# Comparison and verification of bacterial water quality indicator measurement methods using ambient coastal water samples

**ABSTRACT** - More than 30 laboratories routinely monitor water along southern California's beaches for bacterial indicators of fecal contamination. Data from these efforts frequently are combined and compared even though three different methods (membrane filtration (MF), multiple tube fermentation (MTF), and chromogenic substrate (CS) methods) are used. To assess data comparability and quantify variability within method and across laboratories, 26 laboratories participated in an intercalibration exercise. Each laboratory processed three replicates from eight ambient water samples employing the method or methods they routinely use for water quality monitoring. Verification analyses also were conducted on a subset of wells from the CS analysis to confirm or exclude the presence of the target organism. Enterococci results were generally comparable across methods. Confirmation revealed a 9% false positive rate and a 4% false negative rate in the CS verifications for enterococci, though these errors were small in the context of within- and among-laboratory variability. Fecal coliforms also were comparable across all methods, though CS underestimated the other methods by about 10%, probably because it measures only E. coli, rather than the larger fecal coliform group measured by MF and MTF. CS overestimated total coliforms relative to the other methods by several fold and was found to have a 40% false positive rate in verification. Across-laboratory variability was small relative to within- and amongmethod variability, but only after data entry errors were corrected. Nearly 20% of the labs had data entry errors, which were much larger than any method-related errors.

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### INTRODUCTION

Southern California's beaches are monitored extensively to screen for fecal contamination from human activities, such as wastewater discharges, industrial input, and surface runoff (Schiff et al. 2002). More than 30 groups are involved in this monitoring, including city and county health departments, treated wastewater dischargers, stormwater permittees, and non-profit environmental organizations. These groups all measure the same parameters (enterococci, fecal coliforms and total coliforms), but have the option of choosing from a number of different measurement methods. Wastewater dischargers primarily rely on MF. Stormwater agencies and environmental groups primarily use the IDEXX® CS method. Health departments historically have relied on MTF and MF, but have begun to use CS more frequently in the last several years.

Data from these multiple providers are collated and used collectively in several ways. On a daily basis, they are used to assess beach water quality and as the basis for issuance of beach water quality warnings. On a long-term basis, they are integrated to identify chronically contaminated beaches for Section 303(d) listing under the federal Clean Water Act and for development of environmental report cards that compare water quality among locations and over time. Using these data interchangeably assumes that results from multiple laboratories using different measurement methods are comparable, even though the laboratories may have varying levels of

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proficiency and may employ detection methods that rely on widely different products of bacterial growth.

A number of studies have compared the response of MF and MTF, and a few studies have compared these methods to CS (Kinzelman et al. 2003, Francy and Darner 2000, Abbott et al. 1998, Eckner 1998, Budnick et al. 1996, Palmer et al. 1993, Bej et al. 1991, Edberg et al. 1990, Covert et al. 1989). Noble et al. (2003a) was the first to compare results among all three methods and place differences among methods into the context of variability among laboratories that use the same method. However, Noble et al. used fabricated samples created primarily from laboratory strains of bacteria seeded into clean matrices. Natural ambient water samples often contain contaminants, such as humic acids and suspended solids, which have the potential to interfere with these methods. Natural samples also contain native bacteria, such as Aeromonas, Vibrio, Pseudomonas, and Flavobacteria spp., which have been shown to produce positive reactions in substrates containing 4methylumbelliferyl- $\beta$ -glucuronide (MUG) and can lead to false positives in the CS test (Pisciotta et al. 2002, Landre et al. 1998, Davies et al. 1995, Hidalgo et al. 1977).

Here, we present an intercalibration study modeled after Noble *et al.* (2003a), but based on measurement of ambient water samples collected from sites known to have complex matrices. The study assessed comparability of results among 26 southern California laboratories that conduct routine bacterial monitoring using three bacterial indicator measurement methods, evaluated the reliability of CS methods through verification of target organisms, and identified common causes of error in determining bacterial concentrations for water quality monitoring purposes.

# **Methods**

Ambient water samples were collected from eight sites throughout southern California, including open marine beaches, estuaries, and flowing creeks carrying dry-weather urban runoff (Table 1). All samples were collected in sterile, 20 L carboys following Standard Methods 1060 protocol for aseptic sampling techniques (APHA 1995). Samples were then transported on ice to the Orange County Sanitation District (OCSD) laboratory in Fountain Valley, California. Upon arrival, carboys were placed on magnetic stir-plates, a sterile stir bar was added, and samples were stirred continuously for a minimum of 20 minutes to ensure homogeneity. Water from each carboy was dispensed into 26 sets of sterile, pre-labeled 100 mL bottles, which were transported on ice to participating laboratories. Sample processing began simultaneously at all laboratories at a pre-arranged time to eliminate differences due to holding time.

Samples were analyzed for total coliform, fecal coliform, *Escherichia coli*, and enterococci, using the method or methods routinely performed by each laboratory. Three classes of methods were used: chromogenic substrate, multiple-tube fermentation, and membrane filtration. Each laboratory analyzed multiple dilutions of each sample to minimize the number of samples occurring outside of a quantifiable range. All analyses were performed in triplicate.

Laboratories performing CS used IDEXX media and the Quanti-Tray/2000® system for all samples, following the manufacturer's instructions. Colilert®-18 media were used for enumeration of total coliform and *Escherichia coli*, while Enterolert® media were used for enterococci.

Multiple-tube fermentation for total coliform recovery used APHA 9221B (LTB/BGB). Fecal coliform recovery was by either APHA Method 9221E.1 (EC) or APHA Method 9221E.2 (A-1). The enterococcus group was enumerated using the APHA Method 9230B (azide dextrose/PSE; APHA 1995).

Membrane filtration for total coliform recovery was by APHA Method 9222B (M-Endo), both single-step and enrichment-step procedures. Fecal coliform enumeration using MF was by APHA Method 9222D (M-FC). Either the 48-h APHA Method 9230C (M-E) or the 24-h EPA Method 1600 (M-EI) procedures (APHA 1995) were used for enterococci.

Confirmations were conducted for CS by five laboratories. When available, 10 positive wells were selected randomly for confirmation from trays with 80% or more positive wells. In certain instances, where a low number of positive wells was present, a smaller number of wells was selected from the tray. This resulted in confirmation testing for 71 entero-cocci wells, 35 *E. coli* wells, and 153 total coliform wells. In addition, 55 non-fluorescing Enterolert<sup>®</sup> wells and 21 weakly fluorescing wells (scored as not containing enterococci following manufacturer protocols) were subjected to verification analysis to test for false negatives.

Location	Site Description	Enterococci	Fecal Coliforms <i>(E. coli)</i>	Total Coliforms		
	<b>-</b> (	20	100	0.400		
Ballona Wetlands	Estuary	80	130	2,400		
Cabrillo Beach	Embayment	285	500	820		
Doheny Beach	Open Beach	22	20	98		
MBW 6	Urban Creek	130	100	12,033		
MBW 10	Urban Creek	199	41	2,000		
Malibu Beach	Open Beach	10	488	1,194		
Malibu Lagoon	Estuary	122	300	5,191		
Mission Bay	Embayment	120	285	800		

Table 1. Median concentrations of fecal indicator bacteria per 100 mL reported in each sample across all methods.

Bacterial isolates for confirmation testing were obtained by wiping the back of the Quanti-Tray<sup>TM</sup> with 70% isopropanol, puncturing the well with either a sterile syringe or sterile scalpel, and withdrawing the well contents. Total coliform bacteria were confirmed by transferring well contents to either Tryptic Soy (TSB) or Brain Heart Infusion (BHI) broths, incubating at 35°C, transferring again to Brilliant Green Bile broth (BGB), and subsequently plating to MacConkey agar. E. coli was confirmed by transferring well contents to either EC or A-1 broths, incubating at 44.5°C and plating to either MacConkey or LES Endo agars. Enterococcus was confirmed by inoculating either TSB or BHI broth with well contents, incubating at 35°C, then filtering the broth and plating to either mE or mEI agar, and incubating at 41°C. Filters on mE subsequently were transferred to EIA agar. Secondary confirmations were performed on each isolate that tested positive using a second EPA- or APHA-approved biochemical testing method (MF or MTF, as described previously) or by submitting isolates to the Vitek microbial identification system (bioMérieux, Hazelwood, MO).

Statistical differences in median concentrations among methods and within-lab variability among methods were assessed using ANOVA on ranks (Conover and Iman 1981). Median values from replicate samples then were used to compute ranks across methods separately for each station. Where statistical differences among methods were detected (p > 0.05), individual stations were examined for possible station effects using Bonferroni-adjusted significance levels. Within-lab variability among methods was examined by taking the standard deviations of log counts across replicates for each laboratory sample and then ranking them within station. Bonferroni-adjusted significance levels were employed to control overall error rates at 0.05 when testing for station effects.

Estimates of variability for each method, based on the median standard deviation of log counts and the average median log ratios of these counts, were used to estimate the reported bacterial concentrations at which one would be 95% or 99% confident that a single sample analyzed would fall above or below California's standards at which public health warnings are issued.

To assess the importance of individual laboratory error, data were examined to determine how well individual laboratory results agreed with the results from the entire group of laboratories. This was done by identifying the number of samples produced by each laboratory that were more than a half log unit above or below the overall median. This criterion was chosen because it is roughly equal to the average within-lab method variability observed in previous studies (Noble *et al.* 2003a).

# RESULTS

Median concentrations of enterococci exceeded California's single sample standard (104/100 mL) at five of the eight sample sites (Table 1). State standards for fecal coliforms (400/100 mL) and total coliforms (10,000/100 mL) were exceeded at two stations and one station, respectively.

There was no statistical difference in median concentration between MF and MTF for any of the bacterial indicators, though there was a difference between CS and MF for all three indicators (Table 2). For enterococci, CS produced lower concentrations than MF, but most of the difference was attributable to a single station (Figure 1a). The median CS value at Doheny Beach was several-fold lower than that for either MF or MTF, but measured con-

Comparison	All	Enterococci	Fecal	Total Coliform		
	Sites	Without Doheny	All Sites	All Sites		
CS/MF	0.71*	0.86	0.69*	0.91	2.54*	
CS/MTF	0.80	1.03	1.03	0.95	3.80*	
MF/MTF	1.13	1.20	1.46	1.04	1.49	

Table 2. Estimated median ratios of log counts between methods, for each indicator. Asterisk indicates statistically different than 1.

centrations at that site were low for all methods. The median concentration was only 22/100 mL, with more than three-quarters of the CS values and almost half of the MF and MTF values reported as non-detectable values. When the Doheny Beach station was removed from the analysis, there was no statistical difference among methods for enterococci (Table 2).

The median CS values for fecal coliforms were similar to those produced by MTF, but about 30% lower than those produced using MF. Though CS methods are known to underestimate fecal coliform levels because they detect only *E*. coli, the majority of the difference was attributable to low bacterial concentrations at two stations, Doheny Beach and MBW10 (Figure 1b). Median concentrations at the Doheny Beach and MBW10 stations were only 20/100 mL and 41/100 mL, respectively. When these stations were removed from the calculation, the difference between results from CS and MF was only 9%.

For total coliforms, median concentrations from CS were several-fold higher than from either MF or MTF. CS produced the highest median concentration among the three methods at five of the eight stations (Figure 1c).

Verification analyses of CS results revealed a large percentage of false positive wells for total coliforms. Only 93 of 153 positive wells (61%) from IDEXX Quanti-Trays<sup>TM</sup> were confirmed to contain bacteria from the total coliform group, a finding consistent with the higher median concentrations measured using CS. In contrast, *E. coli* was successfully isolated from 100% of the positive wells tested. For Enterolert, 67 of 71 (94%) fluorescing wells and 5 of 55 (9%) empty wells were confirmed to contain enterococci. Only 3 of the 21 tested wells exhibiting weak fluorescence contained enterococci.

MTF exhibited the highest within-laboratory variability of any method for all three indicators, typically twice that of the other two methods (Table 3). MF had the lowest variability among methods for all three indicators. When within-method variability was expressed as a 95% confidence interval for a measurement at existing water quality standards, MTF measurements needed to be three times the water quality standards to be confident that the true value actually exceeded the standard. For MF, a single measured value only needed to be 50% above the standard to be confident that the true value exceeded the standard.

There appeared to be little effect associated with individual laboratories. Only two laboratories had more than one ENT sample for which the result differed by greater than a half log unit from the group median. One of those laboratories later was found to have a defective incubator that did not hold the proper temperature. For FC, two laboratories performing MF and one lab performing MTF reported values above the target range, but this reflected the comparatively large number of CS observations, which measure only E. coli and reduced the grand median. When comparisons were limited to within-method median, no results differed by more than half a log unit from the interlaboratory median. Comparisons were limited to within-method for TC because of CS bias for this indicator. In that comparison, no lab results differed by more than a half log unit from the group median.

#### DISCUSSION

All three methods produced similar results for enterococci. Verifications confirmed both false positives and false negatives using CS, but both rates were small relative to the within-laboratory measurement variability. This is consistent with Fleischer (1990) and Noble *et al* (2003a), but differs from Kinzelman *et al.* (2003), who found nearly 50% false positives for enterococcus verifications with CS. However, Kinzelman *et al.* suggested that their false



Figure 1. Median log counts vs. station for (A) Enterococci, (B) fecal coliforms/E.coli , and (C) total coliforms.

positives occurred primarily for weakly fluorescing wells, which is consistent with our results that 86% of weakly fluorescing wells did not contain enterococci. The CS manufacturer presently recommends that poorly fluorescing wells not be counted, which was the protocol used by laboratories in this study.

The only large difference among methods observed was the severe overestimation of total coliform density using CS relative to the other two methods. The high rate of false positives likely results from interference by non-coliform organisms, such as *Aeromonas*, *Vibrio*, *Pseudomonas*, and *Flavobacteria spp.*, which are known to metabolize MUG (Pisciotta *et al.* 2002, Landre *et al.* 1998, Davies *et al.*1995, Hidalgo *et al.* 1977). In practical terms, though, this overestimation seems to have little effect on beach warning systems in California, as the total coliform standard is so high that the standard is almost never exceeded without fecal coliforms or enterococci also exceeding the standard (Noble *et al* 2003b).

While such a systematic methodological error is of concern, we found that the largest source of error was attributable to data processing. These data entry errors occurred for samples that the laboratories were aware were part of an intercalibration exercise, in which their results would come under greater scrutiny. Prior to the data analysis presented in this article, preliminary screening indicated that results from four of the labs differed by an order of magnitude from those of other labs. Upon inspection of original laboratory data sheets, we discovered that these labs failed to correct for dilution before data submission (which we subsequently corrected before conducting the analysis in this article). We also found that a fifth lab misaligned the sample numbers on the bottle with their internal tracking numbers, leading to values being submitted with the wrong sample number (again confirmed by examination of the original laboratory data sheets and corrected before our data analysis). These labs produced comparable data to all other labs after correcting for data submission errors. However, in typical applications, data from other labs is not available for comparison. These errors would have gone undetected and errant results would have been reported to managers for use in regulatory or public health decisions. These data management errors were far larger and more prevalent than any variability introduced by method or mingling of data across laboratories.

Method	Enterococci			Fecal Coliform/ <i>E.coli</i>			Total Coliform					
	(Std = 104)			(Std = 400)			(Std = 10,000)					
CS MF MTF	Ab 95% 166 146 209	ove 99% 217 178 314	Be 95% 65 73 52	low 99% 50 61 34	Ab 95% 624 767 856	ove 99% 809 1,120 1,331	Be 95% 256 208 186	low 99% 197 143 120	Abo 95% 14,955 14,696 21,409	ove 99% 18,880 18,365 33,266	Be 95% 5,297 5,445 3,000	low 99% 6,681 6,800 4,670

Table 3. Confidence intervals for each method applied to concentrations at California's single sample standards values.

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