., Jr, Hodgson, G. R., & Lind, A. J. (2005). Ecogeography of the herpetofauna of a northern California watershed: linking species patterns to landscape processes. *Ecography*, 28(4), 521–536.

Woelfle-Erskine, C., Larsen, L. G., & Carlson, S. M. (2017). Abiotic habitat thresholds for salmonid over-summer survival in intermittent streams. *Ecosphere*, *8*(2), e01645.

## Microbes

Beauchamp, V. B., Stromberg, J. C., & Stutz, J. C. (2007). Flow regulation has minimal influence on mycorrhizal fungi of a semi-arid floodplain ecosystem despite changes in hydrology, soils, and vegetation. *Journal of Arid Environments*, *68*(2), 188–205.

#### Vegetation

Baron, J. S., LaFrancois, T., & Kondratieff, B. C. (1998). CHEMICAL AND BIOLOGICAL CHARACTERISTICS OF DESERT ROCK POOLS IN INTERMITTENT STREAMS OF CAPITOL REEF NATIONAL PARK, UTAH. *The Great Basin Naturalist*, *58*(3), 250–264.

Beauchamp, V. B., Stromberg, J. C., & Stutz, J. C. (2007). Flow regulation has minimal influence on mycorrhizal fungi of a semi-arid floodplain ecosystem despite changes in hydrology, soils, and vegetation. *Journal of Arid Environments*, 68(2), 188–205.

Caskey, S. T., Blaschak, T. S., Wohl, E., Schnackenberg, E., Merritt, D. M., & Dwire, K. A. (2015). Downstream effects of stream flow diversion on channel characteristics and riparian vegetation in the Colorado Rocky Mountains, USA. *Earth Surface Processes and Landforms*, *40*(5), 586–598.

Cooper, D. J., Kaczynski, K. M., Sueltenfuss, J., Gaucherand, S., & Hazen, C. (2017). Mountain wetland restoration: The role of hydrologic regime and plant introductions after 15 years in the Colorado Rocky Mountains, USA. *Ecological Engineering*, *101*, 46–59.

Gage, E. A., Cooper, D. J., Bultema, B., McKernan, C., & Lichvar, R. (2016). Developing a Field-Tested Wetland Indicator Rating for Blue Spruce (Picea Pungens) in the Southern Rocky Mountains. *Wetlands*, *36*(1), 111–120.

Gosejohan, M. C., Weisberg, P. J., & Merriam, K. E. (2017). Hydrologic Influences on Plant Community Structure in Vernal Pools of Northeastern California. *Wetlands*, *37*(2), 257–268.

Hammersmark, C. T., Rains, M. C., Wickland, A. C., & Mount, J. F. (2009). Vegetation and water-table relationships in a hydrologically restored riparian meadow. *Wetlands*, *29*(3), 785–797.

Hammersmark, C. T., Dobrowski, S. Z., Rains, M. C., & Mount, J. F. (2010). Simulated effects of stream restoration on the distribution of wet-meadow vegetation. *Restoration Ecology*, *18*(6), 882–893.

Harris, R. R. (1987). Occurrence of Vegetation on Geomorphic Surfaces in the Active Floodplain of a California Alluvial Stream. *The American Midland Naturalist*, *118*(2), 393–405.

Henszey, R. J., Skinner, Q. D., & Wesche, T. A. (1991). Response of montane meadow vegetation after two years of streamflow augmentation. *Regulated Rivers: Research & Management*, 6(1), 29–38.

Katz, G. L., Denslow, M. W., & Stromberg, J. C. (2012). The Goldilocks effect: intermittent streams sustain more plant species than those with perennial or ephemeral flow. *Freshwater Biology*, *57*(3), 467–480.

Lowry, C. S., Loheide, S. P., II, Moore, C. E., & Lundquist, J. D. (2011). Groundwater controls on vegetation composition and patterning in mountain meadows. *Water Resources Research*, *47*(10), 30.

McBride, J. R., & Strahan, J. (1984). Establishment and Survival of Woody Riparian Species on Gravel Bars of an Intermittent Stream. *The American Midland Naturalist*, *112*(2), 235–245.

Reynolds, L. V., & Shafroth, P. B. (2017). Riparian plant composition along hydrologic gradients in a dryland river basin and implications for a warming climate. *Ecohydrology*, *10*(6).

Rosentreter, R. (U.S. Department of the Interior, Idaho State Office, Boise, ID). (aug1992). High-water indicator plants along Idaho waterways. *General Technical Report INT - U.S. Department of Agriculture, Forest Service, Intermountain Research Station (USA)*.

Rowlands, P. G., Johnson, H. G., Avery, C. C., & Brian, N. J. (1995). *The Effect of Dewatering a Stream on Its Riparian Stream: A Case Study from Northern Arizona*.

Smith, J. J. (1989). Recovery of riparian vegetation on an intermittent stream following removal of cattle. *In: Abell, Dana L., Technical Coordinator. 1989. Proceedings of the California Riparian Systems Conference: Protection, Management, and Restoration for the 1990s; 1988 September 22-24; Davis, CA. Gen. Tech. Rep. PSW-GTR-110. Berkeley, CA: Pacific Southwest Forest and Range Experiment Station, Forest Service, US Department of Agriculture; P. 217-221, 110.* 

Smith, M. A., Rodgers, J. D., Dodd, J. L., & Skinner, Q. D. (1992). Habitat selection by cattle along an ephemeral channel. *Rangeland Ecology & Management / Journal of Range Management Archives*, *45*(4), 385–390.

Smith, M. A., Dodd, J. L., Skinner, Q. D., & Rodgers, J. D. (1993). Dynamics of vegetation along and adjacent to an ephemeral channel. *Rangeland Ecology & Management / Journal of Range Management Archives*, 46(1), 56–64.

Stromberg, J. C. (2001). Influence of stream flow regime and temperature on growth rate of the riparian tree, Platanus wrightii, in Arizona. *Freshwater Biology*, *46*(2), 227–239.

Stromberg, J. C., Lite, S. J., Marler, R., Paradzick, C., Shafroth, P. B., Shorrock, D., ... White, M. S. (2007). Altered stream-flow regimes and invasive plant species: the Tamarix case. *Global Ecology and Biogeography: A Journal of Macroecology*, *16*(3), 381–393.

U.S. Army Corps of Engineers. (2010). *Regional Supplement to the Corps of Engineers Wetland Delineation Manual: Western Mountains, Valleys, and Coast Region (Version 2.0)*. Retrieved from https://www.nrcs.usda.gov/Internet/FSE\_DOCUMENTS/stelprdb1046494.pdf

Geomorphology

Caskey, S. T., Blaschak, T. S., Wohl, E., Schnackenberg, E., Merritt, D. M., & Dwire, K. A. (2015). Downstream effects of stream flow diversion on channel characteristics and riparian vegetation in the Colorado Rocky Mountains, USA. *Earth Surface Processes and Landforms*, 40(5), 586–598.

Chin, A. (1999). The morphologic structure of step-pools in mountain streams. *Geomorphology*, 27(3), 191–204.

Ford, T. D., & Pedley, H. M. (1996). A review of tufa and travertine deposits of the world. *Earth-Science Reviews*, Vol. 41, pp. 117–175. <u>https://doi.org/10.1016/s0012-8252(96)00030-x</u>

Friedman, J. M., & Lee, V. J. (2002). EXTREME FLOODS, CHANNEL CHANGE, AND RIPARIAN FORESTS ALONG EPHEMERAL STREAMS. *Ecological Monographs*, 72(3), 409–425.

Hall, L. W., Jr, & Killen, W. D. (2005). Temporal and spatial assessment of water quality, physical habitat, and benthic communities in an impaired agricultural stream in California's San Joaquin Valley. *Journal of Environmental Science and Health. Part A, Toxic/hazardous Substances & Environmental Engineering*, 40(5), 959–989.

Hammersmark, C. T., Rains, M. C., & Mount, J. F. (2008). Quantifying the hydrological effects of stream restoration in a montane meadow, northern California, USA. *River Research and Applications*, *24*(6), 735–753.

Hedman, E. R., & Osterkamp, W. R. (1982). *Streamflow characteristics related to channel geometry of streams in western United States*.

Heede, B. H. (1979). Deteriorated watersheds can be restored: A case study. *Environmental Management*, *3*(3), 271–281.

Johnson, B. R., Fritz, K. M., Blocksom, K. A., & Walters, D. M. (2009). Larval salamanders and channel geomorphology are indicators of hydrologic permanence in forested headwater streams. *Ecological Indicators*, *9*(1), 150–159.

Johnson, J. P. L., Whipple, K. X., & Sklar, L. S. (2010). Contrasting bedrock incision rates from snowmelt and flash floods in the Henry Mountains, Utah. *GSA Bulletin*, *122*(9-10), 1600–1615.

Mersel, M. K., & Lichvar, R. W. (2014). *A guide to ordinary high water mark (OHWM) delineation for non-perennial streams in the western mountains, valleys, and coast region of the United States.* ENGINEER RESEARCH AND DEVELOPMENT CENTER HANOVER NH COLD REGIONS RESEARCH AND ENGINEERING LAB. Retrieved from <u>https://erdclibrary.erdc.dren.mil/xmlui/bitstream/handle/11681/5501/ERDC-CRREL-TR-14-</u> <u>13.pdf?sequence=1&isAllowed=y</u>

Mersel, M. K., Lichvar, R. W., Gillrich, J. J., & Lefebvre, L. E. (2014). Occurrence and distribution of ordinary high water mark (OHWM) indicators in non-perennial streams in the western mountains, valleys, and coast region of the United States. ENGINEER RESEARCH AND DEVELOPMENT CENTER HANOVER NH COLD REGIONS RESEARCH AND ENGINEERING LAB. Retrieved from http://www.dtic.mil/docs/citations/ADA608562

Miller, J. P. (1958). High mountain streams: effects of geology on channel characteristics and bed material.

Moore, D. O. (1968). ESTIMATING MEAN RUNOFF IN UNGAGED SEMIARID AREAS. Bulletin of the International Association of Scientific Hydrology, 13(1), 29–39.

O'Dowd, A. P., & Chin, A. (2016). Do bio-physical attributes of steps and pools differ in high-gradient mountain streams? *Hydrobiologia*, 776(1), 67–83.

Osterkamp, W. R. (1998). Processes of fluvial island formation, with examples from Plum Creek, Colorado and Snake River, Idaho. *Wetlands*, *18*(4), 530–545.

Pollock, M. M., Beechie, T. J., & Jordan, C. E. (2007). Geomorphic changes upstream of beaver dams in Bridge Creek, an incised stream channel in the interior Columbia River basin, eastern Oregon. *Earth Surface Processes and Landforms*, *32*(8), 1174–1185.

Polvi, L. E., Wohl, E. E., & Merritt, D. M. (2011). Geomorphic and process domain controls on riparian zones in the Colorado Front Range. *Geomorphology*, *125*(4), 504–516.

Ponce, V. M., & Lindquist, D. S. (1990). MANAGEMENT OF BASEFLOW AUGMENTATION: A REVIEW. *JAWRA Journal of the American Water Resources Association*, *26*(2), 259–268.

Ryan, S. E., Porth, L. S., & Troendle, C. A. (2005). Coarse sediment transport in mountain streams in Colorado and Wyoming, USA. *Earth Surface Processes and Landforms*, *30*(3), 269–288.

Shah, J. J. F., & Dahm, C. N. (2008). Flood regime and leaf fall determine soil inorganic nitrogen dynamics in semiarid riparian forests. *Ecological Applications: A Publication of the Ecological Society of America*, 18(3), 771–788.

Whiting, P. J., & Moog, D. B. (2001). The geometric, sedimentologic and hydrologic attributes of spring-dominated channels in volcanic areas. *Geomorphology*, *39*(3), 131–149.

Wohl, E., & Merritt, D. M. (2008). Reach-scale channel geometry of mountain streams. *Geomorphology*, *93*(3), 168–185.

Wright, J. S. (2000). Tufa accumulations in ephemeral streams: observations from The Kimberley, north-west Australia. *The Australian Geographer*, *31*(3), 333–347.

Hydrology

Abbe, T. B., & Montgomery, D. R. (1996). Large woody debris jams, channel hydraulics and habitat formation in large rivers. *Regulated Rivers: Research & Management*, 12(2-3), 201–221.

Andersen, D. C., Nelson, S. M., & Propst, D. L. (2003). EFFECTS OF RIVER FLOW REGIME ON COTTONWOOD LEAF LITTER DYNAMICS IN SEMI-ARID NORTHWESTERN COLORADO. *The Southwestern Naturalist*, *48*(2), 188–201.

Arp, C. D., Gooseff, M. N., Baker, M. A., & Wurtsbaugh, W. (2006). Surface-water hydrodynamics and regimes of a small mountain stream-take ecosystem. *Journal of Hydrology*, *329*(3-4), 500–513.

Bales, R. C., Molotch, N. P., Painter, T. H., Dettinger, M. D., Rice, R., & Dozier, J. (2006). Mountain hydrology of the western United States. *Water Resources Research*, 42(8). Retrieved from <u>https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1029/2005WR004387</u> Baron, J. S., LaFrancois, T., & Kondratieff, B. C. (1998). CHEMICAL AND BIOLOGICAL CHARACTERISTICS OF DESERT ROCK POOLS IN INTERMITTENT STREAMS OF CAPITOL REEF NATIONAL PARK, UTAH. *The Great Basin Naturalist*, *58*(3), 250–264.

Cooper, D. J., Kaczynski, K. M., Sueltenfuss, J., Gaucherand, S., & Hazen, C. (2017). Mountain wetland restoration: The role of hydrologic regime and plant introductions after 15 years in the Colorado Rocky Mountains, USA. *Ecological Engineering*, *101*, 46–59.

Faustini, J. M., & Jones, J. A. (2003). Influence of large woody debris on channel morphology and dynamics in steep, boulder-rich mountain streams, western Cascades, Oregon. *Geomorphology*, 51(1), 187–205.

Fonstad, M. A. (2003). Spatial variation in the power of mountain streams in the Sangre de Cristo Mountains, New Mexico. *Geomorphology*, 55(1), 75–96.

Gippel, C. J. (1995). Environmental Hydraulics of Large Woody Debris in Streams and Rivers. *Journal of Environmental Engineering*, *121*(5), 388–395.

Godsey, S. E., & Kirchner, J. W. (2014). Dynamic, discontinuous stream networks: hydrologically driven variations in active drainage density, flowing channels and stream order: SCIENTIFIC BRIEFING. *Hydrological Processes*, *28*(23), 5791–5803.

Heede, B. H. (1985). Channel adjustments to the removal of log steps: an experiment in a mountain stream. *Environmental Management*, 9(5), 427–432.

Lester, R. E., & Wright, W. (2009). Reintroducing wood to streams in agricultural landscapes: changes in velocity profile, stage and erosion rates. *River Research and Applications*, *25*(4), 376–392.

Mason, R. R., Jr, Simmons, C. E., & Watkins, S. A. (1990). *Effects of channel modifications on the hydrology of Chicod Creek basin, North Carolina, 1975-87* (No. 90-4031). U.S. Geological Survey ; Books and Open-File Reports Section [distributor], <u>https://doi.org/10.3133/wri904031</u>

Oswald, E. B., & Wohl, E. (2008). Wood-mediated geomorphic effects of a jökulhlaup in the Wind River Mountains, Wyoming. *Geomorphology*, *100*(3), 549–562.

Shields F. Douglas, & Gippel Christopher J. (1995). Prediction of Effects of Woody Debris Removal on Flow Resistance. *Journal of Hydraulic Engineering*, *121*(4), 341–354.

Smith, R. D., Sidle, R. C., Porter, P. E., & Noel, J. R. (1993). Effects of experimental removal of woody debris on the channel morphology of a forest, gravel-bed stream. *Journal of Hydrology*, *152*(1), 153–178.

Turner, D. S., & Richter, H. E. (2011). Wet/dry mapping: using citizen scientists to monitor the extent of perennial surface flow in dryland regions. *Environmental Management*, 47(3), 497–505.

Wohl, E., & Goode, J. R. (2008). Wood dynamics in headwater streams of the Colorado Rocky Mountains: WOOD DYNAMICS IN COLORADO. *Water Resources Research*, 44(9), 201.

Wohl, E., Madsen, S., & MacDonald, L. (1997). Characteristics of log and clast bed-steps in step-pool streams of northwestern Montana, USA. *Geomorphology*, 20(1), 1–10.

Other Topics

Austin, G., & Cooper, D. J. (2016). Persistence of high elevation fens in the Southern Rocky Mountains, on Grand Mesa, Colorado, USA. *Wetlands Ecology and Management*, 24(3), 317–334.

Bartos, D. L. (2001). Landscape dynamics of aspen and conifer forests. Retrieved from https://digitalcommons.usu.edu/cgi/viewcontent.cgi?article=1786&context=aspen bib

Berkowitz, J. F. (2011). Recent Advances in Wetland Delineation—Implications and impact of Regionalization. *Wetlands*, *31*(3), 593–601.

Caruso, B. S. (2014). GIS-Based Stream Classification in a Mountain Watershed for Jurisdictional Evaluation. *JAWRA Journal of the American Water Resources Association*, 50(5), 1304–1324.

Caruso, B. S., & Haynes, J. (2010). Connectivity and Jurisdictional Issues for Rocky Mountains and Great Plains Aquatic Resources. *Wetlands*, *30*(5), 865–877.

Clow, D. W. (2010). Changes in the Timing of Snowmelt and Streamflow in Colorado: A Response to Recent Warming. *Journal of Climate*, *23*(9), 2293–2306.

Deacon, J. R., Mize, S. V., & Spahr, N. E. (1999). Characterization of selected biological, chemical, and physical conditions at fixed sites in the Upper Colorado River basin, Colorado, 1995-98. Retrieved from https://pubs.usgs.gov/wri/wri99-4181/pdf/wrir99-4181.pdf

Geiger, S. T., Daniels, J. M., Miller, S. N., & Nicholas, J. W. (2014). Influence of Rock Glaciers on Stream Hydrology in the La Sal Mountains, Utah. *Arctic, Antarctic, and Alpine Research*, *46*(3), 645–658.

Harrelson, C. C., Rawlins, C. L., & Potyondy, J. P. (1994). Stream channel reference sites: an illustrated guide to field technique. *Gen. Tech. Rep. RM-245. Fort Collins, CO: US Department of Agriculture, Forest Service, Rocky Mountain Forest and Range Experiment Station. 61 P., 245.* Retrieved from https://www.fs.usda.gov/treesearch/pubs/20753

Papadaki, C., Soulis, K., Ntoanidis, L., Zogaris, S., Dercas, N., & Dimitriou, E. (2017). Comparative Assessment of Environmental Flow Estimation Methods in a Mediterranean Mountain River. *Environmental Management*, 60(2), 280–292.

Patterson, L., & Cooper, D. J. (2007). The use of hydrologic and Ecological indicators for the restoration of drainage ditches and water diversions in a Mountain Fen, Cascade Range, California. *Wetlands*, *27*(2), 290–304.

Peckarsky, B. L., McIntosh, A. R., Horn, S. C., McHugh, K., Booker, D. J., Wilcox, A. C., ... Alvarez, M. (2014). Characterizing disturbance regimes of mountain streams. *Freshwater Science*, *33*(3), 716–730. Serreze, M. C., Clark, M. P., & Frei, A. (2001). Characteristics of large snowfall events in the montane western United States as examined using snowpack telemetry (SNOTEL) data. *Water Resources Research*, *37*(3), 675–688.

Shoutis, L., Patten, D. T., & McGlynn, B. (2010). Terrain-based Predictive Modeling of Riparian Vegetation in a Northern Rocky Mountain Watershed. *Wetlands*, *30*(3), 621–633.

Wohl, E., Cooper, D., Poff, L., Rahel, F., Staley, D., & Winters, D. (2007). Assessment of stream ecosystem function and sensitivity in the Bighorn National Forest, Wyoming. *Environmental Management*, *40*(2), 284–302.