# **Supplemental Information**

# Genomic and phenotypic response of honyhead turbot exposed to municipal wastewater effluents

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### Microarray data analysis

The improved multispecies array contained an expanded probe content including targets with defined roles in endocrine pathways and processes (Baker, 2005, 2011; Blanchette et al., 2007; Filby et al., 2007; Ollikainen et al., 2006), in addition to well-defined biomarkers for contaminant exposure (Hutchinson et al., 2006; Katchamart et al., 2002). Statistical analysis of the microarray experiment involved two steps: normalization of microarray data as described previously using an mloess algorithm, and sorting of the genes according to interest (Baker et al., 2009). To investigate alterations in gene expression of controls and exposed fish, we used the Significance Analysis of Microarrays [SAM] procedure as described by Tusher et al. (Tusher et al., 2001) and its implementation in the official statistical package SAMR.

In order to not be unduly impressed by accidental small variances, we set the percentile of standard deviation values used for the exchangeability factor s0 in the test statistic to 90. We used a false discovery rate (FDR) threshold of 0.2 as the criterion of significance, except in the case of 0.5% SD effluent comparison to seawater, in which case we used FDR < 0.3. In the 0.5% LA effluent comparison to seawater, we do not estimate FDR at all; rather, we set a cutoff for the test statistic itself based on our experience with the other comparisons. Special treatment of these two cases was necessitated by a combination of smaller biological effect and a smaller group size (n = 1 in the 0.5% LA effluent group). SAM helped identify genes with statistically significant changes in expression by assigning a score on the basis of its change in gene expression relative to the standard deviation of repeated measurements for that gene.

Genes with scores greater than the predetermined threshold were considered to be significant. Scatter plots of the signal intensity for each microarray spot were log2 transformed

SI-2

for comparison. Thus, the log2 signal for each gene for the sea water control was compared with that of signals from the corresponding gene in fish exposed to ammonia, 0.5% effluent, 5% effluent and 30 nM E2 for two weeks.

The FDR is a q-value (a measure of false positives) rather than a p-value (a measure of statistical significance). It should be emphasized that an FDR of 0.2 is very relevant when dealing with a relatively small number of differentially expressed transcripts, as in the case of the array data analyzed here. If there are 20 differentially expressed transcripts, with a q-value of 0.2, one would expect two false positives. If we were dealing with a larger gene set, say 2000 genes, and then a q-value of 0.2 would be unacceptable. However, based on the small number of differentially expressed transcripts being analyzed, a q-value of 0.2 is acceptable, (given that key findings are validated by qPCR).

#### **Microarray Probe Validation**

In a previous study by members of this group (Rouse et al., 2008, BMC Res Notes. 2008 Jul 11; 1:45) we examined the effects of probe variability, signal dynamic range and feature morphology and from this prior work have an excellent sense of the sensitivity of the array used in this study. We evaluated multiple probes for the same gene on this 15K array, using diagnostic heat maps to ensure probe validity and uniformity of performance before the data was considered in our downstream analysis. Data were averaged across replicate probes. As averaging is additive in nature, the non-detection of low abundance transcripts due to a dilution effect did not pose a problem. The replicate features can be considered as one giant or large probe and this additive effect provides advantages for low level transcripts, enabling their detection. The microarray is based on the Agilent *in situ* synthesized oligonucleotide microarray, which is a highly sensitive platform (Hardiman, 2004).

#### Amplification and sequencing of hornyhead turbot mRNAs

We amplified partial turbot transcripts using conserved sequences from other fish species to guide the choice of primer design. Gene-specific primers were designed using Primer3 software (Rozen and Skaletsky, 2000). A 117 bp sequence was amplified from the *VTG1* transcript using the following primers; forward 5'- ATGAAGGGACAGACCTGTGG and reverse 5'- AACCCAGGAATGAGCATAGC. A 127 bp sequence was amplified from *VTG2* transcript using the following primers; forward 5'- ACTGGATGAGAGGCCAGACTT and reverse 5'- GGTAGAACCCAGGAATGAGC.

A 92 bp sequence was amplified from *CYP1A* transcript using the following primers; forward 5'- TCACTGTGAGGACAGGAAGCT and reverse 5'-CAGCACCGAACAGGTCATT. A 98 bp sequence was amplified from  $\beta$ -actin using the following primers; forward 5'- TCCCTGGAGAAGAGCTACGA and reverse 5'-AGGAAGGAAGGCTGGAAGAG.

For amplification and sequencing, oligonucleotide primers were obtained from Invitrogen. PCR was carried out in 50µl volume, 5µl (10X) reaction buffer 200mMTrisHCl/100mMKCl/100mM(NH<sub>4</sub>)<sub>2</sub>SO<sub>4</sub>/20mMMgSO<sub>4</sub>/1%TritonX/BSA/1mg/m, 0.1mM dNTPs, 2.5 units of *Pfu*, 1µl 1µM forward primers, 1µl 1µM reverse primers, 200ng turbot liver cDNA. After a denaturing step for 10 min at 96°C, touchdown amplification was performed with 35 cycles of 45s at 96°C, 45s at 55° to 50°C, in one degree increments and 1 min at 72°C.

SI-4

All reactions were evaluated on a 1% agarose gel stained with ethidium bromide to validate the reaction and the products were directly sequenced using the respective forward and reverse PCR primers. All of the sequencing reads were subjected to a series of quality control measures, including a Phred quality score > 20, and manual trace inspection. The identity of each sequence was confirmed by performing BLAST searches of GenBank.

### Quantitative real-time PCR analysis

Relative turbot mRNA transcript levels were measured by real-time quantitative (qPCR) in a LightCycler 480. Total RNA was extracted from hepatic turbot samples as described above and reverse-transcribed using the Roche Transcriptor kit and 50 ng cDNA were quantified using LightCycler 480 SYBR Green Master kit. For selection of housekeeping genes appropriate for normalizing the qPCR array data, we investigated three housekeeping genes,  $\beta$ -actin, G3PDH and 18srRNA. Of these three, G3PDH and  $\beta$ -actin performed equally well. The  $\beta$ -actin transcript was chosen to serve as an internal control for normalization.

Each sample was run in duplicate using the primer pairs described above and mean values were reported. Amplification and melting curve analyses for these assays indicated that the amplifications were specific. Normalized gene expression values were obtained using LightCycler Relative Quantification software. Relative gene copy numbers were derived using the formula  $2^{\Delta CT}$  where  $\Delta CT$  is the difference in amplification cycles required to detect amplification product from equal starting concentrations of turbot liver RNA (Untergasser et al., 2007).

## REFERENCES

Baker, M.E., 2005. Xenobiotics and the evolution of multicellular animals: emergence and diversification of ligand-activated transcription factors. Integrative and Comparative Biology 45, 172-178.

Baker, M.E., 2011. Origin and diversification of steroids: co-evolution of enzymes and nuclear receptors. Mol Cell Endocrinol 334, 14-20.

Baker, M.E., Ruggeri, B., Sprague, L.J., Eckhardt-Ludka, C., Lapira, J., Wick, I., Soverchia, L., Ubaldi, M., Polzonetti-Magni, A.M., Vidal-Dorsch, D., Bay, S., Gully, J.R., Reyes, J.A., Kelley, K.M., Schlenk, D., Breen, E.C., Sasik, R., Hardiman, G., 2009. Analysis of Endocrine Disruption in Southern California Coastal Fish Using an Aquatic Multispecies Microarray. Environmental Health Perspectives 117, 223-230.

Blanchette, B., Feng, X., Singh, B.R., 2007. Marine glutathione S-transferases. Mar Biotechnol (NY) 9, 513-542.

Filby, A.L., Santos, E.M., Thorpe, K.L., Maack, G., Tyler, C.R., 2007. Gene expression profiling for understanding chemical causation of biological effects for complex mixtures: a case study on estrogens. Environ Sci Technol 41, 8187-8194.

Hardiman, G., 2004 Microarray platforms--comparisons and contrasts. Pharmacogenomics 5, 487-502.

Hutchinson, T.H., Ankley, G.T., Segner, H., Tyler, C.R., 2006. Screening and testing for endocrine disruption in fish-biomarkers as "signposts," not "traffic lights," in risk assessment. Environ Health Perspect 114 Suppl 1, 106-114.

Katchamart, S., Miranda, C.L., Henderson, M.C., Pereira, C.B., Buhler, D.R., 2002. Effect of xenoestrogen exposure on the expression of cytochrome P450 isoforms in rainbow trout liver. Environ Toxicol Chem 21, 2445-2451.

Ollikainen, N., Chandsawangbhuwana, C., Baker, M.E., 2006. Evolution of the thyroid hormone, retinoic acid, and liver X receptors. Integrative and Comparative Biology 46, 815-826.

Rozen, S., Skaletsky, H.J., 2000. Primer3 on the WWW for general users and for biologist programmers, in: Misener, S., A., K.S. (Eds.), Bioinformatics Methods and Protocols: Methods in Molecular Biology. Humana Press Inc, Totowa (NJ), pp. 365-386.

Tusher, V.G., Tibshirani, R., Chu, G., 2001. Significance analysis of microarrays applied to the ionizing radiation response. Proc Natl Acad Sci U S A 98, 5116-5121.

Untergasser, A., Nijveen, H., Rao, X., Bisseling, T., Geurts, R., Leunissen, J.A., 2007. Primer3Plus, an enhanced web interface to Primer3. Nucleic Acids Res 35, W71-74.

| Gene ID  | Gene Name   | Function  |
|--|---|---|
| FXR  | Farnesoid X receptor  | Bile biosynthesis/transcription factor            |
| HSP70  | Heat shock protein 70   | Cellular stress response                          |
| HSP90  | Heat shock protein 90   | Cellular stress response                          |
| AHR  | Aryl hydrocarbon receptor   | Contaminant response/Detoxification               |
| CYP1A  | Cytochrome P450, family 1, subfamily A  | Contaminant response/Detoxification               |
| СҮРЗА  | Cytochrome P450, family 3, subfamily A  | Contaminant response/Detoxification               |
| GSTa   | Glutathione S-transferase   | Contaminant response/Detoxification               |
| MT1  | Metallothionein   | Contaminant response/Detoxification               |
| IGFBP1   | Insulin-like growth factor building protein1  | Development/Growth                                |
| IGFBP2   | Insulin-like growth factor-binding protein 2  | Development/Growth                                |
| IGFBP5   | Insulin-like growth factor-binding protein 5  | Development/Growth                                |
| VEGF   | Vascular endothelial growth factor  | Development/Growth                                |
| TRα  | Thyroid receptor (alpha)  | Development/mediates thyroid activities           |
| TRβ  | Thyroid receptor (beta)   | Development/mediates thyroid activities           |
| G3PDH  | Glyceraldehyde 3-phosphate dehydrogenase  | Glycolysis/nuclear functions                      |
| GRF1   | Glucocorticoid receptor 1   | Hormone signaling                                 |
| HAMP   | Hepcidin  | Infection response                                |
| PPARy  | Ry Peroxisome proliferator activated receptor gamma Lipid metabolism/development/transcription factor |   |
| PPARα  | Peroxisome proliferator activated receptor alpha  | Lipid metabolism/development/transcription factor |
| LXR  | Liver X receptor  | Lipid metabolism/transcription factor             |
| ZP2  | Zona pellucida glycoprotein 2   | Reproduction/egg: sperm binding                   |
| ZP3  | Zona pellucida glycoprotein 3   | Reproduction/egg: sperm binding                   |
| VTG1   | Vitellogenin A  | Reproduction/egg: yolk protein                    |
| VTG2   | Vitellogenin B  | Reproduction/egg: yolk protein                    |
| CYP19  | Cytochrome P450, family 19 (Aromatase)  | Reproduction/Steroidogenesis                      |
| HSD17B1Hydroxysteroid (17-beta) dehydrogenase type 1 |   | Reproduction/Steroidogenesis                      |
| CYP11B   | Cytochrome P450, family 11, subfamily B1  | Reproduction/Steroidogenesis/stress               |
| HSD11B1  | Hydroxysteroid (11-beta) dehydrogenase 1  | Reproduction/Steroidogenesis/stress               |
| AR   | Androgen receptor   | Reproduction/transcription factor                 |
| ER1  | Estrogen receptor alpha   | Reproduction/transcription factor                 |
| ER2  | Estrogen receptor beta  | Reproduction/transcription factor                 |
| PR   | Progesterone receptor   | Reproduction/transcription factor                 |
| RAR  | Retinoic acid receptor  | Retinoid hormone transcription factor             |
| RXR  | Retinoid receptor   | Retinoid hormone transcription factor             |
| SHBG   | Sex hormone-binding globulin  | Sex steroid hormone binding                       |
| MR   | Mineralocorticoid receptor  | Steroid hormone transcription factor              |
| VDR  | Vitamin D receptor  | Steroid hormone transcription factor              |
| StAR   | Steroidogenesis acute regulator   | Steroidogenesis/other endocrine tissues           |

# Table SI-1. Genes contained in the microarray (additional functions are not described).

Table SI-2. List of analytes investigated and their reporting limits (RL), grouped as current-use pesticides (CUP), hormones, industrial and commercial compounds (ICC), and pharmaceuticals and personal care products (PPCP).

| Group    | Chemical Name                            | RL     |
|----------|--|--------|
| CUP      | Atrazine                                 | 0.0003 |
| CUP      | Lindane                                  | 0.001  |
| Hormones | Estradiol                                | 0.0005 |
| Hormones | Estrone                                  | 0.0002 |
| Hormones | Etynylestradiol                          | 0.001  |
| Hormones | Progesterone                             | 0.0005 |
| Hormones | Testosterone                             | 0.0005 |
| ICC      | Benzophenone                             | 0.053  |
| ICC      | Bisphenol A                              | 0.05   |
| ICC      | Nonylphenol                              | 0.08   |
| ICC      | Tris(1-chloro-2-propyl) phosphate (TCPP) | 0.01   |
| ICC      | Tris(2-chloroethyl) phosphate (TCEP)     | 0.1    |
| PPCP     | Atenolol                                 | 0.001  |
| PPCP     | Atorvastatin                             | 0.003  |
| PPCP     | Caffeine                                 | 0.005  |
| PPCP     | Carbamazepine                            | 0.005  |
| PPCP     | DEET                                     | 0.001  |
| PPCP     | Diazepam                                 | 0.0003 |
| PPCP     | Diclofenac                               | 0.003  |
| PPCP     | Dilantin                                 | 0.001  |
| PPCP     | Fluoxetine                               | 0.0005 |
| PPCP     | Gemfibrozil                              | 0.003  |
| PPCP     | Ibuprofen                                | 0.05   |
| PPCP     | lopromide                                | 0.011  |
| PPCP     | Meprobamate                              | 0.003  |
| PPCP     | Musk Ketone                              | 0.025  |
| PPCP     | Naproxen                                 | 0.005  |
| PPCP     | Primidone                                | 0.0005 |
| PPCP     | Sulfamethoxazole                         | 0.003  |
| PPCP     | Triclosan                                | 0.01   |
| PPCP     | Trimethoprim                             | 0.003  |



0.5% Effluent Treatments





Figure SI-1. Average concentrations for compounds detected in advanced-primary treated effluent (PL) and full-secondary treated effluent (HTP) in exposure tanks.